4.1 Background

Chino Basin groundwater is not only a critical resource to overlying water producers; it is a critical resource to the entire Santa Ana Watershed. From a regulatory perspective, the use of Chino Basin groundwater to serve potable demands is limited by drinking water standards, groundwater basin water quality objectives, and Santa Ana River water quality objectives. In August 1999, Phase 1 of the OBMP established that groundwater monitoring must be conducted in order to obtain current water quality and water level data in Chino Basin (WEI, 1999). These data are necessary for defining and evaluating specific strategies and locations for the mitigation of nitrate, TDS, and other Constituents of Potential Concern (COPCs); new recharge sites; and pumping patterns that result from the implementation of the OBMP.

In the past, various entities have collected groundwater quality data. Municipal and agricultural water supply entities have collected groundwater quality data to comply with the Department of Health Services' requirements in the California Code of Regulations, Title 22, or for programs that range from irregular study-oriented measurements to long-term periodic measurements. Groundwater quality observations have been made by the DWR, by participants in the 1969 Judgment on the Santa Ana River (Orange County Water District vs. City of Chino et al.), by dischargers under orders from the RWQCB, and by the County of San Bernardino. The DWR and the San Bernardino County Flood Control District were very active in collecting groundwater quality data in the Chino Basin prior to the adjudication of the Chino Basin. After the Judgment was entered in 1978, monitoring south of State Route 60 stopped almost completely with the exception of that conducted by the Cities of Chino, Chino Hills, and Norco; the Jurupa Community Services District (JCSD); and the Santa Ana River Water Company. Most of the pre-1978 measurements were digitized by the DWR. In 1986, the MWDSC conducted the first comprehensive survey of groundwater quality, covering all constituents regulated under Title 22.

Watermaster initiated a regular monitoring program for Chino Basin in 1989. Groundwater quality data has been obtained periodically since 1990.

4.2 Water Quality Monitoring Programs

Watermaster began conducting a more robust monitoring program as part of the initial OBMP implementation. Watermaster's program relies on municipal producers, government agencies, and private consultants to supply their groundwater quality data on a cooperative basis. Watermaster supplements these data with data obtained through its own sampling and analysis program of private wells in the area generally south of State Route 60. Water quality data are also obtained from special studies and monitoring programs that take place under the orders of the RWQCB, the California Department of Toxic Substances Control (DTSC), and others. Watermaster has combined previously digitized groundwater quality data from all known sources into a comprehensive database.



4.2.1 Water Quality Monitoring Programs for Wells Owned by Municipal Water Suppliers

Water quality samples are collected from Appropriative Pool wells and some overlying Non-Agricultural Pool wells as part of formalized monitoring programs. Constituents include (i) those regulated for drinking water purposes in the California Code of Regulations, Title 22; (ii) those regulated in the 1995 Water Quality Control Plan for the Santa Ana River Basin (Basin Plan); or (iii) those that are of special interest to the pumper.

4.2.2 Water Quality Monitoring Programs for Private Water Supply Wells

Historically, private wells were sampled less methodically and less frequently than wells owned by members of the Appropriative Pool. As a result, there is little historical (pre-1999) groundwater quality information for most of the 600 private wells in the southern part of the Chino Basin. As mentioned above, the MWDSC conducted an assessment of water quality and water levels in the private wells south of State Route 60 in 1986. This assessment was a component of the Chino Basin groundwater storage program Environmental Impact Report (MWDSC et al., 1988). Nevertheless, the historical quality of groundwater produced at the majority of the wells in the southern Chino Basin is unknown.

In 1999, the Comprehensive Monitoring Program initiated the systematic sampling of private wells south of State Route 60 in the Chino Basin. Over a three-year period, Watermaster sampled all available wells at least twice to develop a robust baseline data set. This program has since been reduced to approximately 110 private key wells, and about half of these wells are sampled every other year. Groundwater quality samples are analyzed for general minerals, physical properties, and for regional COPCs (e.g. perchlorate, and volatile organic chemicals [VOCs] in the vicinity of known VOC plumes). This key well monitoring program provides a good representation of the areal groundwater quality in this portion of the basin.

4.2.3 Water Quality Monitoring Programs Conducted Pursuant to Regulatory Orders

Groundwater monitoring is conducted by private and public entities as part of regulatory orders and voluntary cleanups. These programs consist of networks of monitoring wells designed specifically to delineate and characterize the extent of the responsible party's contamination. These monitoring programs may include monthly, quarterly, and/or annual sampling frequencies. The following is a summary of all the regulatory and voluntary contamination monitoring in Chino Basin:

Plume: Alumax Aluminum Recycling Facility
 Constituent of Concern: TDS, sulfate, nitrate, chloride
 Order: RWQCB Cleanup and Abatement Order 99-38

• Plume: Chino Airport

Constituent of Concern: VOCs

Order: RWQCB Cleanup and Abatement Order 90-134



Plume: California Institute for Men Constituent of Concern: VOCs Order: Voluntary Cleanup Monitoring

Plume: Crown Coach International Facility
 Constituent of Concern: VOCs and Solvents

Order: Voluntary Cleanup Monitoring

• Plume: General Electric Flatiron Facility

Constituent of Concern: VOCs
Order: Voluntary Cleanup Monitoring

• Plume: General Electric Test Cell Facility

Constituent of Concern: VOCs
Order: Voluntary Cleanup Monitoring

• Plume: Kaiser Steel Fontana Site

Constituent of Concern: TDS/total organic carbon (TOC)

Order: See discussion in Section 4.36.7.

Plume: Milliken Sanitary Landfill
 Constituent of Concern: VOCs
 Order: RWQCB Order No. 81-003

Plume: Upland Sanitary Landfill
Constituent of Concern: VOCs
Order RWQCB Order No 98-99-07

• Plume: Ontario International Airport (VOC Plume – South of Ontario Airport)

Constituent of Concern: VOC

Order: This plume is currently being voluntarily investigated by a group of potentially responsible parties.

• Plume: Stringfellow National Priorities List (NPL) Site

Constituent of Concern: VOCs, perchlorate, N-nitrosodimethylamine (NDMA), heavy metals **Order:** The Stringfellow Site is the subject of US Environmental Protection Agency (EPA) Records of Decision (RODs): EPA/ROD/R09-84/007, EPA/ROD/R09-83/005, EPA/ROD/R09-87/016, and EPA/ROD/R09-90/048.

4.2.4 Other Water Quality Monitoring Programs

In a letter dated July 13, 2000, the RWQCB expressed their concern to the IEUA that the historical recharge of recycled water at IEUA Regional Plant No. 3 (RP3) may have caused groundwater contamination at down-gradient wells. Other sources of groundwater contamination in the area include the Kaiser Steel Mill, Alumax, other industries, and historical agricultural activities, including citrus groves and hog feed lots. Several municipal wells have been shut down in MZ3 due to perchlorate and nitrate in groundwater. MZ3 includes areas that underlie all or part of the Fontana Water Company, the Marygold Mutual Water Company, the CVWD, and the City of Ontario. MZ3 groundwater is tributary to wells owned by the JCSD.

To characterize groundwater levels and quality in MZ3, Watermaster and the IEUA



performed an investigation. The objectives of this investigation were to develop a groundwater sampling program, install two sentry wells at the distal end of the Kaiser plume, and perform further characterization of groundwater quality. Sampling was conducted at twenty-two selected key wells from late 2005 to 2007. Where possible, four quarterly samples and one annual sample were collected. In 2007, two triple-nested wells (MZ3-1 and MZ3-2) were installed down gradient of the Kaiser plume. These wells were sampled quarterly for one year. The sampling results provided data to further characterize the water quality patterns for contaminants of concern in the study area, including TDS, nitrate, sulfate, chloride, and perchlorate. And, the results from well MZ3-1/3 redefined the extent of the Kaiser plume.

4.2.5 Information Management

As with groundwater level and production data, Watermaster manages groundwater quality data in order to perform the requisite scientific and engineering analyses required to ensure that the goals of the OBMP are being met. Watermaster's relational database contains well location, construction, lithology, specific capacity, groundwater level, and water quality data. Historical water quality data for the period prior to the mid-1980s were obtained from the DWR and supplemented with data from producers in the Appropriative and Overlying Non-Agricultural Pools and others. For the period from the mid-1980s forward, Watermaster has QA/QC'd and uploaded water quality data from its own sampling programs, the State of California Department of Public Health (CDPH, formerly the Department of Health Services) database, and other cooperating parties to its relational database. Occasionally, problems have been found with CDPH data, usually occurring in the form of incorrect constituent identification. In 2003, Watermaster launched the Chino Basin Relational Database effort to collect water quality data directly from each member agency and thereby circumvent past data problems. Cooperating parties provide all data (including geologic, geophysical, water levels, water quality, production, and recharge) to Watermaster on a routine basis. These data are delivered in electronic format directly from the laboratory or from the cooperating party.

4.3 Groundwater Quality in Chino Basin

Figure 4-1 shows all wells with groundwater quality monitoring results for the 5-year period of July 2003 to June 2008.

Inorganic and organic constituents detected in groundwater samples from wells in the Chino Basin through June 2008 were analyzed synoptically. This analysis included all available data from production and monitoring wells. Hence, the data do not represent a programmatic investigation of potential sources nor do they represent a randomized study that was designed to ascertain the water quality status of the Chino Basin. These data do, however, represent the most comprehensive information available to date.

Monitoring wells targeted at potential sources tend to have greater concentrations than municipal or agricultural production wells. Wells with constituent concentrations greater than one-half of the MCL represent areas that warrant concern and inclusion in a long-term monitoring program. In addition, groundwater in the vicinity of wells with samples greater than the MCL may be impaired from a beneficial use standpoint.



Numerous water quality standards have been put in place by federal and state agencies. Primary MCLs are enforceable criteria that are set due to health effects. Secondary standards are related to the aesthetic qualities of the water, such as taste and odor. For some chemicals, there are "Notification Level" criteria that are set by the CDPH. When notification levels are exceeded, the CDPH recommends that the utility inform its customers and consumers about the presence of the contaminant and any health concerns associated with exposure. The level at which the CDPH recommends the drinking water system remove the affected drinking water source from service is the "Response Level." These levels range from 10 to 100 times the notification level, depending on the chemical. The following constituents exceeded at least one water quality criteria in more than 10 wells within the Chino Basin for the period of July 2003 through June 2008:

Analyte Group/Constituent	Wells with Exceedance
Inorganic Constituents	
Total Dissolved Solids	221
Nitrate-Nitrogen	395
Aluminum	153
Arsenic	24
Chloride	25
Chromium	30
Iron	185
Manganese	58
Perchlorate	188
Sulfate	41
Vanadium	25
General Physical	
Color	21
Odor	28
рН	14
Specific Conductance	121
Turbidity	78
Chlorinated VOCs	
1,1-Dichloroethane	11
1,1-Dichloroethene	31
1,2,3-Trichloropropane	23
1,2-Dichloroethane	17
cis-1,2-Dichloroethene	10
Tetrachloroethene (PCE)	37
Trichloroethene (TCE)	115

For all figures (Section 4 and Appendix B) that depict water quality distributions in the Chino Basin, the following convention is typically followed in setting class intervals in the legend (where WQS is the applicable water quality standard [see table below]). Variations of this convention may be employed to highlight certain aspects of the data.



Symbol	Class Interval
0	Not Detected
•	<0.5x WQS, but detected
•	0.5x WQS to WQS
0	WQS to 2x WQS
	2x WQS to 4x WQS
	> 4x WQS

4.3.1 Total Dissolved Solids

In Title 22, TDS is regulated as a secondary contaminant. The California secondary drinking water MCL for TDS is 500 mg/L. Figure 4-2 shows the distribution of the maximum TDS concentrations in Chino Basin from July 2003 through June 2008. During this period, maximum TDS concentrations ranged from 48 mg/L to 4,790 mg/L with average and median concentrations of approximately 550 mg/L and 380 mg/L, respectively. The highest concentrations are located south of State Route 60 where the impacts from agriculture are greatest, which is consistent with the data reported in the 2006 State of the Basin Report.

The impacts of agriculture on TDS in groundwater are primarily caused by dairy waste disposal, consumptive use, and fertilizer use on crops. As irrigation efficiency increases, the impact of consumptive use on TDS in groundwater also increases. For example, if source water has a TDS concentration of 250 mg/L and the irrigation efficiency is about fifty percent (flood irrigation), the resulting TDS concentration in returns to groundwater would be 500 mg/L, which is exclusive of the mineral increments from fertilizer. If irrigation efficiency is increased to seventy-five percent, the resulting TDS concentration in the returns to groundwater would be 1,000 mg/L, which is also exclusive of the mineral increments from fertilizer. For modern irrigated agriculture, the TDS impacts of consumptive use are more significant than mineral increments from fertilizers.

Wells with low TDS concentrations in close proximity to wells with higher TDS concentrations suggests a vertical stratification of water quality. However, there is a paucity of information concerning well construction/perforation intervals; Thus, the vertical differences in water quality are currently unverifiable.

4.3.2 Nitrate-Nitrogen

In Title 22, the primary MCL for nitrate as nitrogen (NO3-N) in drinking water is 10 mg/L. By convention, all nitrate values are expressed in this report as NO3-N. Figure 4-3 displays the distribution of maximum NO3-N concentrations in the Chino Basin from July 2003 through June 2008.

Areas with significant irrigated land use or dairy waste disposal histories overlie groundwater with elevated nitrate concentrations. The primary areas of nitrate degradation were formerly or are currently overlain by:



- Citrus (the northern parts of the Chino-North MZ)
- Dairy and irrigated agriculture (the southern parts of the Chino-North MZ, the Chino-South MZ, the Chino-East MZ, and the Prado Basin MZ [PBMZ])

Nitrate concentrations in groundwater have increased slightly or remained relatively constant in the northern parts of the Chino-North MZ from 1960 to present. These areas were formerly occupied by citrus groves and vineyards. The nitrate concentrations underlying these areas rarely exceed 10 mg/L (as nitrogen). Over the same period, nitrate concentrations increased significantly in the southern parts of the Chino-North MZ, the Chino-South MZ, the Chino-East MZ, and the PBMZ. In these areas, land use was progressively converted from irrigated/non-irrigated agricultural land to dairies, and nitrate concentrations typically exceed the 10 mg/L MCL and frequently exceed 40 mg/L.

4.3.3 Other Constituents of Potential Concern

Section 4.3.3 discusses the constituents with water quality standards that were exceeded in ten or more wells in Chino Basin with the exception of nitrate and TDS. The details of these exceedances are displayed graphically in Figures 4-4 through 4-17, and in Appendix B.

A query was developed to analyze water quality data in the Chino Basin from July 2003 through June 2008 that is in exceedance of any water quality standard. The results of this query are provided in a summary table in Appendix C, including:

- Chemical Constituents (listed alphabetically)
- Reporting Units
- Water Quality Standards (detailed explanations are provided in the table's footnote):
 - EPA Primary MCL
 - EPA Secondary MCL
 - California Primary MCL
 - California Secondary MCL
 - California Notification Level
- Minimum the minimum concentration of the given constituent for the given time period. Non-detect values were assigned a value of zero.
- Lower or First Quartile the first value that divides the items of a frequency distribution or ordered data set into four classes with each containing one fourth of the total population.
- Median or Second Quartile the second value that divides the items of a frequency distribution or ordered data set into four classes with each containing one fourth of the total population.
- Upper or Third Quartile the third value that divides the items of a frequency distribution or ordered data set into four classes with each containing one fourth of the total population.



- Maximum the maximum concentration of the given constituent for the given time period. Non-detect values were assigned a value of zero.
- Average the average concentration of the given constituent for the given time period. Non-detect values were assigned a value of zero.
- Number of Samples the total number of samples for the given constituent for the given time period.
- Number of Wells Sampled the number of wells sampled in the given time period, not the number of samples collected.
- Number of Wells with Detects the number of wells in the period wherein the constituent was detected at any concentration.
- Number of Wells with Exceedances the number of wells in the given time period with any value that exceeded any of the five water quality standards.

4.3.3.1 VOCs

The following seven VOCs were detected at or above their MCL in more than 10 wells in the Chino Basin:

- 1,1-dichloroethane (1,1-DCA)
- 1,1-dichloroethene (1,1-DCE)
- 1,2,3-trichloropropane (1,2,3-TCP)
- 1,2-dichloroethane (1,2-DCA)
- *cis*-1,2-dichloroethene (cis-1,2-DCE)
- tetrachloroethene (PCE)
- trichloroethene (TCE)

4.3.3.1.1 Trichloroethene and Tetrachloroethene

Trichloroethene (TCE) and tetrachloroethene (PCE) were/are widely used industrial solvents. Both PCE and TCE are used as metal degreasers in the automotive and other metal working industries. PCE is commonly used in the dry-cleaning industry. TCE was commonly used as a food extractant. The areal distributions of TCE and PCE are shown in Figures 4-4 and 4-5, respectively. In general, PCE is below the detection limit for wells in the Chino Basin. Wells with detectable levels tend to occur in clusters, such as those around the Milliken Landfill, south and west of the Ontario Airport, and along the margins of the Chino Hills. The spatial distribution of TCE resembles that of PCE. TCE was not detectable in most of the wells in the basin, and similar clusters of wells occur around the Milliken Landfill, south and west of Ontario International Airport (OIA), south of Chino Airport, and in the Stringfellow plume.

Figure 4-19 shows the ratio of TCE, PCE, and their breakdown products in monitoring wells associated with the VOC plumes in the southern Chino Basin. The unique characteristics of these plumes can be seen by comparing TCE and PCE concentrations and dispersion. For example, the Milliken Landfill plume and the GE plumes near Ontario Airport have significant concentrations of both TCE and PCE while the Chino Airport and Stingfellow



plumes have significant concentrations of TCE and only minor detections of PCE, and the OIA plume is characterized solely by TCE. These unique characteristics allow for differentiation between the plumes and determining the intermingling of plumes.

4.3.3.1.2 1,1-Dichloroethene, 1,2-Dichloroethane, and cis-1,2-Dichloroethene

1,1-Dichloroethene (1,1-DCE), 1,2-Dichloroethane (1,2-DCA), and cis-1,2-Dichloroethene (cis-1,2-DCE) are degradation by-products of PCE and TCE (Dragun, 1988) that are formed by reductive dehalogenation. The areal distributions of 1,1-DCE, 1,2-DCA, and cis-1,2-DCE are shown in Figures 4-6 through 4-8, respectively. 1,1-DCE, 1,2-DCA, and cis-1,2-DCE have not been detected in the majority of wells in the Chino Basin. 1,1-DCE is found near the Milliken Landfill, south and west of OIA, at the former Crown Coach Facility, and at the head of the Stringfellow plume. 1,2-DCA and cis-1,2-DCE are found in the same general locations.

4.3.3.1.3 1,1-Dichloroethane

1,1,-Dichloroethane (1,1-DCA) is a colorless oily liquid that is used as a solvent for plastics, as a degreaser, as a halon in fire extinguishers, and in the cementing of rubber, and is a degradation by-product of 1,1,1-TCA. Figure 4-9 shows the areal distribution of 1,1-DCA in the Chino Basin. Eleven wells were in exceedance of the primary CA MCL of 5 μ g/L for 1,1-DCA for the period of July 2003 through June 2008. The majority of these wells are monitoring wells at the former Crown Coach Facility.

4.3.3.1.4 1,2,3-Trichloropropane

1,2,3-TCP is a colorless liquid that is used primarily as a chemical intermediate in the production of polysulfone liquid polymers and dichloropropene, and in the synthesis of hexafluoropropylene and as a cross linking agent in the synthesis of polysulfides. It has been used as a solvent, an extractive agent, a paint and varnish remover, and a cleaning and degreasing agent, and it has been formulated with dichloropropene in the manufacturing of soil fumigants, such as D-D.

The current California State Notification Level for 1,2,3-TCP is $0.005~\mu g/L$. The adoption of the Unregulated Chemicals Monitoring Requirements regulations occurred before a method capable of achieving the required detection limit for reporting (DLR) was available. According to the CDPH, some utilities moved ahead with monitoring, and samples were analyzed using higher DLRs. Unfortunately, findings of non-detect with a DLR higher than $0.005~\mu g/L$ do not provide the CDPH with the information needed for setting a standard. New methodologies with a DLR of $0.005~\mu g/L$ have since been developed, and the CDPH has requested that any utility with 1,2,3-TCP findings of non-detect with reporting levels of $0.01~\mu g/L$ or higher do follow-up sampling using a DLR of $0.005~\mu g/L$. Because 1,2,3-TCP may be a basin-wide water quality issue, private and public wells are continuing to be retested at the lower detection limit $(0.005~\mu g/L)$.

Figure 4-10 shows the distribution of 1,2,3-TCP in Chino Basin, based on the data limitations discussed above. High 1,2,3-TCP values are associated with the Chino Airport Plume. Of particular note, there is a cluster of wells with 1,2,3-TCP concentrations greater than the Notification Level in the Jurupa region and a scattering of wells that exceed the Notification



Level on the western margins of the basin. Watermaster will continue to monitor and investigate this constituent.

4.3.3.2 Iron, Arsenic, and Vanadium

Iron, arsenic, and vanadium concentrations depend on mineral solubility, ion exchange reactions, surface complexations, and soluble ligands. These speciation and mineralization reactions, in turn, depend on pH, oxidation-reduction potential, and temperature.

4.3.3.2.1 Iron

In general, iron is not detected across the Chino Basin, but there are some scattered detectable concentrations that are above regulatory limits (see Appendix B). Iron concentrations are elevated in the vicinity of the Stringfellow Plume. Outside of the Stringfellow Plume, there were 85 wells with iron concentrations that exceed the MCL. Nevertheless, these exceedances may be an artifact of sampling methodology; relatively high concentrations of iron and trace metals are often the result of the dissolution of aluminosilicate particulate matter and colloids, which is caused by the acid preservative in unfiltered samples.

4.3.3.2.2 Arsenic

The US EPA implemented a new primary MCL for arsenic in 2006, decreasing the MCL from 50 µg/L to 10 µg/L. In November 2008, the Primary CA MCL was also changed from 50 µg/L to 10 µg/L. Figure 4-11 shows the distribution of arsenic in the Chino Basin. Eleven wells in the basin had arsenic concentrations that exceeded the MCL. Of these wells, three are associated with the Stringfellow Plume, and three are associated with Chino Airport Plume. Higher concentrations of arsenic are found in the Chino/Chino Hills area in the lower aquifer at depths greater than about 350 ft-bgs.

4.3.3.2.3 Vanadium

In the Chino Basin, vanadium has been detected above regulatory limits in some scattered wells. In groundwater, vanadium can result from mining and industrial activities or be of natural occurrence. While elemental vanadium does not occur in nature, vanadium compounds are found in fossil fuels and exist in over 60 different mineral ores. The primary industrial use of vanadium is in the steel industry where it is used to strengthen steel. Figure 4-12 shows the areal distribution of vanadium in the Chino Basin. The majority of the 25 wells in exceedance of the California Notification Level (0.05 mg/L) are associated with the Stringfellow Plume. Other exceedances are found near the Milliken Landfill, in deep wells in the Chino/Chino Hills area, and in one well near the Jurupa Mountains.

4.3.3.3 Perchlorate

Perchlorate has recently been detected in several wells in the Chino Basin (Figure 4-13), in other basins in California, and in other states in the west. The most probable reason why perchlorate was not detected in groundwater until recently is that analytical methodologies that could attain a low enough detection limit did not previously exist. Prior to 1996, the



method detection limit for perchlorate was 400 μ g/L. In March 1997, an ion chromatographic method was developed with a detection limit of 1 μ g/L and a reporting limit of 4 μ g/L.

As an environmental contaminant, perchlorate (ClO4-) originates from the solid salts of ammonium perchlorate (NH4ClO4), potassium perchlorate (KClO4), or sodium perchlorate (NaClO4). Perchlorate salts are quite soluble in water. The perchlorate anion (ClO4-) is exceedingly mobile in soil and groundwater environments. Because of its resistance to react with other available constituents, it can persist for many decades under typical groundwater and surface water conditions. Perchlorate is a kinetically stable ion, which means that reduction of the chlorine atom from a +7 oxidation state in perchlorate to a -1 oxidation state as a chloride ion requires activation energy or the presence of a catalyst to facilitate the reaction. Since perchlorate is chemically stable in the environment, natural chemical reduction is not expected to be significant.

Possible sources of perchlorate contamination are synthetic (ammonium perchlorate used in the manufacturing of solid propellant used for rockets, missiles, and fireworks) and natural (perchlorate derived from Chilean caliche that was used for fertilizer).

Fertilizers derived from Chilean caliche are currently used in small quantities on specialized crops, including tobacco, cotton, fruits, and vegetables (Renner, 1999). However, evidence suggests that usage may have been widespread for citrus crops in Southern California from the late 1800s through the 1930s.

The current CDPH Notification Level for perchlorate is 6 µg/L, which was established on March 11, 2004.

Perchlorate has been detected in 188 wells in the Chino Basin at levels greater than 6 μ g/L. Perchlorate Notification Level exceedances occur in the following areas of the Chino Basin (Figure 4-13):

- Rialto-Colton Basin (There is a significant perchlorate plume in the Rialto-Colton Basin. The RWQCB is investigating the source of this plume, which appears to be near the Mid-Valley Sanitary Landfill. According to the RWQCB, several companies—including B.F. Goodrich, Kwikset Locks, American Promotional Events, and Denova Environmental—operated nearby and used or produced perchlorate. These companies were located on a 160-acre parcel at T1N R5W S21 SW1/4. Denova Environmental also operated on a 10-acre lot at T1N R5W S20 S1/2 (along the boundary between Sections 20 and 29). Perchlorate in the Fontana area of Chino Basin may be the result of (i) the Rialto-Colton perchlorate plume migrating across the Rialto-Colton fault, (ii) other point sources in Chino Basin, and/or (iii) the non-point application of Chilean nitrate fertilizer in citrus groves.)
- Downgradient of the Stringfellow Superfund Site (Concentrations have exceeded 600,000 µg/L at onsite observation wells. The plume has likely reached the Pedley Hills and may extend as far as Limonite Avenue.)
- City of Pomona well field (source[s] unknown)
- Wells in the City of Ontario water service area, south of OIA (source[s] unknown)
- Scattered wells in the Monte Vista water service area (source[s] unknown)



• Scattered wells in the City of Chino water service area (source[s] unknown)

A forensic isotope study was conducted to determine the source of perchlorate in Chino Basin groundwater. This forensic technique was developed using comprehensive stable isotope analyses (37Cl/35Cl and 18O/17O/16O) of perchlorate to determine the origin of the perchlorate (synthetic vs. naturally occurring). Stable isotope analyses of perchlorate from known man-made (e.g. samples derived from electrochemically synthesized ammonium- and potassium-perchlorate salts) and natural (e.g. samples from the nitrate salt deposits of the Atacama Desert in Chile) sources reveal systematic differences in isotopic characteristics that are related to the formation mechanisms (Bao & Gu, 2004; Böhlke et al., 2005; Sturchio et al., 2006). There is considerable anecdotal evidence that large quantities of Chilean nitrate fertilizer were imported into the Chino Basin in the early 1900s for the citrus industry, which covered the north, west and central portions of the basin.

The perchlorate isotope study consisted of 10 groundwater samples that were collected throughout the Chino Basin. The sampling points included private wells and municipal production wells. Samples were collected using a flow-through column with a highly perchlorate-selective anion-exchange resin. The exchange resin concentrates low levels of perchlorate in groundwater such that a sufficient amount can be acquired and for isotopic analysis. Results confirmed that most of the perchlorate in the west and central portions of the Chino Basin was derived from Chilean nitrate fertilizer. One sample collected south of the OIA is a potential mixture of natural and synthetic sources.

4.3.3.4 Total Chromium and Hexavalent Chromium

Figure 4-14 shows the areal distribution of total chromium in the Chino Basin. Thirty wells were found to be in exceedance of the CA MCL of 50 $\mu g/L$. The majority of these wells are associated with the Milliken Sanitary Landfill, the Stringfellow Plume, and the GE Test Cell Plume. The remaining wells include isolated wells near the Jurupa Mountains and in the southern Chino Basin and City of Pomona wells. Chromium in groundwater results from natural and anthropogenic sources.

Hexavalent chromium is currently regulated under the MCL for total chromium. In 1999, the CDPH identified that hexavalent chromium needed an individual MCL, and concerns over its carcinogenicity grew. Subsequently, the CDPH included it on the list of unregulated chemicals that require monitoring. California Health and Safety Codes (§116365.5 and §1163659a) compelled the adoption of a hexavalent chromium MCL by January 1, 2004, and required it to be close to the public health goals (PHG) established by the Cal/EPA Office of Environmental Health Hazard Assessment (OEHHA). At present, the PHG has not been established, and the CDPH cannot proceed with the MCL process. Figure 4-15 shows the areal distribution of hexavalent chromium in the Chino Basin. Only three wells in the Chino Basin were in exceedence of the CA MCL for total chromium. In the near future hexavalent chromium may become a more significant contaminant of concern in the Chino Basin when a lower MCL is determined by CDPH, and more wells are sampled for hexavalent chromium.

4.3.3.5 Chloride and Sulfate



Chloride and sulfate both exceeded secondary MCLs. As discussed previously, secondary MCLs apply to chemicals in drinking water that adversely affect its aesthetic qualities and are not based on the direct health effects associated with the chemical. Chloride and sulfate are major anions associated with TDS. All wells in the basin had detectable levels of sulfate (Figure 4-16), but most had concentrations that were less then 125 mg/L (one-half the water quality standard). A total of 41 wells had concentrations at or above the sulfate secondary MCL. In general, these wells are distributed in the southern portion of the basin, in the Stringfellow plume, and along the margins of the Chino Hills. All wells had detectable levels of chloride (Figure 4-17), but most had concentrations that were less 125 mg/L (one-half the MCL). The secondary MCL for chloride was exceeded in 25 wells; almost all of which are located in the southern portion of the basin.

4.3.3.6 Color, Odor, and Turbidity

In the last 5 years, color, odor, and turbidity have been detected above their secondary MCLs in more than 10 wells within the Chino Basin (see Appendix B). These parameters are monitored purely for aesthetic reasons and should not substantially impair water quality in the Chino Basin.

4.3.4 Point Sources of Concern

The water quality discussion above described water quality conditions across the entire basin. The discussion below describes the water quality plumes associated with known point source discharges to groundwater. Figure 4-18 shows the locations of various point sources and associated areas of water quality degradation. Figure 4-19 shows the VOC plumes and features pie charts that display the relative percent of TCE, PCE, and other VOCs detected at groundwater wells within plume impacted areas. The pie charts demonstrate the chemical differentiation between the VOC plumes in the southern portion of Chino Basin.

4.3.4.1 Alumax Aluminum Recycling Facility

Between 1957 and 1982, an 18-acre aluminum recovery facility was operated in the City of Fontana. The byproducts of aluminum recycling are aluminum oxide wastes and brine water. During this 25-year period, solid wastes were stockpiled onsite. Process water containing sodium and potassium chloride salts was discharged onsite and allowed to percolate into native soil and groundwater. Discharge ceased in 1982, and the solid wastes were removed in 1992. Onsite groundwater monitoring was initiated in 1993 by then owner Alumax, Inc. The site was subsequently capped to prevent the future mobilization of salts offsite. Alcoa Davenport Works (Alcoa) purchased Alumax in 1998.

Currently, there are two onsite monitoring wells: MW-1 is located in the northeast corner of the property, and MW-2 is located in the southwest corner. These wells have steel casings and have experienced chloride corrosion and extensive accumulation of iron hydroxide scale. Rehabilitation efforts in 2001 failed to adequately clear the well screens. Both wells subsequently experienced partial casing constrictions or screen collapses. In 2007, it was discovered that over ten feet of iron oxide scale and sediment had accumulated in the bottom



of MW-1. MW-2 was abandoned and replaced in 2008 as it could no longer be sampled.

Offsite monitoring began with the construction of four monitoring wells (AOS-1, AOS-2, AOS-3, and AOS-4) between 1999 and 2000. These wells are all located downgradient of the site and were constructed of PVC in an effort to avoid the scale and corrosion experienced at the onsite wells. In April 2008, the RWQCB stated that Alcoa would no longer be required to monitor offsite monitoring wells AOS-1, AOS-2, and AOS-3 unless elevated levels of salts were detected at upgradient well AOS-4 (RWQCB, 2008). Alcoa is currently evaluating the ownership transfer of wells AOS-1, AOS-2, and AOS-3 to Watermaster to allow for continued monitoring.

The plume emanating from the Alumax site is characterized by elevated concentrations of sulfate, nitrate, chloride, potassium, and sodium. Consequently, the TDS concentrations at the onsite wells are high, ranging from about 500 mg/L to over 2,000 mg/L. Offsite monitoring has yielded observed TDS concentrations that range from about 100 mg/L to 700 mg/L. Note that these TDS values are higher than those observed at up-gradient wells, which typically range from 200 to 300 mg/L.

4.3.4.2 Chino Airport

The Chino Airport is located approximately four miles east of the City of Chino and six miles south of the OIA and occupies about 895 acres. From the early 1940s until 1948, the airport was owned by the federal government and used for flight training and aircraft storage. The County of San Bernardino acquired the airport in 1948 and has operated and/or leased portions of the facility ever since. Since 1948, businesses and activities at the airport have included: the modification of military aircraft; crop-dusting; aircraft-engine repair; aircraft painting, stripping, and washing; dispensing of fire-retardant chemicals to fight forest fires; and general aircraft maintenance. The use of organic solvents for various manufacturing and industrial purposes has been widespread throughout the airport's history. From 1986 to 1988, a number of groundwater quality investigations were performed in the vicinity of the Chino Airport. Analytical results from groundwater sampling revealed the presence of VOCs above MCLs in six wells downgradient of the Chino Airport. The most common VOC detected above its MCL is TCE, as shown in Figure 4-19. TCE concentrations in the contaminated wells ranged from 6 to 75 µg/L.

In 1990, Cleanup and Abatement Order (CAO) No. 90-134 was issued to address groundwater contamination emanating from the Chino Airport. During 2003, five groundwater monitoring wells were installed onsite; and in 2005, an additional four groundwater monitoring wells were installed onsite for further characterization. During June and July of 2006, Watermaster conducted a focused sampling event of 25 wells within the vicinity of the Chino Airport plume. In 2007, the San Bernardino County Department of Airports began to focus their investigation on offsite characterization of the plume. In 2008, the RWQCB issued a CAO (No. R-8 2008-0064) to the San Bernardino County Department of Airports in order to define the lateral and vertical extent of the VOCs in groundwater and to prepare a remedial action plan. In late 2008, nine offsite monitoring wells were completed in three locations. Initial sampling of these wells was done in August 2009.

Figure 4-18 shows the approximate areal extent of TCE in groundwater at concentrations in



exceedance of the MCL in the vicinity of the Chino Airport as of 2008. The plume is elongate in shape, up to 3,600 feet wide, and extends approximately 12,100 feet from the airport's northern boundary in a south to southwestern direction. From July 2003 to June 2008, the maximum TCE concentration detected at an individual well within the Chino Airport plume was $910 \, \mu g/L$.

4.3.4.3 California Institute for Men

The California Institution for Men (CIM) is a state correctional facility located in the City of Chino and has been in existence since 1939. The property occupies approximately 1,500 acres, and is bounded by Eucalyptus Avenue to the north, Euclid Avenue to the east, Kimball Avenue to the south, and Central Avenue to the west. Site use includes agricultural operations, inmate housing, and correctional facilities. The Heman G. Stark Youth Correctional Facility occupies the eastern portion of the property (Geomatrix Consultants, 2005).

In 1990, PCE was detected at a concentration of 26 μg/L at CIM drinking water supply Well 1. Analytical results have indicated that the most common VOCs detected in groundwater underlying CIM are PCE and TCE. The maximum PCE concentration in groundwater detected at an individual monitoring well (MW-7) was 1990 μg/L, and the maximum TCE concentration in groundwater detected at an individual monitoring well (MW-6) was 160 μg/L (Geomatrix Consultants, 2007). Other detected VOCs include 1,2-DCE, bromodichloromethane, 1,1,1-TCA, carbon tetrachloride, chloroform, and toluene.

In 1992, construction began on a groundwater monitoring network of approximately 40 wells. These wells were sampled intermittently through 2007. An Interim Remedial Measure (IRM) was implemented to resume production at Well 1, treat extracted water to reduce VOC concentrations, and use that water as part of the CIM potable water distribution system. Since the implementation of the IRM, the concentrations of PCE and TCE in groundwater have decreased considerably. Of the 39 wells sampled in 2007, 6 wells in the shallow aquifer had PCE concentrations in exceedance of the MCL, and TCE was detected at one shallow monitoring well (Geomatrix Consultants, 2007). CIM submitted a Request for No Further Action (NFA) for groundwater PCE remediation to the RWQCB.

Figure 4-18 shows the approximate areal extent of VOCs in groundwater at concentrations exceeding their MCLs as of 2008. The plume is up to 2,900 feet wide and extends about 5,800 feet from north to south. As Figure 4-19 illustrates, the CIM plume is primarily characterized by PCE. From July 2003 to June 2008, the maximum PCE and TCE concentrations in groundwater detected at an individual well within the CIM plume were $57 \,\mu g/L$ and $26 \,\mu g/L$, respectively.

4.3.4.4 Crown Coach

The former Crown Coach site, located at 13799 Monte Vista Ave in the City of Chino, was used by the General Electric Corporation (GE) for the manufacturing and maintenance of semi-tractors and buses from the early 1970s onward. In 1987, it was discovered that twelve underground storage tanks were leaking lube oils, diesel, antifreeze, waste oil, and waste



solvents. All 12 tanks were removed by 1988, and the release of spent solvents in the underlying soil and groundwater was reported (Rosengarten Smith & Associates, 1992). Since 1988, sampling at 22 monitoring wells has determined the concentration and areal extent of the VOC plume. Contaminated soil and groundwater are contained onsite. The most common VOCs detected are TCE, PCE, and 1,1-DCE, as shown in Figure 4-19.

Concurrent with groundwater monitoring, a series of remediation activities have occurred on the property. Starting in June 1990, extracted groundwater was discharged to an onsite sewer connection, operating under an industrial wastewater discharge permit. A soil-vapor extraction system was brought onsite in 1992 to address vadose zone contamination. Starting in 2005, a Dual Phase Extraction Treatment System (DPETS) was used to remediate groundwater and soil. In May 2008, Duke Reality began redevelopment activities on the property. During construction, DPETS operations ceased, and Edible Oil Solution (EOS) was injected into ten monitoring and extraction wells as a remediation replacement.

Figure 4-18 shows the approximate areal extent of VOCs in groundwater at concentrations exceeding their MCLs near the Crown Coach Facility as of 2008. The plume is approximately 500 feet in length and 250 feet wide. The last monitoring event in 2008 indicated that the lateral boundaries of the plume are decreasing, and PCE, TCE, and 1,1 DCE were not detected in deep aquifer wells (Rosengarten Smith & Associates, 2008). From July 2003 to June 2008, the maximum PCE and TCE concentrations detected at an individual well within the Crown Coach VOC plume were 182 μg/L and 125 μg/L, respectively.

In June 2009, GE submitted a report to the Regional Board evaluating the effectiveness of the EOS injections and the need for additional remedial measures. In this report GE concluded that the hydrogeologic conditions beneath the site are sufficient to protect the beneficial uses of groundwater in the regional aquifer and that no further monitoring and remediation activity is warranted at this site. A response from the Regional Board on this report is pending.

4.3.4.5 General Electric Flatiron Facility

The General Electric Flatiron Facility (Flatiron Facility) occupied the site at 234 East Main Street, Ontario, California from the early 1900s to 1982. Its operations primarily consisted of manufacturing clothes irons. Currently, the site is occupied by an industrial park. The RWQCB issued an investigative order to GE in 1987 after an inactive well in the City of Ontario was found to contain TCE and chromium above drinking water standards. Analytical results from groundwater sampling have indicated that VOCs and total chromium are the major groundwater contaminants. The most common VOC detected at levels significantly above its MCL is TCE, as shown in Figure 4-19. TCE has reached a measured maximum concentration of 5,620 µg/L. Other VOCs—including PCE, toluene, and total xylenes—are periodically detected but commonly below their MCLs (Geomatrix Consultants, 1997).

The facility's eighteen monitoring wells are part of a quarterly monitoring program that began in 1991. Remediation activities began in 1995 with RWQCB Waster Discharge Requirement Order No. 95-62 for the pump and treat of groundwater at two extraction wells, EW-01 and EW-02. The operation of the extraction wells and remediation system is also referred to as the Final Remediation Measures (FRM). Groundwater from EW-01 is treated for VOCs, and groundwater from EW-02 is treated for VOCs and chromium. The two sources of treated



water join, are pipelined to the West Cucamonga Channel and ultimately to the Ely Basins, where it percolates into the Chino Basin Aquifer. In late 2009 or early 2010, an injection well and pipeline will be completed, and treated groundwater will be injected into the Chino Basin. In addition to the remediation measures discussed above, a Soil Vapor Extraction (SVE) system has been in operation since 2003 to remove VOCs from impacted soil.

Figure 4-18 shows the approximate areal extent of TCE in groundwater at concentrations exceeding the MCL as of 2008. The plume is up to 3,400 feet wide and extends about 9,000 feet south-southwest (hydraulically downgradient) from the southern border of the site. From July 2003 to June 2008, the maximum TCE concentration detected at an individual well within the Flatiron Facility plume was 5,620 µg/L, and the maximum total chromium concentration detected at an individual well was 485 µg/L.

4.3.4.6 General Electric Test Cell Facility

The GE Engine Maintenance Center Test Cell Facility (Test Cell Facility) is located at 1923 East Avion, Ontario, California. From 1956 to present, primary operations at the Test Cell Facility have included the testing and maintenance of commercial and military aircraft engines. Historically, hazardous waste was disposed of in dry wells. In 1987, results of a preliminary investigation indicated the presence of VOCs in soils near the dry wells. In 1991, a soil and groundwater investigation and subsequent quarterly groundwater quality monitoring showed the presence of VOCs in the soil and groundwater beneath the Test Cell Facility and that the VOCs had migrated offsite (Dames & Moore, 1996). Subsequent investigations indicated that the most common and abundant VOC detected in groundwater beneath the site was TCE. The historical maximum TCE concentration measured at an onsite monitoring well (directly beneath the Test Cell Facility) was 1,240 µg/L. The historical maximum TCE concentration measured at an offsite monitoring well (downgradient) was 190 µg/L 1997). Other detected VOCs include (BDM International, PCE, cis-1,2-DCE, 1,2-dicholoropropane, 1,1-DCE, 1,1-DCA, and chloroform, among others.

A Consent Order between General Electric and CDPH was signed September 28, 1988 for groundwater and soil remediation (Docket No. 88/89-009CO). The groundwater investigation and cleanup is under the oversight of the RWQCB. Vapor extraction treatment system operations began in 1996 (Docket No. HAS 97/98-014). Quarterly monitoring and operations status reports have been submitted to the DTSC and the RWQCB since remediation commenced. Recently a study was conducted to evaluate the effectiveness of the soil remediation program. The results of this study were submitted to the DTSC in October 2008 (Geosyntec Consultants, 2008). In some regions of the facility, shallow soils have reached acceptable closure levels; however, remediation activities will continue until sufficient data can be evaluated.

Figure 4-18 shows the approximate areal extent of VOCs in groundwater at concentrations exceeding federal MCLs as of 2008. The plume is elongate in shape, up to 2,400 feet wide, and extends approximately 10,300 feet from the Test Cell Facility in a southwesterly direction. As Figure 4-19 illustrates, the GE Test Cell Facility plume is characterized primarily by TCE, PCE, cis-1,2-DCE, and 1,1-DCE. From July 2003 to June 2008, the maximum TCE and PCE concentrations in groundwater detected at an individual well within the Test Cell Facility



plume were 900 µg/L and 16 µg/L, respectively.

4.3.4.7 Kaiser Steel Fontana Steel Site

Between 1943 and 1983, the Kaiser Steel Corporation (Kaiser) operated an integrated steel manufacturing facility in Fontana. During the first 30 years of operations (1945-1974), a portion of the Kaiser brine wastewater was discharged to surface impoundments and allowed to percolate into the soil. In the early 1970s, the surface impoundments were lined to eliminate percolation to groundwater (Wildermuth, 1991). In July of 1983, Kaiser initiated a groundwater investigation that revealed the presence of a plume of degraded groundwater beneath the facility. In August 1987, the RWQCB issued CAO Number 87-121, requiring additional groundwater investigations and remediation activities. The results of those investigations showed that the major constituents of release to groundwater were inorganic dissolved solids and low molecular weight organic compounds. The wells sampled during the groundwater investigations had TDS concentrations ranging from 500 to 1,200 mg/L and TOC concentrations ranging from 1 to 70 mg/L. By November 1991, the plume had migrated almost entirely off the Kaiser site.

In 1993, Kaiser and the RWQCB entered into a settlement agreement; Kaiser was required to mitigate any adverse impacts caused by its plume at existing and otherwise useable municipal wells. Pursuant to the settlement, the RWQCB rescinded its earlier order 91-40, and Kaiser was granted capacity in the Chino II Desalter to intercept and remediate the Kaiser plume within the Chino Basin. In an effort to further characterize the plume, during 2005, a network of 22 public and private supply wells were selected for quarterly groundwater sampling for one year and annual sampling thereafter. In addition, two triple nested monitoring wells, MZ3-1 and MZ3-2, were installed between the distal edge of the plume and municipal supply wells in 2007. Well MZ3-1/3 was found to have elevated concentrations of TDS, sulfate, and TOC. Based on this finding, the Kaiser plume was extended to include this well.

Figure 4-18 shows the approximate areal extent of the TDS/TOC groundwater plume as of 2008. Based on a limited number of wells, including Kaiser monitoring wells MP-2 and KOSF, City of Ontario Wells 27 and 30, and monitoring wells MZ3-1 and MZ3-2, the plume is up to 7,000 feet wide and extends about 18,500 feet from the northeast to the southwest.

4.3.4.8 Milliken Sanitary Landfill

The Milliken Sanitary Landfill (MSL) is an inactive Class III Municipal Solid Waste Management Unit, located near the intersections of Milliken Avenue and Mission Boulevard in the City of Ontario. This facility is owned by the County of San Bernardino and managed by the County's Waste System Division. The facility operated from 1958 to 1999. Groundwater monitoring at the MSL began in 1987 with five monitoring wells as part of a Solid Waste Assessment Test (SWAT) investigation (IT, 1989). The results of this investigation indicated that the MSL had released organic and inorganic compounds to underlying groundwater. Based on this finding, the MSL conducted an Evaluation Monitoring Program (EMP) investigation. At the completion of the EMP, a total of 29 monitoring wells were drilled to evaluate the nature and extent of the groundwater impacts identified in the vicinity of the MSL (GeoLogic Associates, 1998). Analytical results have indicated that VOCs



are the major constituents of release. The most commonly detected VOCs are TCE, PCE, and dichlorodifluoromethane. Other VOCs that have been detected above MCLs include vinyl chloride, benzene, 1,1-dichloroethane, and 1,2-dichloropropane. Historically, the maximum total VOC concentration in an individual monitoring well was $159.6\,\mu\text{g/L}$ (GeoLogic Associates, 1998).

Figure 4-18 shows the approximate areal extent of VOCs in groundwater at concentrations exceeding MCLs as of 2008. The plume is up to 1,800 feet wide and extends about 2,100 feet south of the MSL's southern border. As Figure 4-19 illustrates, the MSL plume is characterized by a mixture of PCE, TCE, and their degradation products. From July 2003 to June 2008, the maximum TCE and PCE concentrations detected at an individual well within the MSL plume were $12 \,\mu g/L$ and $8.4 \,\mu g/L$, respectively.

4.3.4.9 Municipal Wastewater Disposal Ponds

Historically, treated municipal wastewater was disposed of in ponds located near the current IEUA Regional Plant 1 (RP1), located in south Ontario, and the former Regional Plant 3 (RP3) disposal ponds, located in south Fontana. The ponds located just east of RP1, commonly referred to as the Cucamonga ponds, were used to dispose of untreated effluent collected by the Cucamonga County Water District (now the CVWD) and the IEUA. The RP3 disposal ponds are located on the southwest corner of Beech and Jurupa Avenues in the City of Fontana. The discharge of treated wastewater to the Cucamonga ponds and the RP3 ponds ceased between the early 1970s and the mid-1980s. The contaminant plumes emanating from these ponds have never been characterized.

4.3.4.10 Upland Sanitary Landfill

The Upland Sanitary Landfill (USL) is located on the site of a former gravel quarry at the southeastern corner of 15th Street and Campus Avenue in the City of Upland. The facility operated from 1950 to 1979 as an unlined Class II and Class III municipal solid waste disposal site. In 1982, the entire USL disposal site was covered with a 10-inch thick, low permeability layer of sandy silt (GeoLogic Associates, 1997). Groundwater monitoring began at the USL in 1988, and there are now three onsite monitoring wells: an upgradient well, a cross-gradient well, and a downgradient well (City of Upland, 1998). Monitoring results indicate that the USL and inorganic compounds organic to underlying (GeoLogic Associates, 1997). Groundwater samples from the downgradient monitoring well consistently contain higher concentrations of organic and inorganic compounds than samples from the upgradient and cross-gradient wells. Historical groundwater samples have indicated that VOCs are the major constituents of release, and all three monitoring wells have shown detectable levels of VOCs. The most common VOCs detected above MCLs are dichlorodifluoromethane, PCE, TCE, and vinyl chloride. Other VOCs that have been periodically detected above MCLs include methylene chloride, cis-1,2-DCE, 1,1-DCA, and benzene. For the 1990 to 1995 period, the average total VOC concentration at the downgradient monitoring well was 125 µg/L (GeoLogic Associates, 1997). And, for the July 2003 to June 2008 period, the maximum TCE and PCE concentrations detected at USL monitoring wells were 0.6 µg/L and 3.5 µg/L, respectively.



Figure 4-18 shows the approximate areal extent of VOCs at concentrations exceeding MCLs as of 2008. Please note that this plume is only defined by three onsite monitoring wells. The extent of the plume may be greater than currently depicted in Figure 4-18.

4.3.4.11 VOC Plume - South of the OIA

A VOC plume, containing TCE, exists south of the OIA. This plume extends approximately from State Route 60 on the north and Haven Avenue on the east to Cloverdale Road on the south and South Grove Avenue on the west. It is up to 11,300 feet wide and 20,500 feet long. By the late 1980s, the RWQCB determined TCE was present in numerous private wells in the area south of the OIA, and identified past activities at the airport as a likely source of TCE (RWQCB, 2005b). By 2005, TCE in exceedance of the CA MCL (5µg/L) was detected in 92 of the 167 private wells in the area. In July 2005, Draft CAOs were issued by the RWQCB to six parties identified as former TCE dischargers on the OIA property: Aerojet, the Boeing Company (Boeing), the Department of Defense, the Lockheed Martin Corporation (Lockheed), and the Northrop Grumman Corporation (Northrop). On a voluntary basis, Lockheed, GE, Boeing, and Aerojet are funding current investigative work on the extent and source of the TCE plume. Three triple nested monitoring wells were constructed in 2008 between the OIA and the VOC plume. A fourth well will be completed in 2009.

Final CAOs will likely be issued in the future. Watermaster has been working closely with the RWQCB and the identified parties, providing any available information to assist in the investigation. Remediation of the plume will likely be achieved using the CDA's Chino Basin Desalter I facilities . Watermaster is currently seeking a settlement with the companies to recover treatment costs associated with the VOC plume.

Figure 4-18 shows the approximate areal extent of the plume as of 2008. As Figure 4-19 illustrates, the OIA plume is characterized solely by TCE. During the July 2003 to June 2008 period, the maximum TCE concentration detected at an individual well within this plume was $38 \,\mu g/L$.

4.3.4.12 Stringfellow NPL Site

One facility in the Chino Basin, the Stringfellow site, is on the current NPL of Superfund Sites. This site is located in Pyrite Canyon north of Highway 60 near the community of Glen Avon in Riverside County (see Figure 4-18). From 1956 until 1972, this 17-acre site was operated as a hazardous waste disposal facility. More than 34-million gallons of industrial waste—primarily from metal finishing, electroplating, and pesticide production—were deposited at the site (US EPA, 2001). A groundwater plume of site-related contaminants exists underneath portions of the Glen Avon area. Groundwater at the site contains various VOCs, perchlorate, NDMA, and trace metals, such as cadmium, nickel, chromium, and manganese. In the original disposal area, soil is contaminated with pesticides, polychlorinated biphenyls (PCBs), sulfates, perchlorate, and trace metals. The original disposal area is covered by a clay cap, fenced, and guarded by security services.



Contamination at the Stringfellow site has been addressed by cleanup remedies described in four EPA RODs. Since 1986, cleanup actions have focused on controlling the source of contamination, installing an onsite pretreatment plant, the cleanup of the lower part of Pyrite Canyon, and the cleanup of the community groundwater area below Highway 60. In 1996, the DTSC assumed responsibility for the maintenance of the Stringfellow Superfund Site through a Cooperative Agreement with the USEPA. In December 2007, the DTSC submitted the Draft Final Supplemental Feasibility Study (SFS), which identified and evaluated the final remedial alternatives for cleanup. The 2007 Draft SFS is a revised version of an earlier 2000 draft; reconsideration was required after perchlorate and other new contaminates were discovered in 2001. Once finalized, the SFS will be used by the US EPA to select a final remedial strategy and prepare a draft ROD. The draft ROD is anticipated in December 2009.

Figure 4-18 shows the approximate areal extent of the Stringfellow VOC plume as of 2008. The VOC plume is elongate in shape, up to 1,500 feet wide, and extends approximately 14,500 feet from the original disposal area in a southwesterly direction. The most common VOC detected at levels above the MCL is TCE. There are approximately 70 extraction wells throughout the length of the plume, which have been effective in stopping plume migration and removing TCE contamination. South of Highway 60, there are only a few isolated areas where TCE exceeds 5 μg/L (DTSC, 2008). During the 2003 to 2008 period, the maximum TCE concentration detected in the Stringfellow plume was 170 μg/L.

High levels of perchlorate associated with the Stringfellow site were detected in community groundwater south of Highway 60 in 2001. Residents connected to the JCSD water service were provided bottled water, and the DTSC contracted to install water mains and hook ups at each residence. Concurrent with the SFS, the DTSC is conducting a Remedial Investigation and Feasibility Study of remedial alternatives for perchlorate in the downgradient community area. As with TCE, the operation of the groundwater treatment system has resulted in a reduction of perchlorate. Since the discovery in 2001, perchlorate concentrations have been reduced by 30% to 50% throughout the monitored area (DTSC, 2008). Figure 4-18 shows the approximate areal extent of perchlorate concentrations exceeding the Notification Level (6 μg/L) as of 2008. The perchlorate plume is elongated in shape, up to 2,000 feet wide, and extends approximately 25,000 feet to the southwest from the original disposal area. During the 2003 to 2008 period, the maximum perchlorate concentration detected in the Stringfellow plume was 870 μg/L.

4.3.5 Water Quality by Management Zone

Figure 4-20 shows the locations of wells with groundwater quality time histories discussed herein and the five Chino Basin management zone boundaries. Wells were selected based on length of record, completeness of record, quality of data, and geographical distribution. Wells are identified by their local name (usually owner abbreviation and well number) or their X Reference ID (X Ref ID) if privately owned. The HCMP wells were selected because they are sampled at multiple depths and have a consistent water quality record for the past four years. Figures 4-21 through 4-28 are TDS and NO3-N time histories for the wells shown in Figure 4-20 from 1970 to 2008. These time histories illustrate water quality variation and trends within each management zone and the current state of water quality compared to



historical trends.

4.3.5.1 Management Zone 1

MZ1 is an elongate region in the westernmost part of the Chino Basin. Figures 4-21 and 4-22 show TDS and NO3-N time histories for three wells representative of the northern portion of MZ1 (City of Upland well 8 [Upland 08], Monte Vista Water District well 5 [MVWD 05], and City of Upland well 20 [Upland 20]), two wells representative of the central region (City of Chino 5 [Chino 05] and City of Pomona well 23 [Pomona 23]), and two wells representative of the southern portion (Chino Institution for Men well 13 [CIM 13] and HCMP 3). In the northern portion of MZ1, NO3-N and TDS values have remained steady or decreased slightly over the time period depicted. Upland 08 exhibits NO3-N concentrations above the MCL (10 mg/L); however, slightly towards the west, near the Upland, Montclair, and College Heights Recharge Basins, NO3-N values drop below the MCL, as demonstrated by MVWD 05. TDS levels also decrease near the recharge basins. In the central region of MZ1, TDS and NO3-N concentrations have increased slightly over the last 30 years, but they are still below the MCLs. In the southern portion, NO3-N and TDS concentrations have increased significantly since 1990 and are above the MCLs, which is the trend seen in the majority of wells south of Highway 60. Quarterly sampling at HCMP 3 shows that TDS and NO3-N concentrations have remained stable over the past four years. HCMP 3 also shows the variation of water quality from the shallow to deeper aquifers. Overall, NO3-N and TDS concentrations in MZ1 escalate from north to south but have not increased over the last five years.

4.3.5.2 Management Zone 2

MZ2 is an elongate region in the center part of the Chino Basin. Figures 4-23 and 4-24 show TDS and NO3-N time histories for two wells representative of the northern portion of MZ2 (CVWD Well 5 [CVWD 05] and City of Ontario well 24 [ONT 24]), one well representative of the central region (City of Ontario well 17 [ONT 17]), and three wells representative of the southern portion (X Ref 29, HCMP 1, and X Ref 5333). Similar to MZ1, NO3-N and TDS values increase from north to south. Over the time period depicted, NO3-N and TDS concentrations have remained stable in the northern portion of MZ2, increased slightly in the central region, and increased considerably in the southern portion. At X Ref 5333 and HCMP 1, in the southern portion of MZ2, TDS concentrations are currently greater than twice the MCL (500 mg/L), and NO3-N concentrations are twice the MCL (10mg/L) or greater. In addition, HCMP 1 exemplifies the variation of high TDS and NO3-N levels in the shallow aquifer and low levels in the deeper aquifer. Overall, NO3-N and TDS concentrations have not increased over the last five years with the exception well X Ref 5333.

4.3.5.3 Management Zone 3

MZ3 is an elongate region that borders the majority of the Chino Basin's eastern boundary. Figures 4-25 and 4-26 show TDS and NO3-N time histories for one well representative of the northern portion (City of Fontana 37A [F37A]), one well representative of the central region (City of Ontario well 31 [ONT 31]), and two wells representative of the southern portion (Jurupa Community Service District well 16 [JCSD 16], and X Ref 5736). Similar to MZ1 and



MZ2, NO3-N and TDS values increase from north to south. In the northern and central areas of MZ3, TDS values have slightly increased since 1980 but still remain below the MCL (500 mg/L). Over the time period depicted, NO3-N concentrations increase in all regions of MZ3. Well F37A, in the northern region, exhibits NO3-N concentrations slightly above the MCL (10 mg/L). In the southern portion of MZ3, current TDS and NO3-N concentrations are near double the MCLs. At JCSD 16, NO3-N and TDS concentrations have increased significantly since 1990. In general, NO3-N and TDS concentrations have not increased over the last five years.

4.3.5.4 Management Zone 4

MZ4 – also known as Chino-East – is a wedge shaped region, bounded by the Jurupa Hills to the northeast, the Pedley Hills to the southeast, Management Zone 5 to the south, and Management Zone 3 to the west. Figures 4-27 and 4-28 show TDS and NO3-N time-histories for one well representative of the western region (HCMP-9), one well representative of the northern region (Jurupa Community Service District Well 24 [JCSD 24]), and one well representative of the eastern region (CDPH Stringfellow monitoring well [CTP-TW1]). In the western portion of MZ4, at HCMP-9, TDS and NO3-N concentrations are above the MCLs in the shallow aquifer but quite low in the deeper aquifer. The TDS and NO3 concentrations at JCSD 24 are slightly lower than those in the western portion, but they are slightly below or equal to the MCLs. In the eastern portion, at CTP-TW1, TDS and NO3-N concentrations are significantly above the MCLs. High TDS and NO3-N concentrations in the eastern portion of MZ4 are predominantly associated with the Stringfellow plume. Pre-1990 water quality data was not available for wells in this region. Since 1990, MZ4 TDS and NO3-N levels have remained relatively stable and decreased slightly over the last few years.

4.3.5.5 Management Zone 5

MZ5 – also known as Chino-South – is a small region towards the southeastern boundary of the Chino Basin. It is bordered by MZ4 to the north and MZ3 to the east. Figures 4-27 and 4-28 show TDS and NO3-N time histories for three wells representative of the northern portion of MZ5 (San Ana River Water Company Well 1A [SARWC 01A], JCSD 01, and HCMP-8). None of the wells in the southern region of MZ5 have sampling records that are complete enough to be considered representative. At JCSD 01 and SARWC 01A, TDS concentrations have historically been above the MCL (500 mg/L) and began to notably increase in 1990. Starting in 1995, NO3-N concentrations at JCSD 01 and SARWC 01A began to increase slightly above the MCL. Water quality sampling at these two wells ceased around 2005; however, HCMP-8 shows that TDS and NO3-N concentrations have decreased significantly since then.

4.3.6 Current State of Groundwater Quality in Chino Basin

The groundwater quality in Chino Basin is generally very good with better groundwater quality found in the north where recharge occurs. In the southern portion of the basin, TDS and NO3-N concentrations increase. Between July 2003 and June 2008, 32 percent of the wells sampled south of Highway 60 had TDS concentrations below the secondary MCL, an



improvement from the 20 percent reported in the 2006 State of the Basin Report (period of July 2001 through June 2006). In some places, wells with low TDS concentrations are proximate to wells with higher TDS concentrations, suggesting a vertical stratification of water quality. Between July 2003 and June 2008, about 69 percent of the wells sampled south of Highway 60 had NO3-N concentrations greater than the MCL, an improvement from the 80 percent reported in the 2006 State of the Basin Report (period of July 2001 through June 2006). However, please note that these statistical improvements may be an artifact of sampling occurrence and frequency.

Other constituents that impact groundwater quality from a regulatory or Basin Plan standpoint include certain VOCs, arsenic, and perchlorate. As discussed in Section 4.3.4, there are a number of point source releases of VOCs in the Chino Basin that are in various stages of investigation or cleanup. There are also known point source releases of perchlorate (MVSL area, Stringfellow, etc.), and non-point source related perchlorate contamination appears to have resulted from natural and anthropogenic sources. Arsenic at levels above the WQS appears to be limited to the deeper aquifer zone near the City of Chino Hills. Hexavalent chromium, while not currently a groundwater quality issue in the Chino Basin, may become so, depending on the promulgation of future standards.

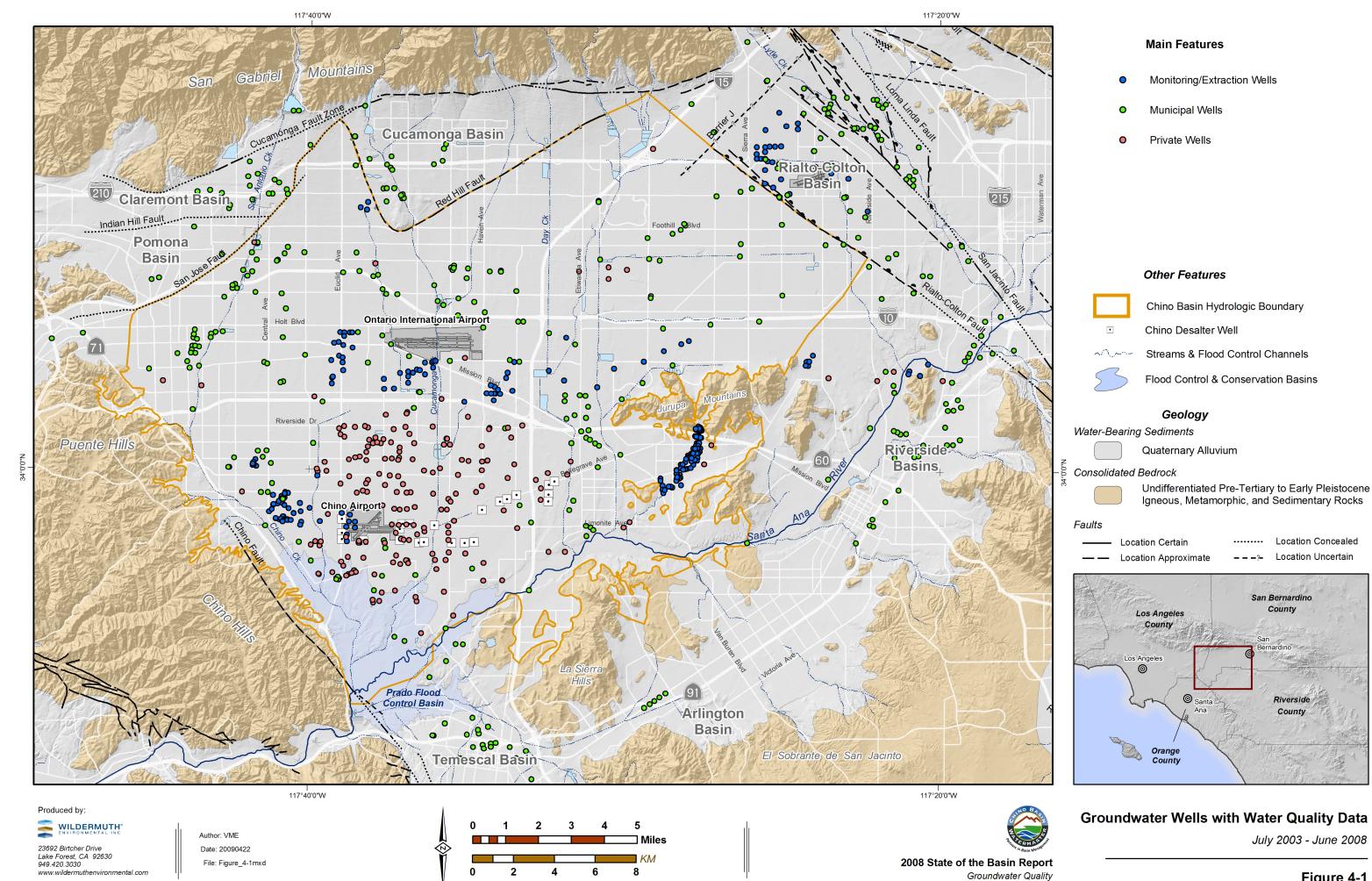
4.4 Conclusions and Recommendations

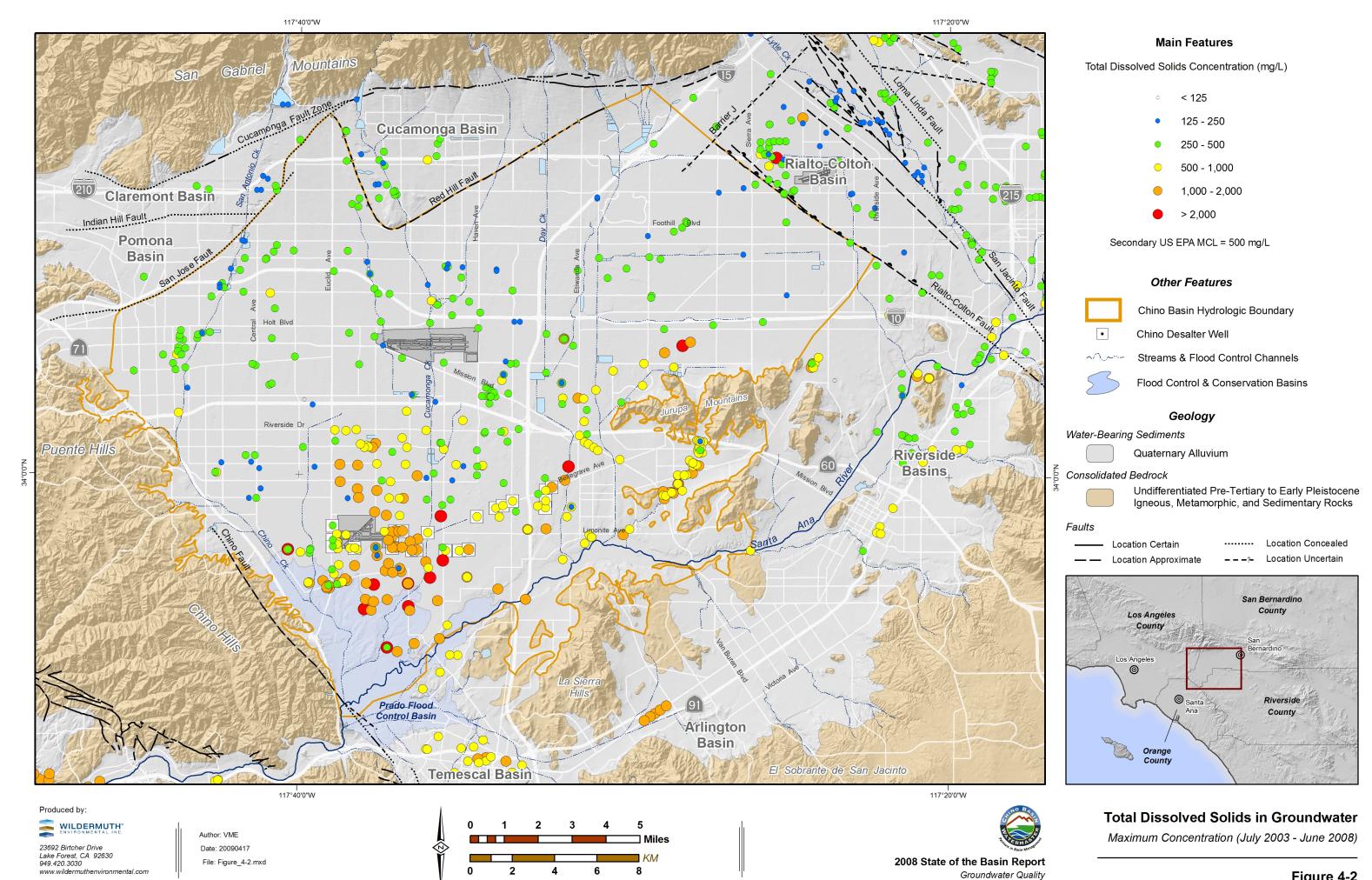
The Initial State of the Basin, and the 2004 and 2006 State of the Basin Reports discussed the need for future, long-term monitoring. Due to commercial and residential development in the Chino Basin area; many of the private agricultural wells that have been used for monitoring activities are destroyed as land is developed.

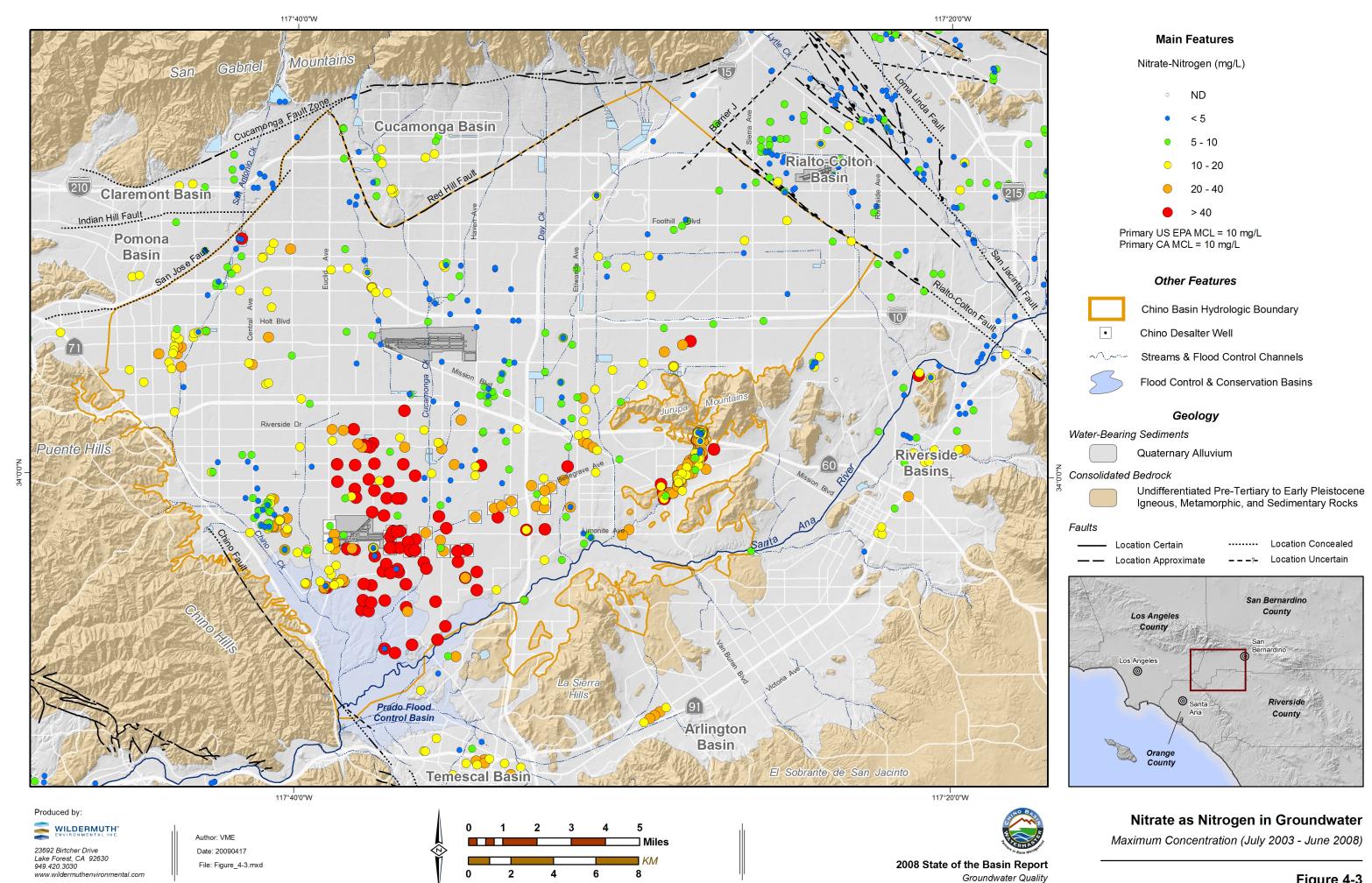
In response to the loss of historically utilized wells, Watermaster developed a water quality key well program. This program designates a series of wells across a wide areal distribution for long-term monitoring activities. To establish the well network, a grid was overlain the basin, and, where possible, at least one well was chosen per grid cell. Wells that are part of the water level monitoring program and/or on property that is not likely to be developed were preferentially chosen. Details of the Key Well Groundwater Quality Monitoring Program are available in the 2008 Chino Basin Maximum Benefit Annual Report and in Section 4.2.2 of this report. Key well sampling began in fall 2005 and runs in two-year cycles. Sampling results are added to the Watermaster database.

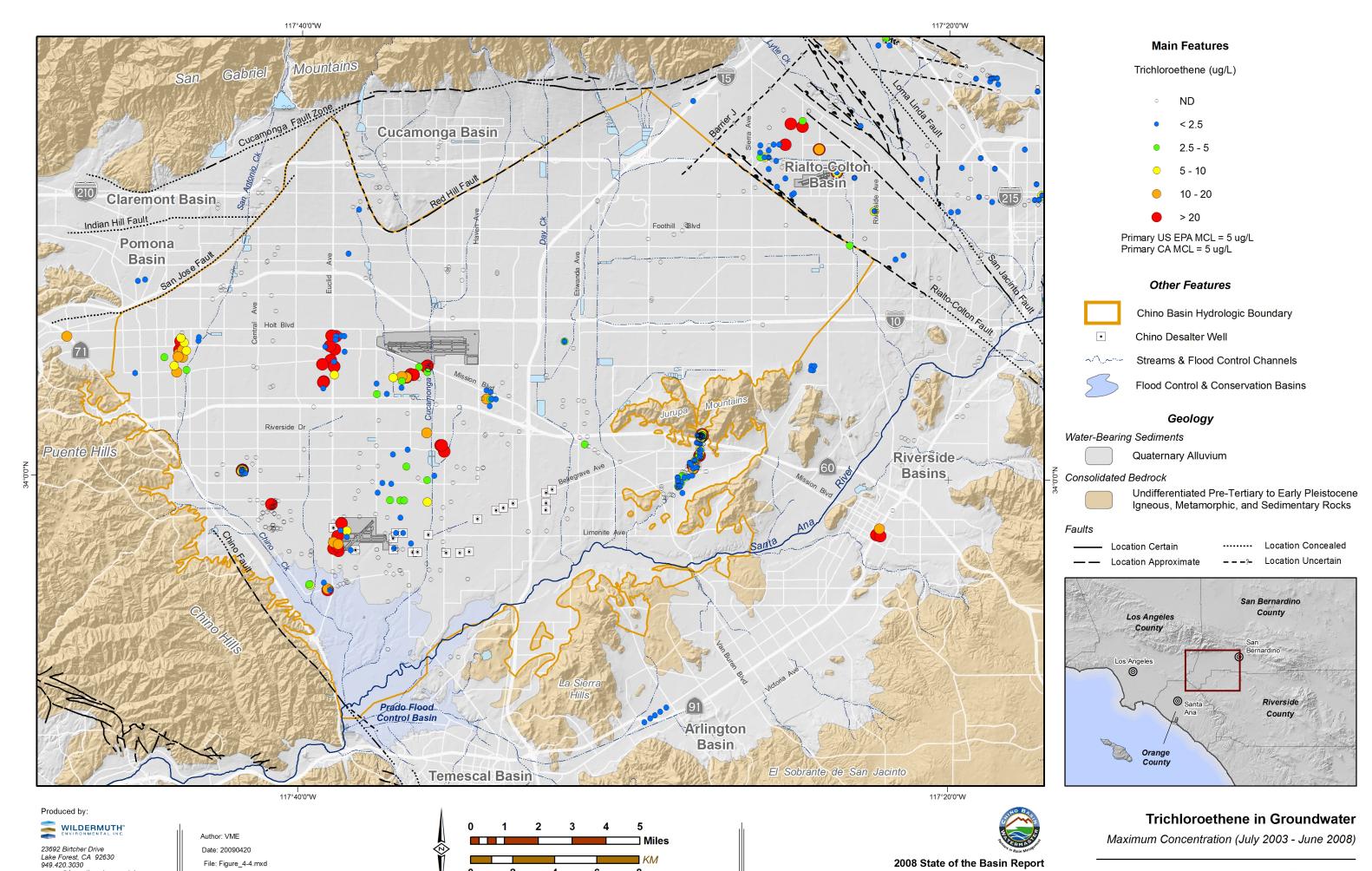
Point sources of concern are critical to the overall quality of Chino Basin groundwater. To ensure that Chino Basin groundwater remains a sustainable resource, it is of the utmost importance that Watermaster closely monitor point sources and emerging contaminates. It is recommended that Watermaster continue to work closely with the RWQCB and potentially responsible parties within the Chino Basin. This will allow for up-to-date understanding of groundwater quality, investigations, remediation activities, and potential mutually beneficial remedial options through Chino Basin desalting facilities.



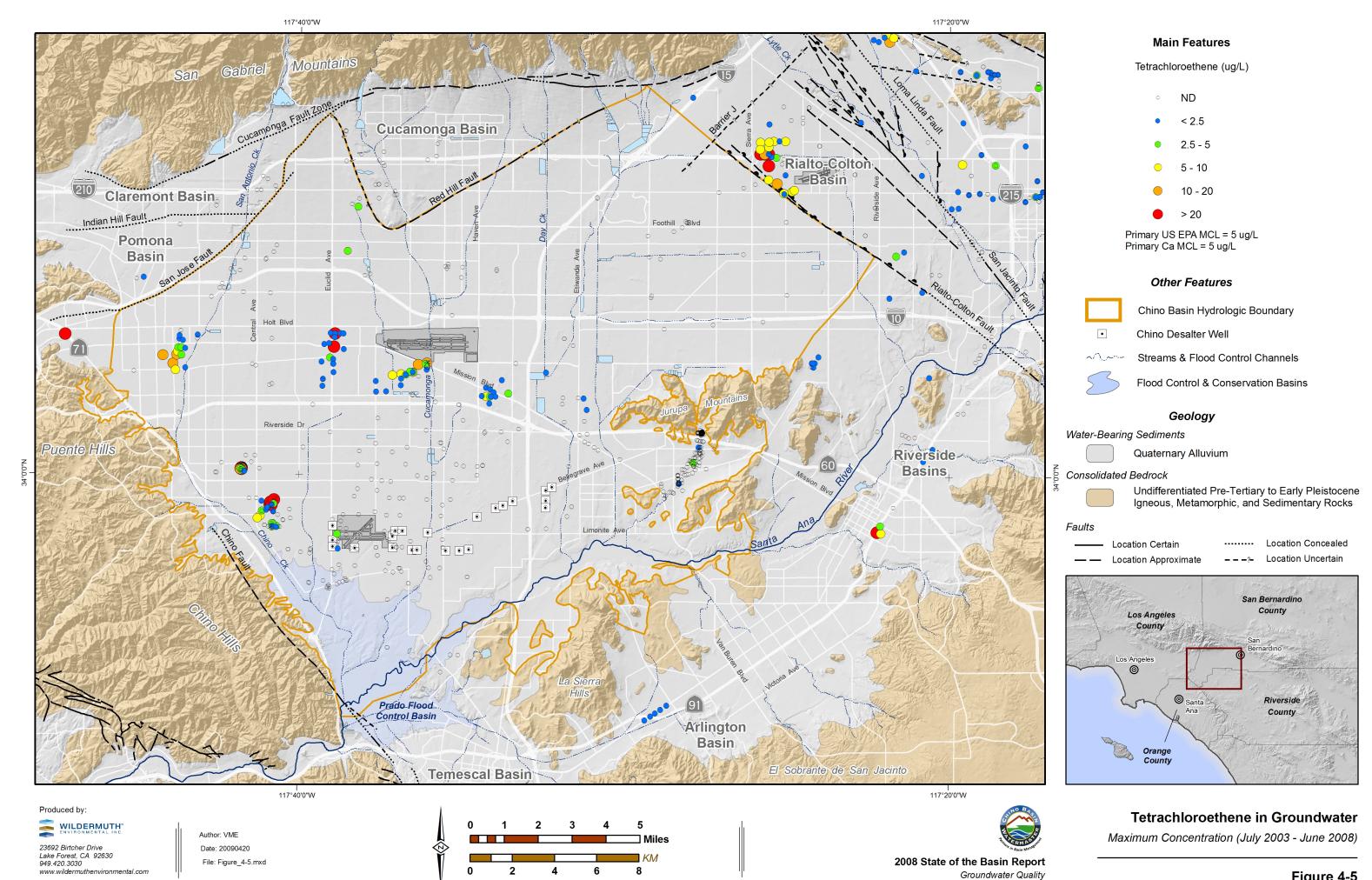


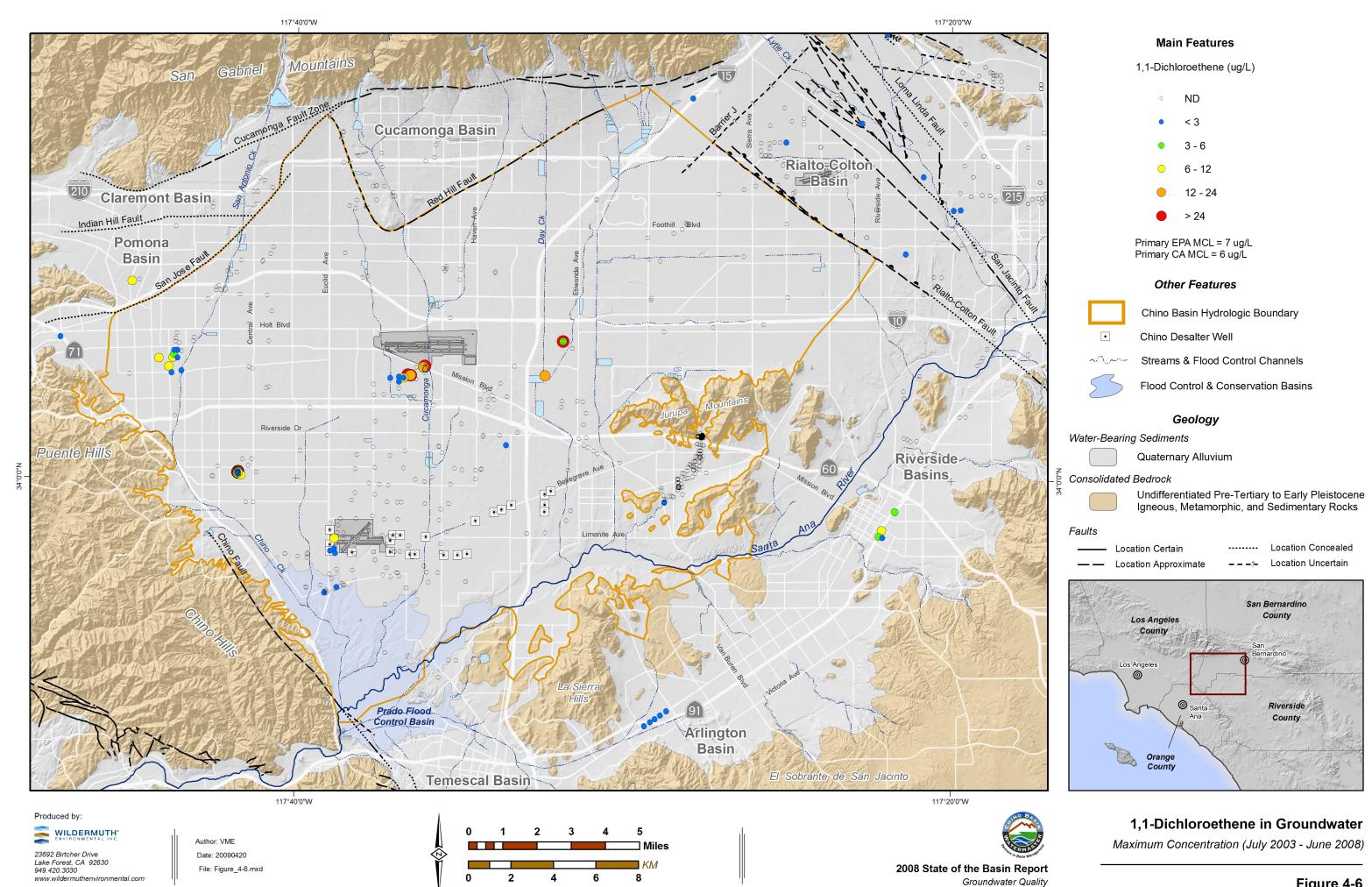


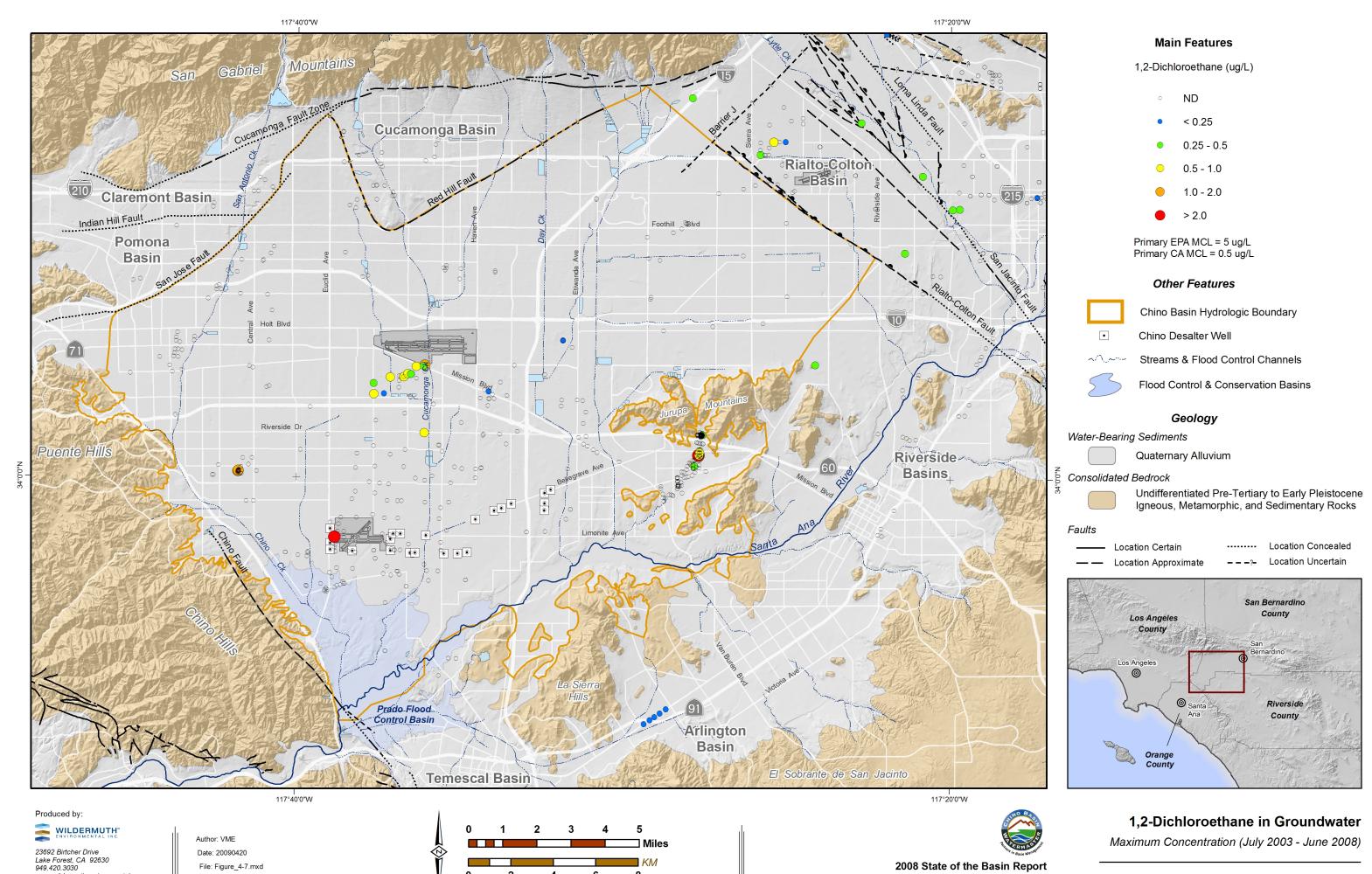




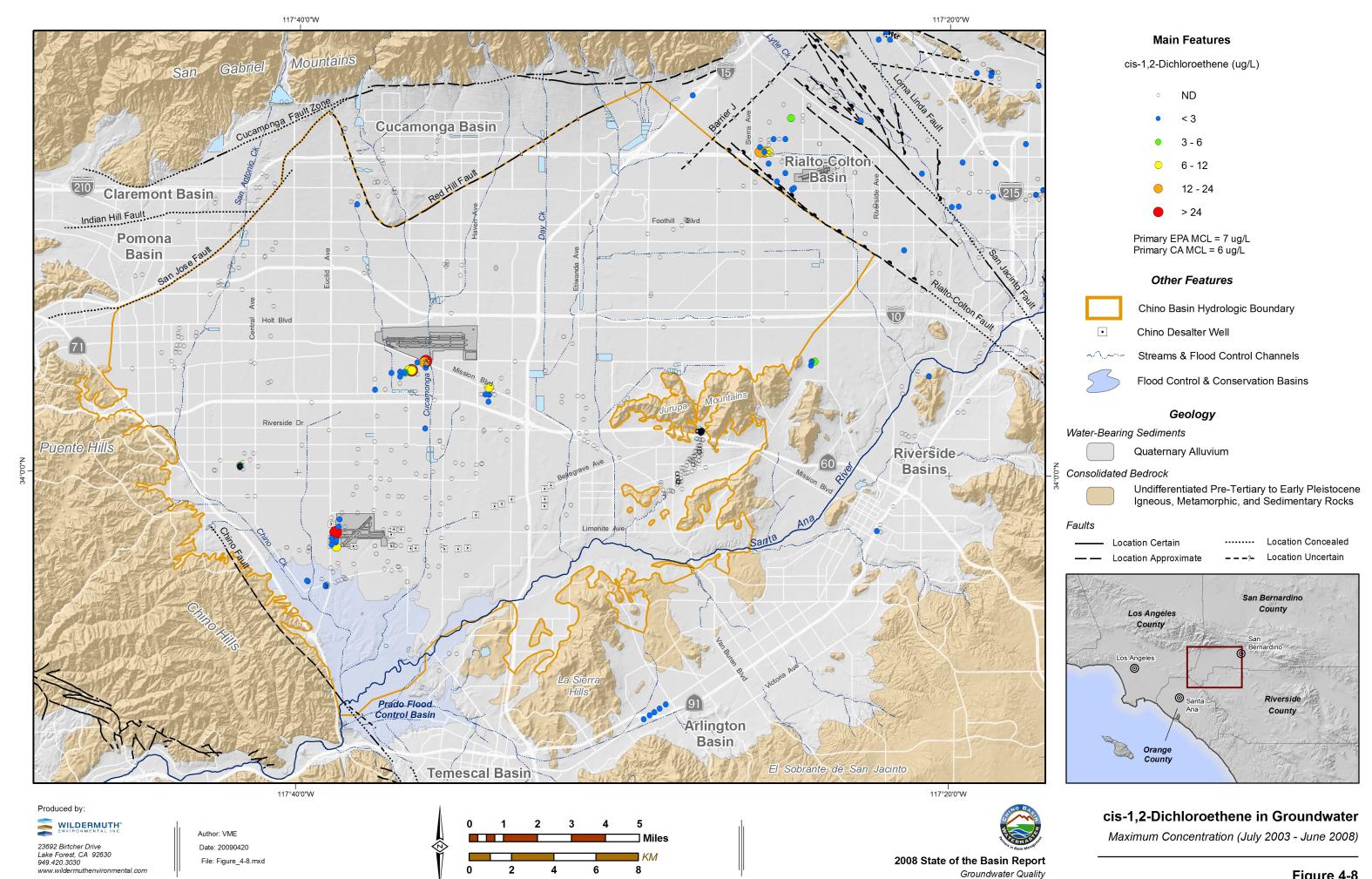
Groundwater Quality

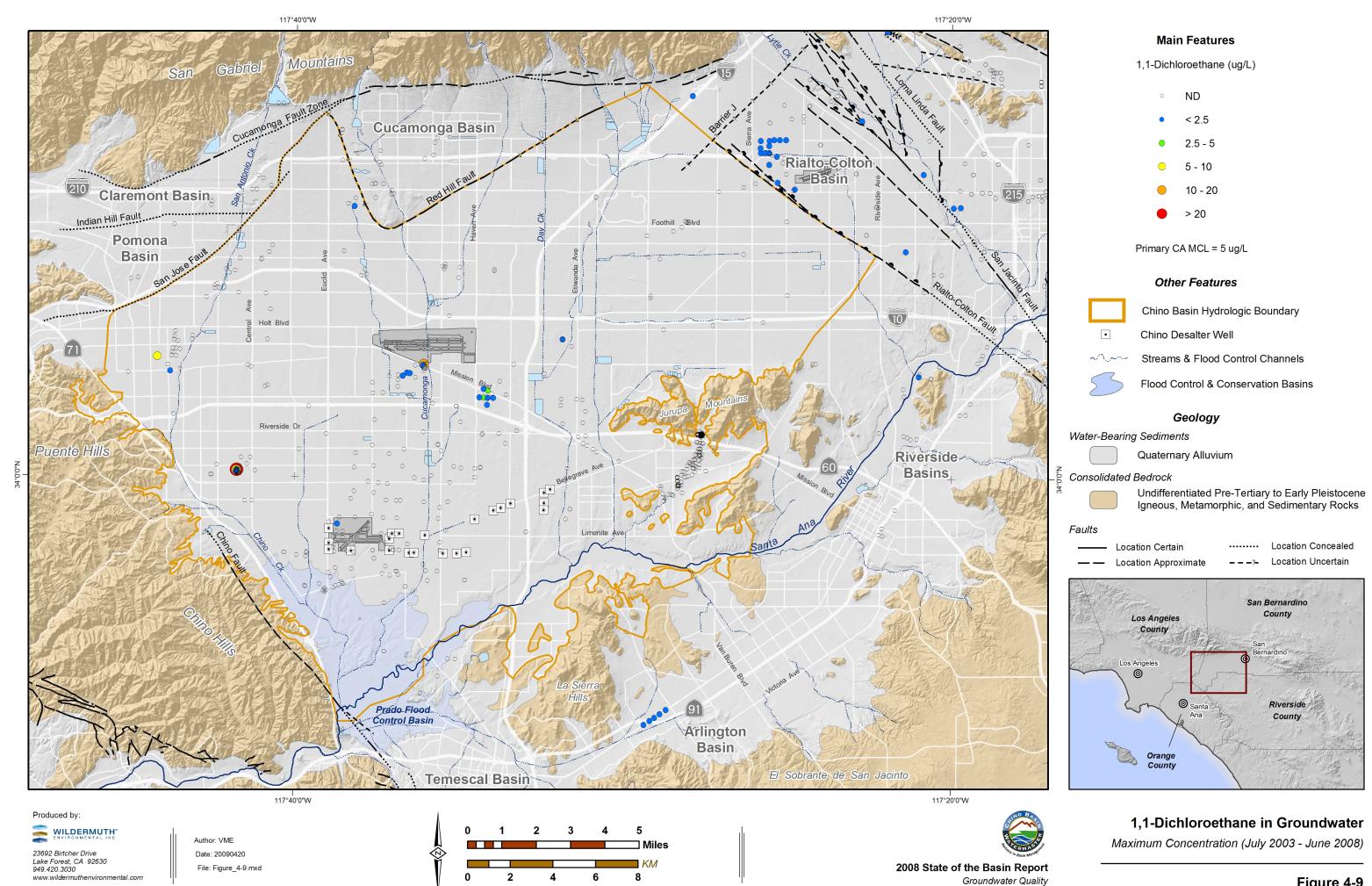


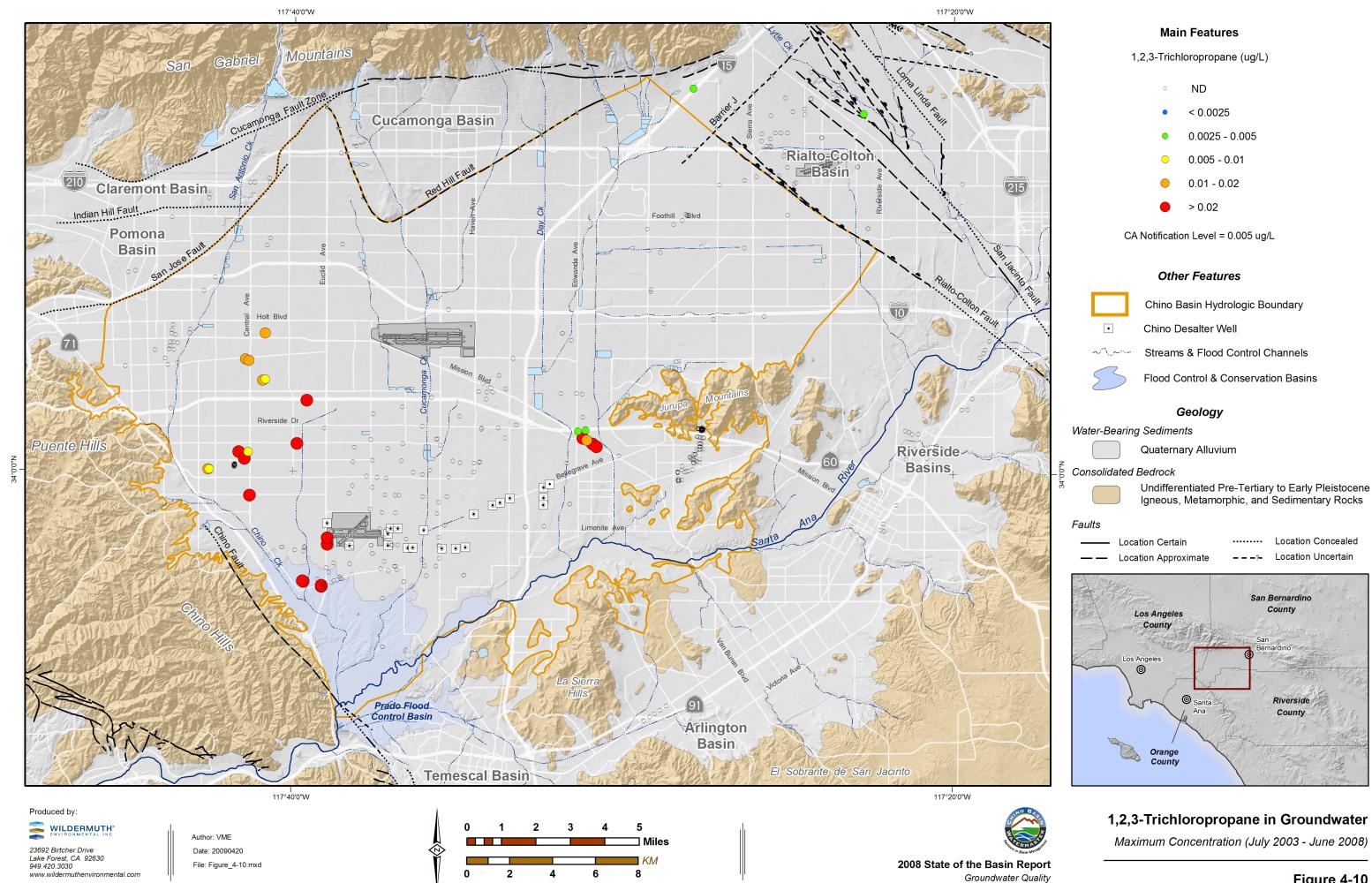


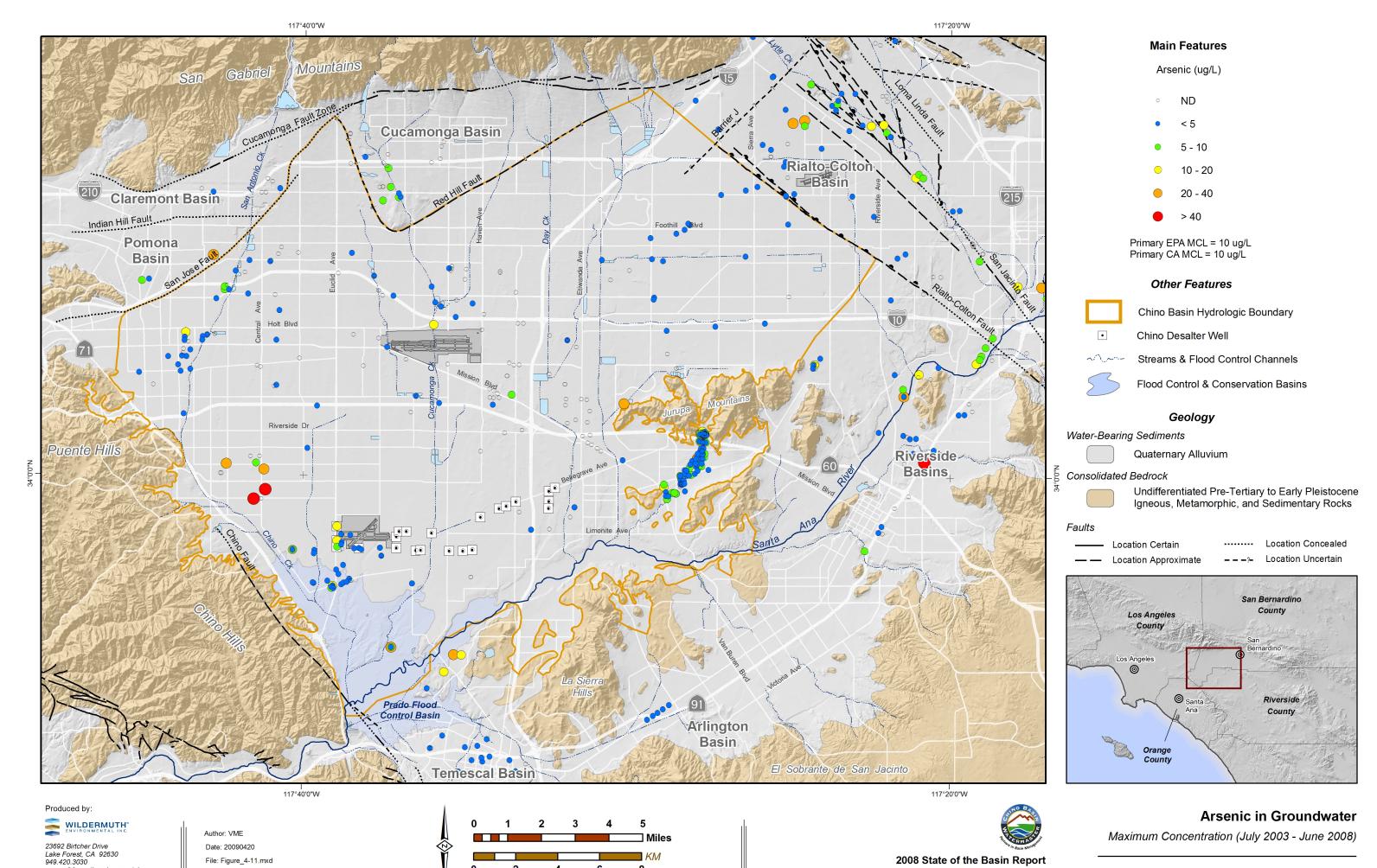


Groundwater Quality

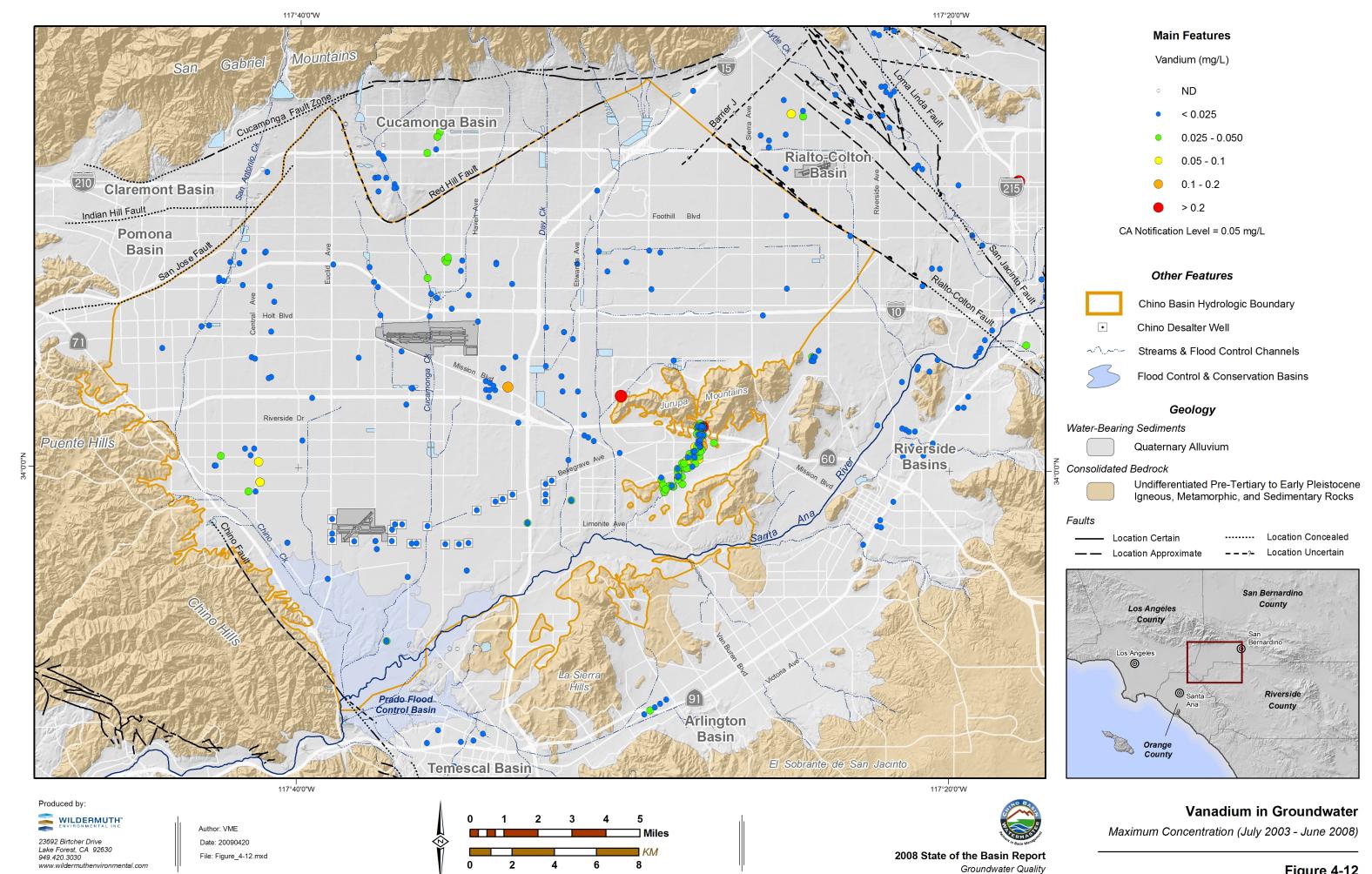


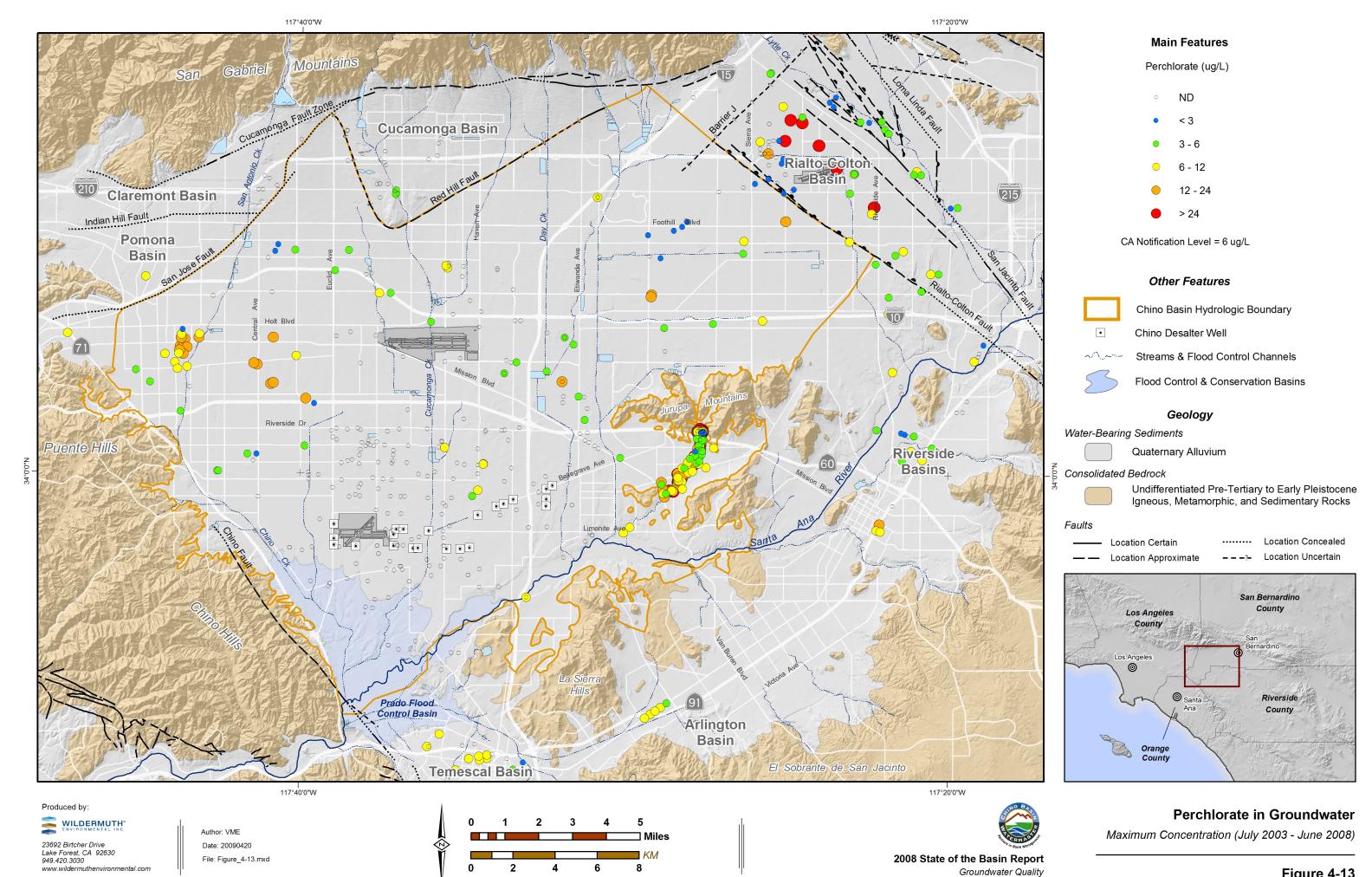


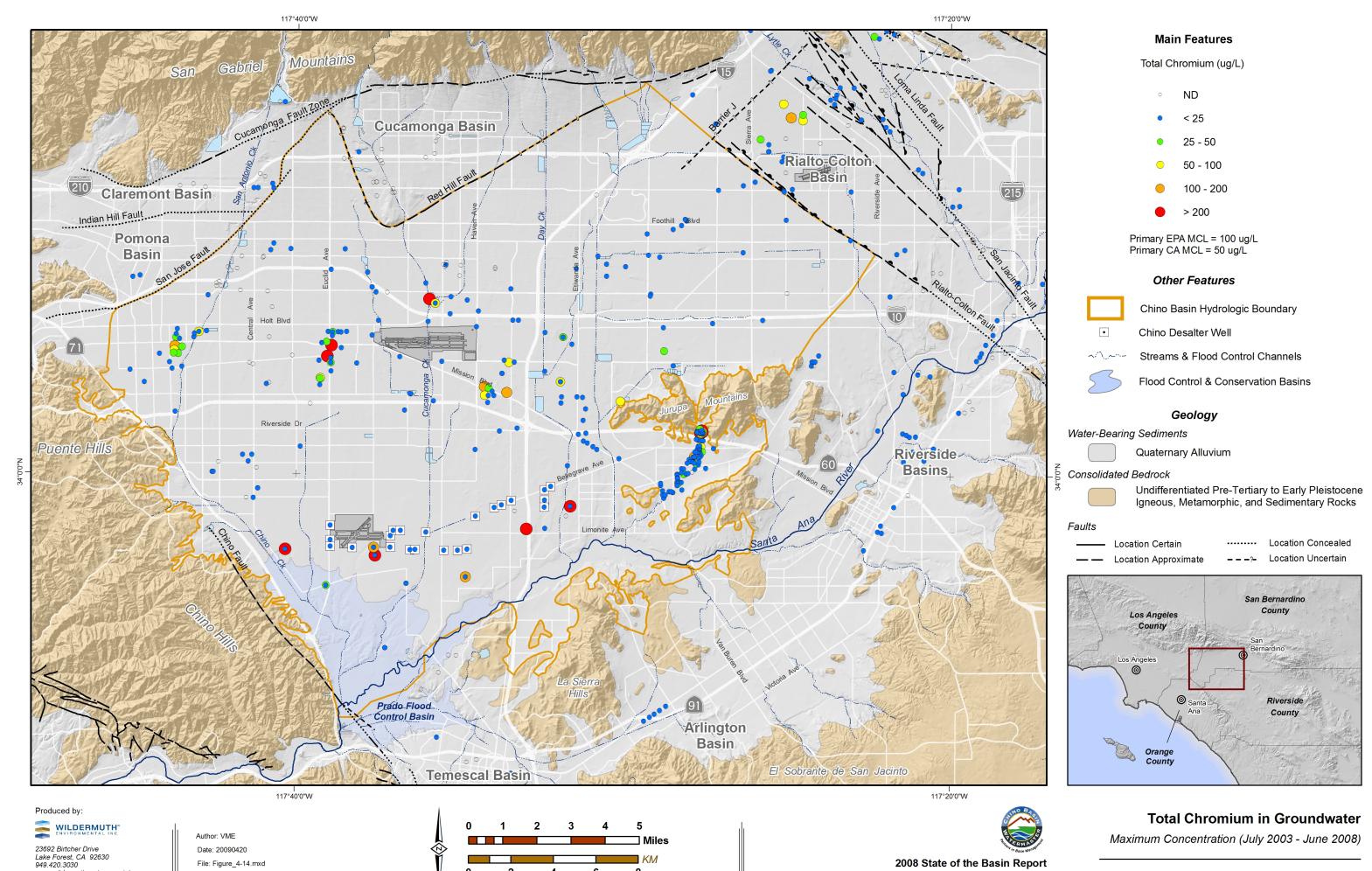


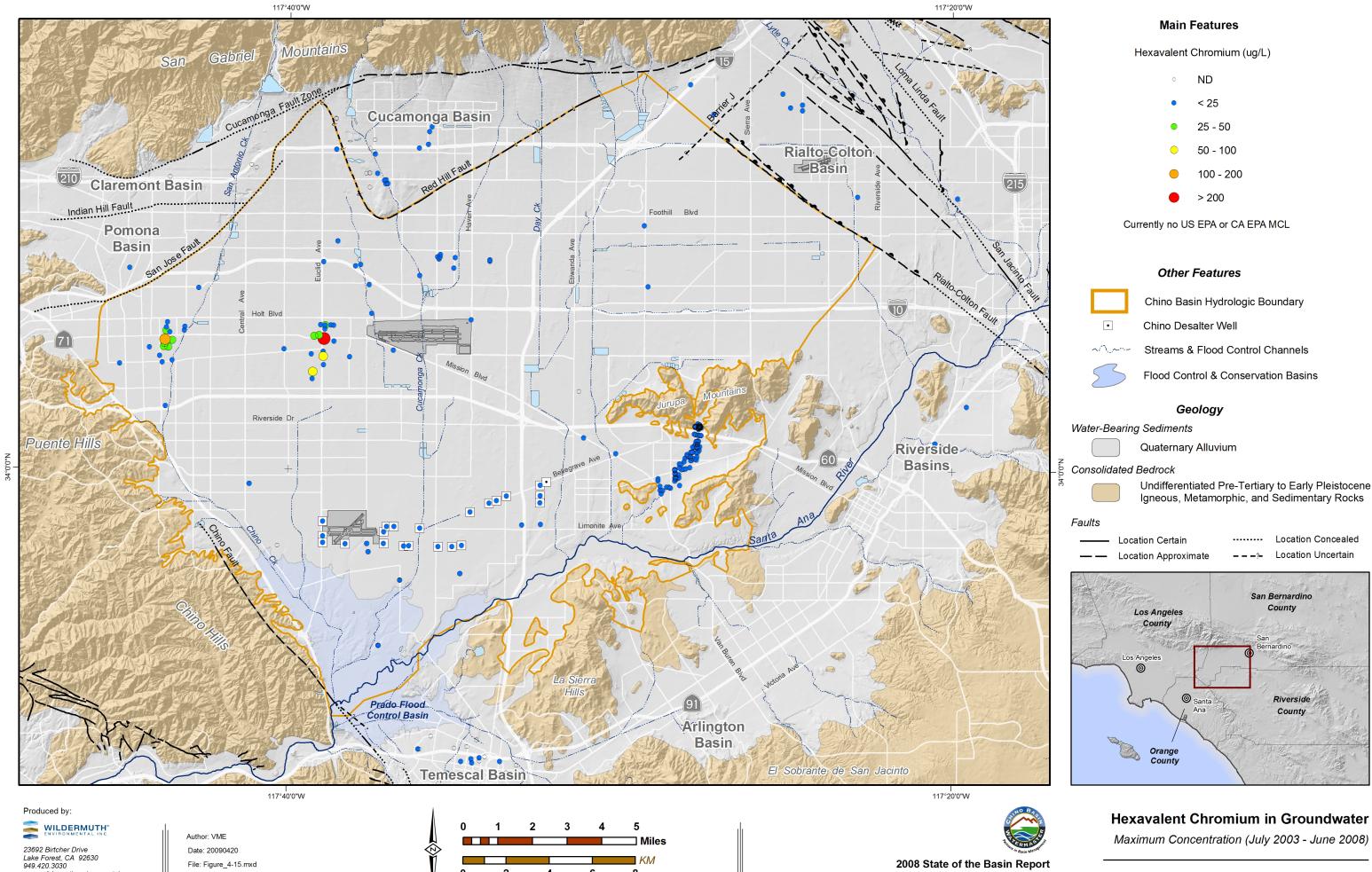


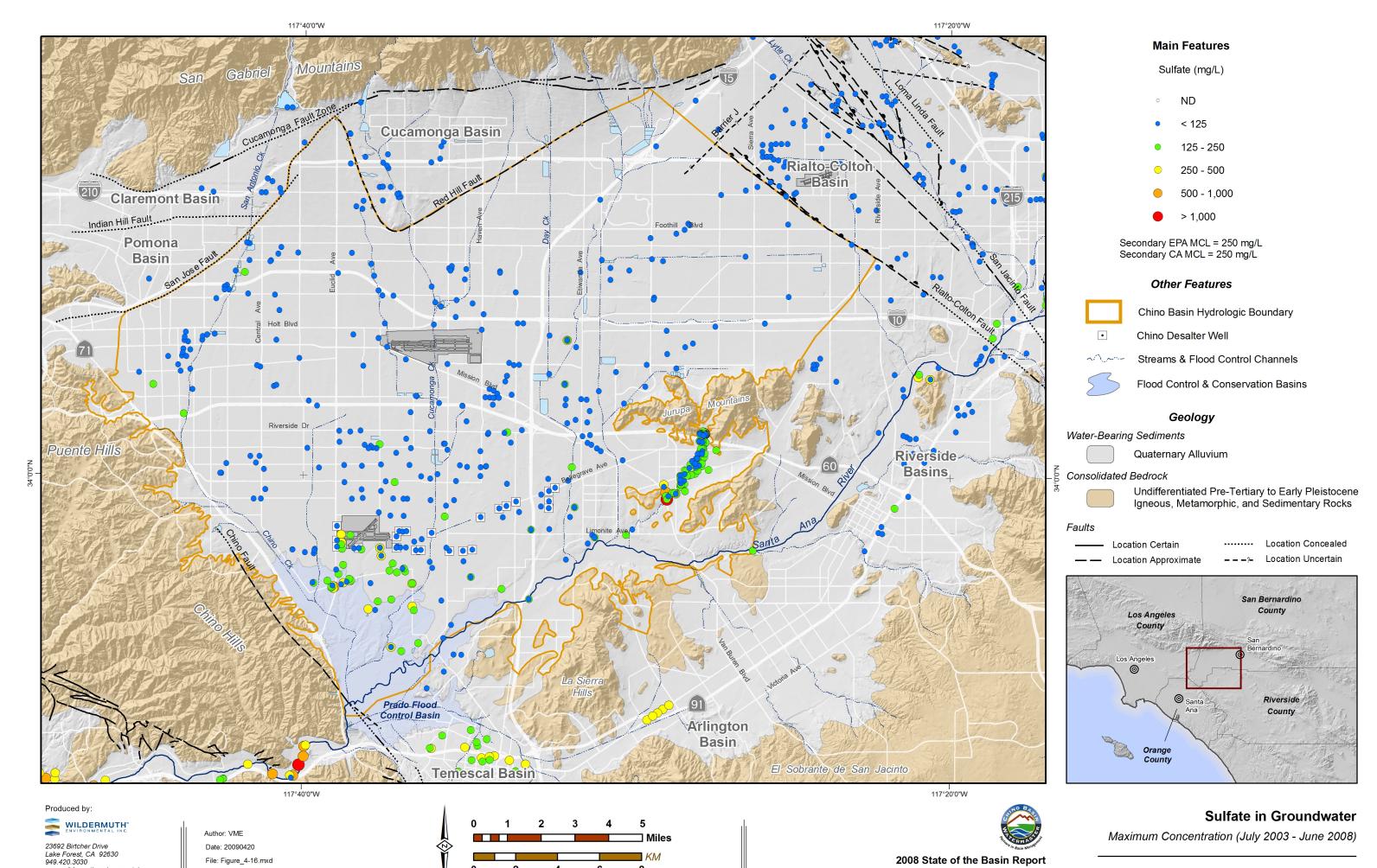
Groundwater Quality

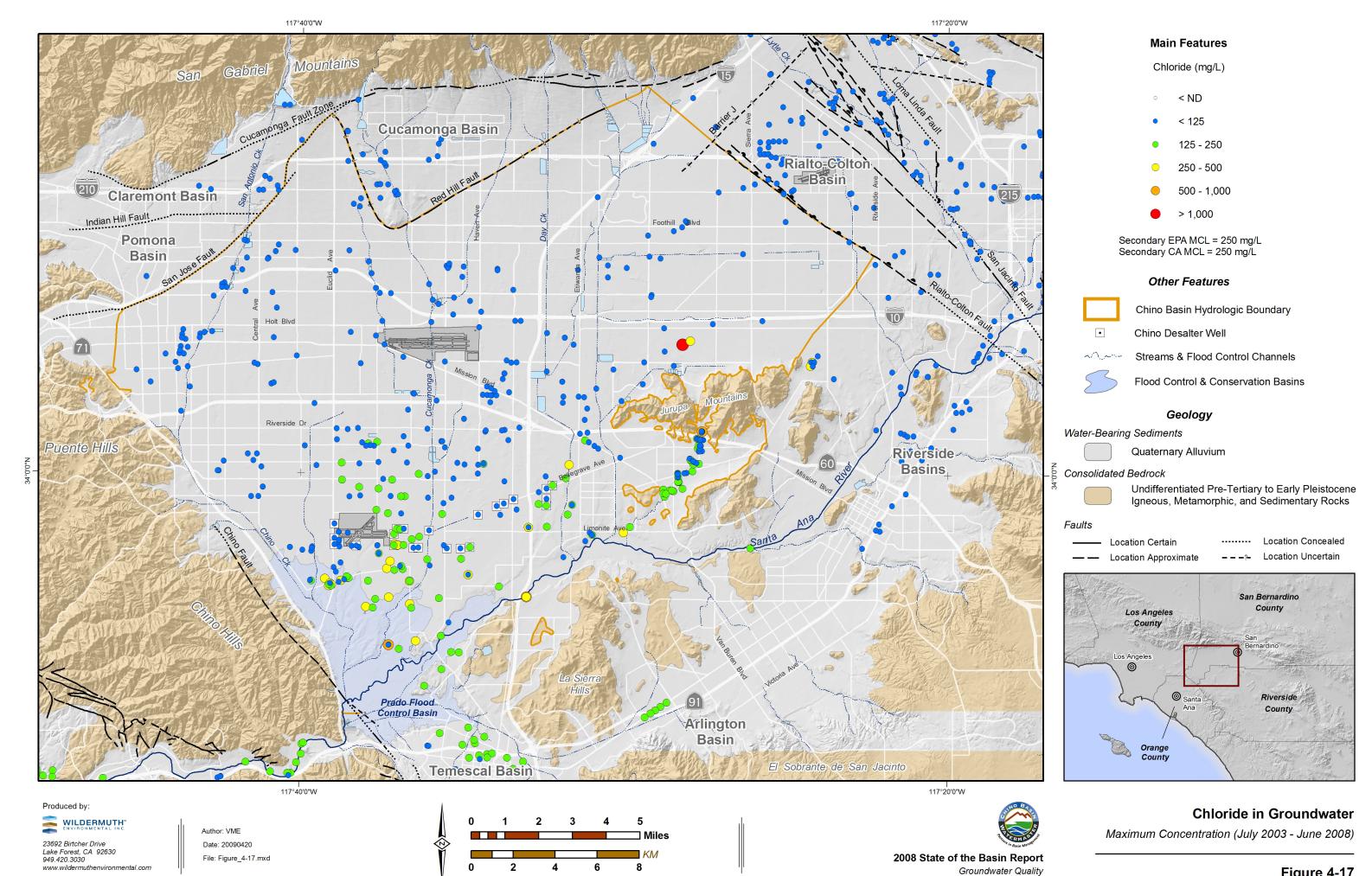


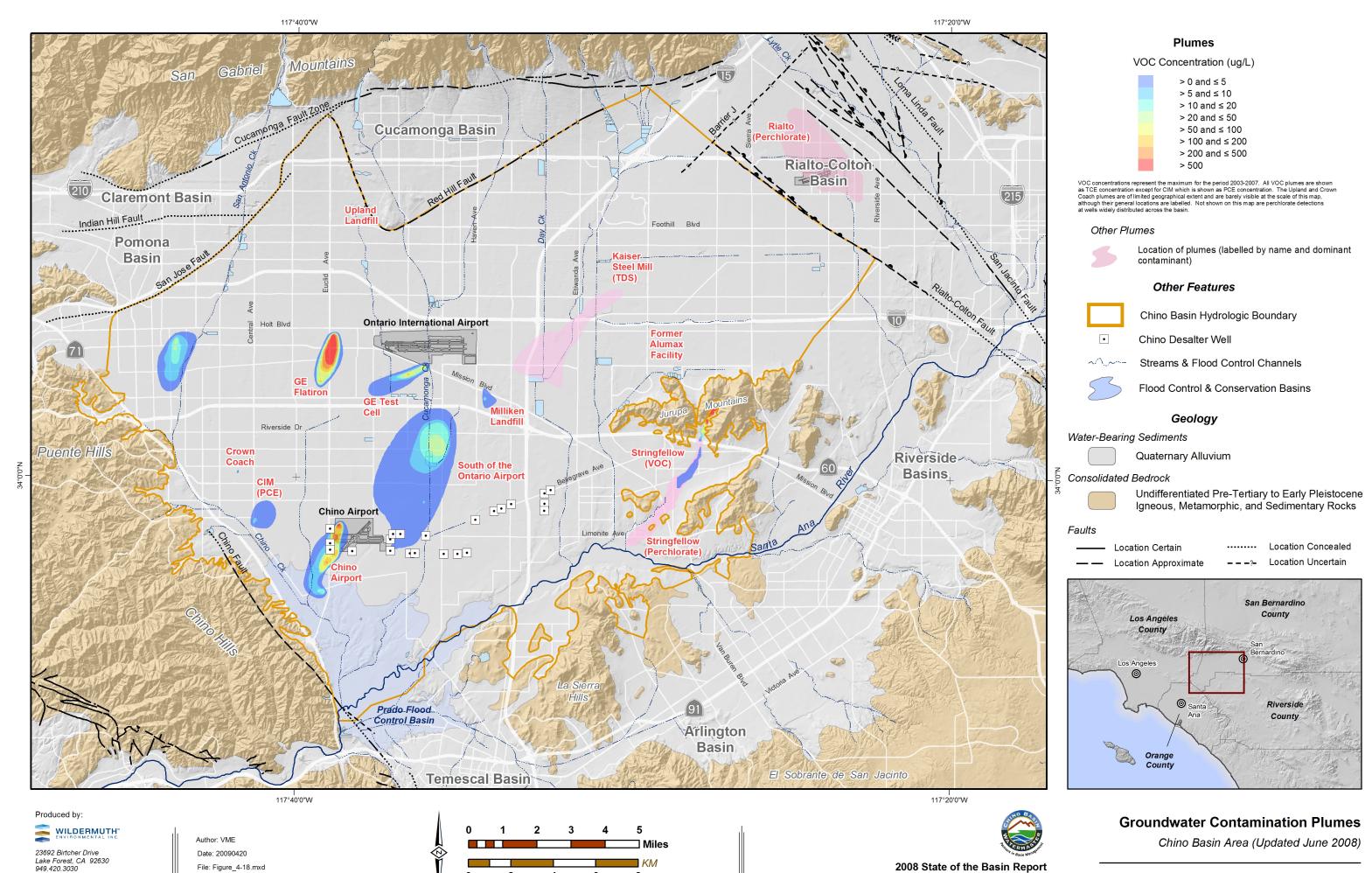


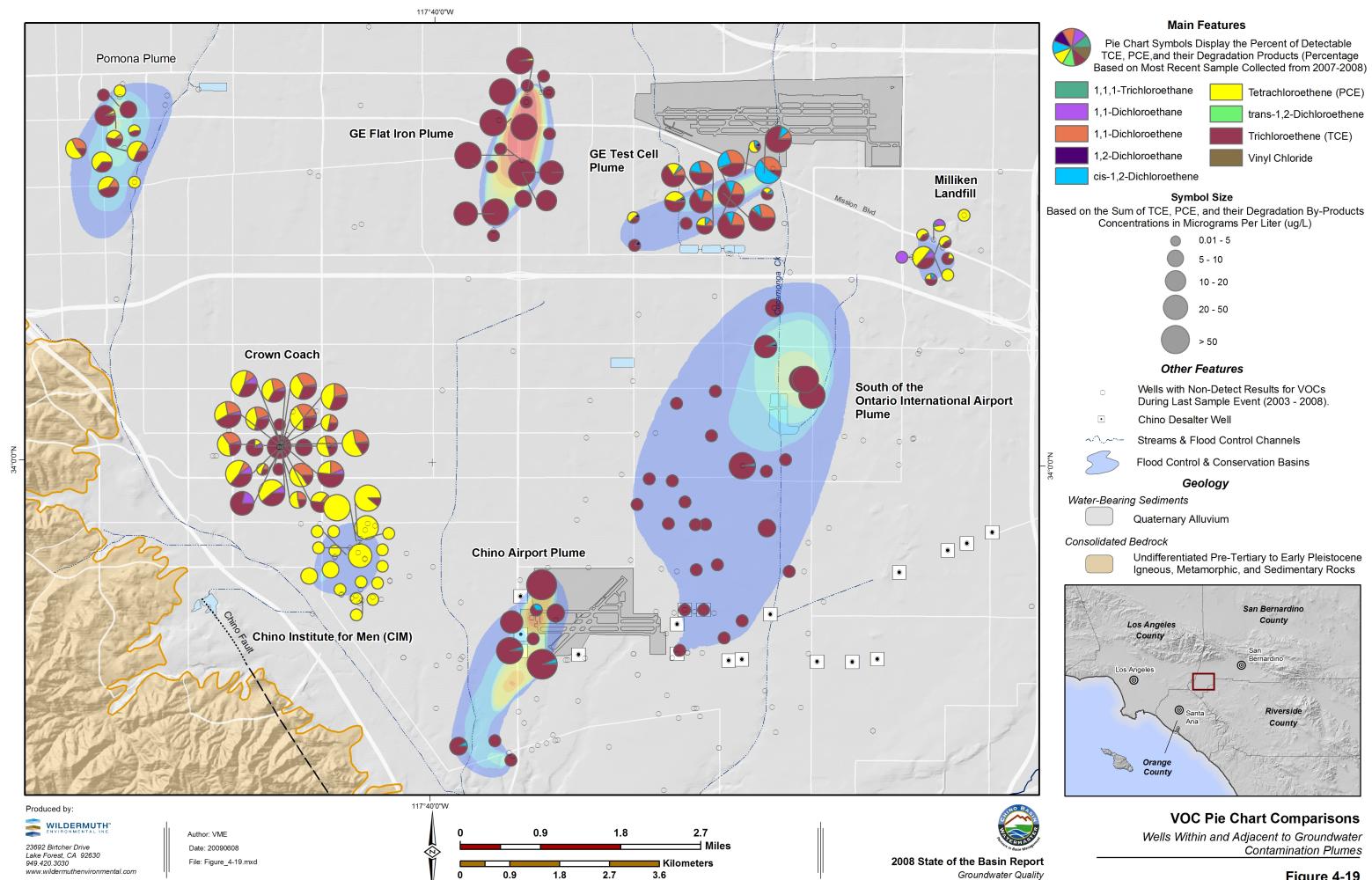










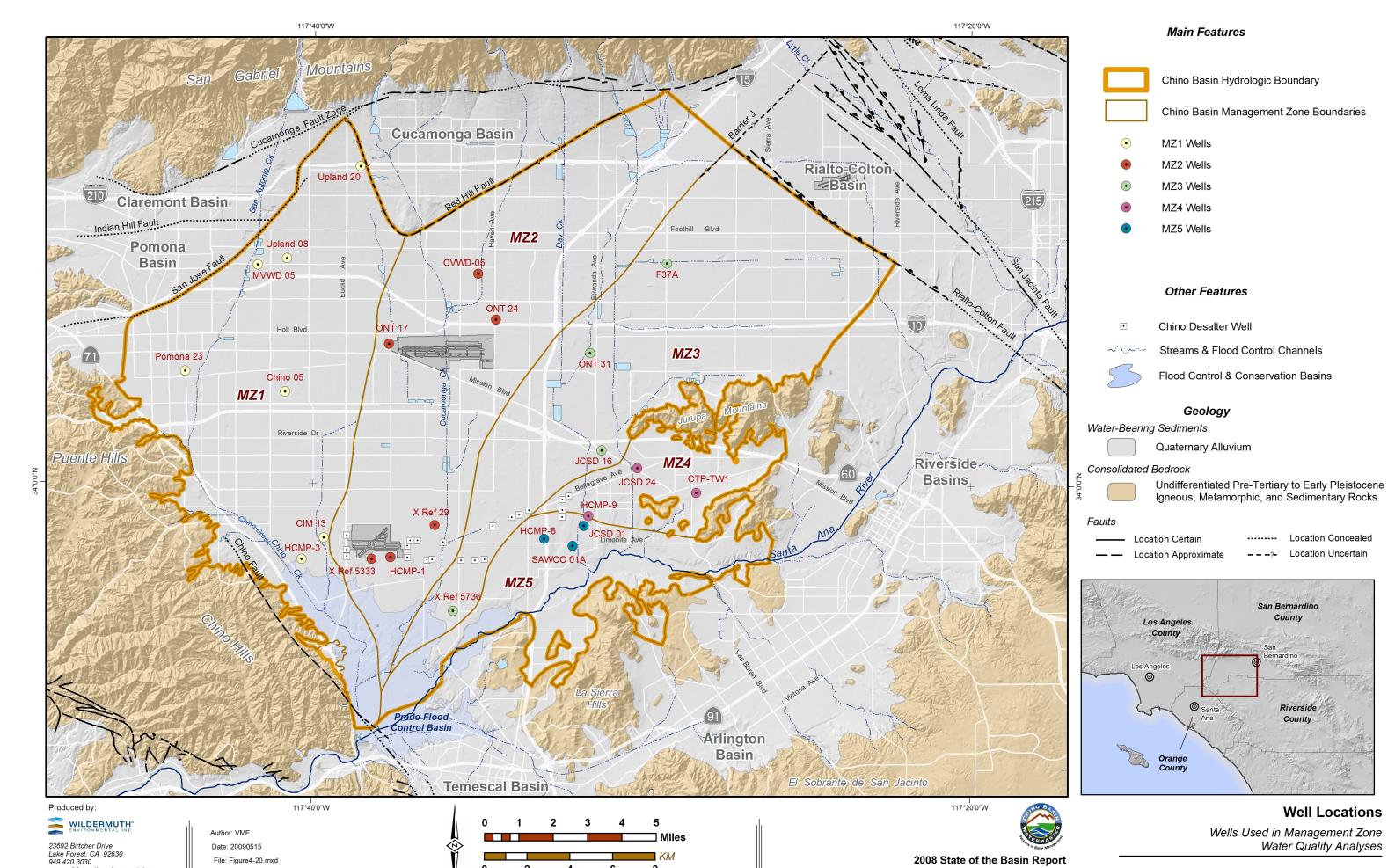


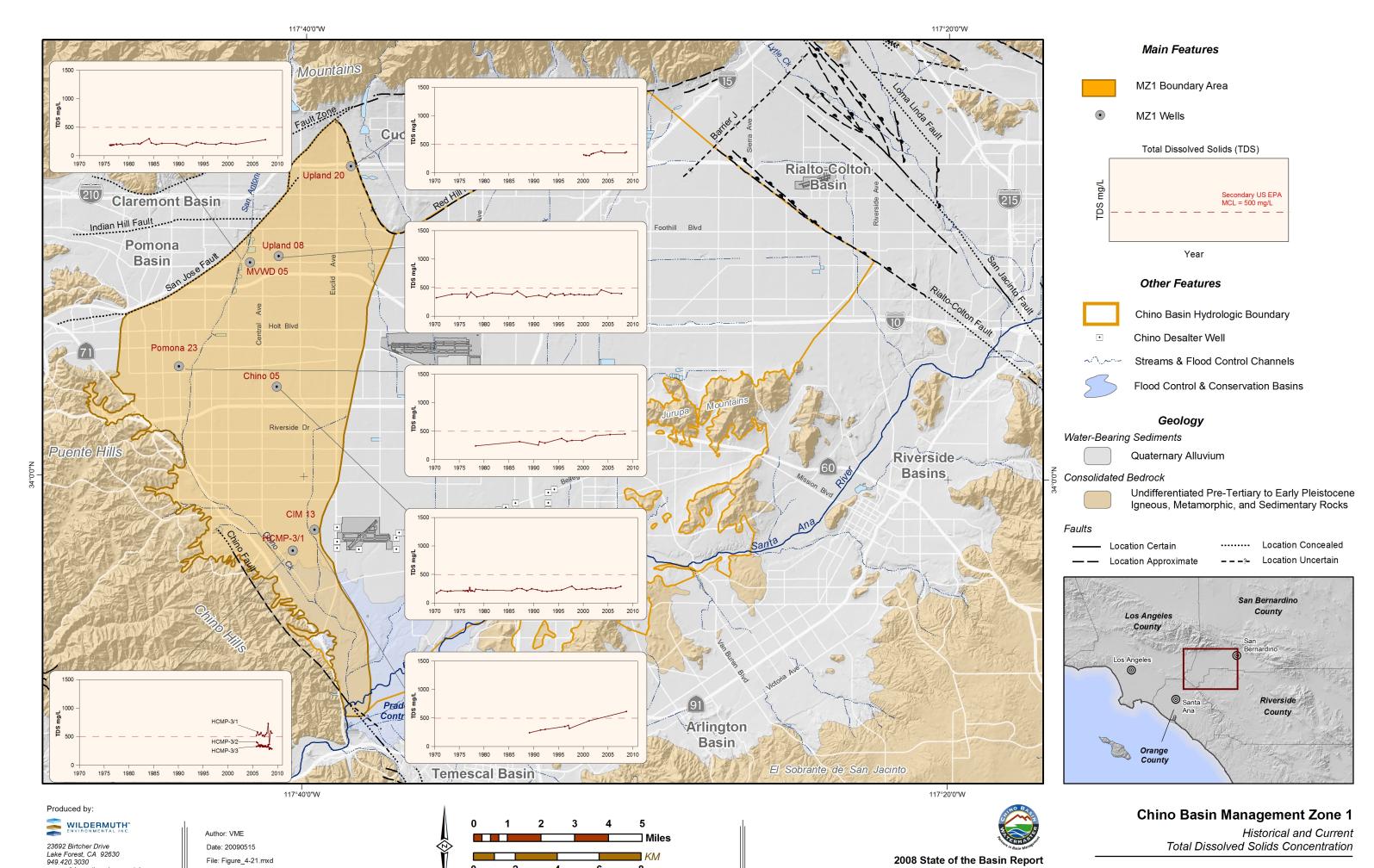
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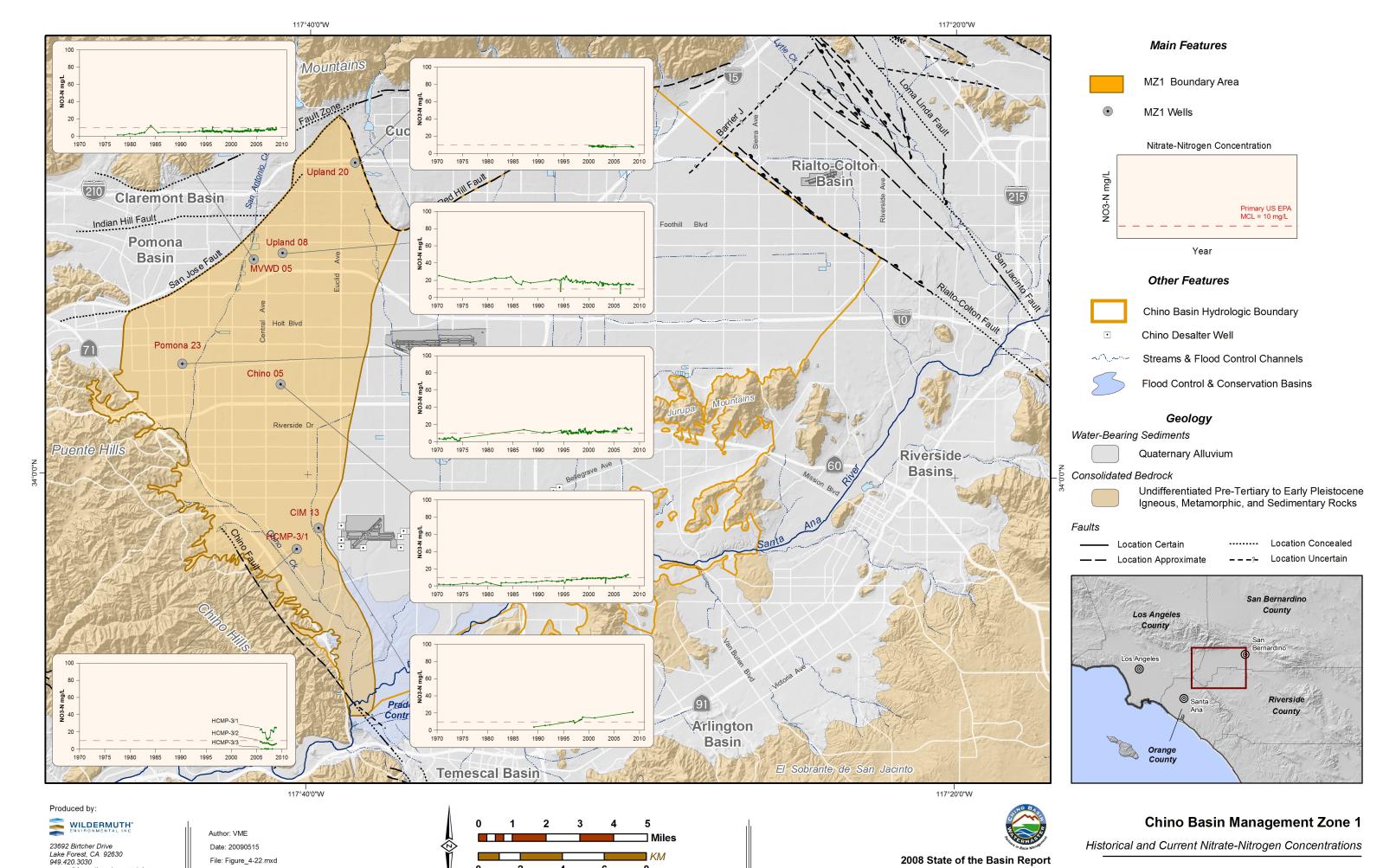
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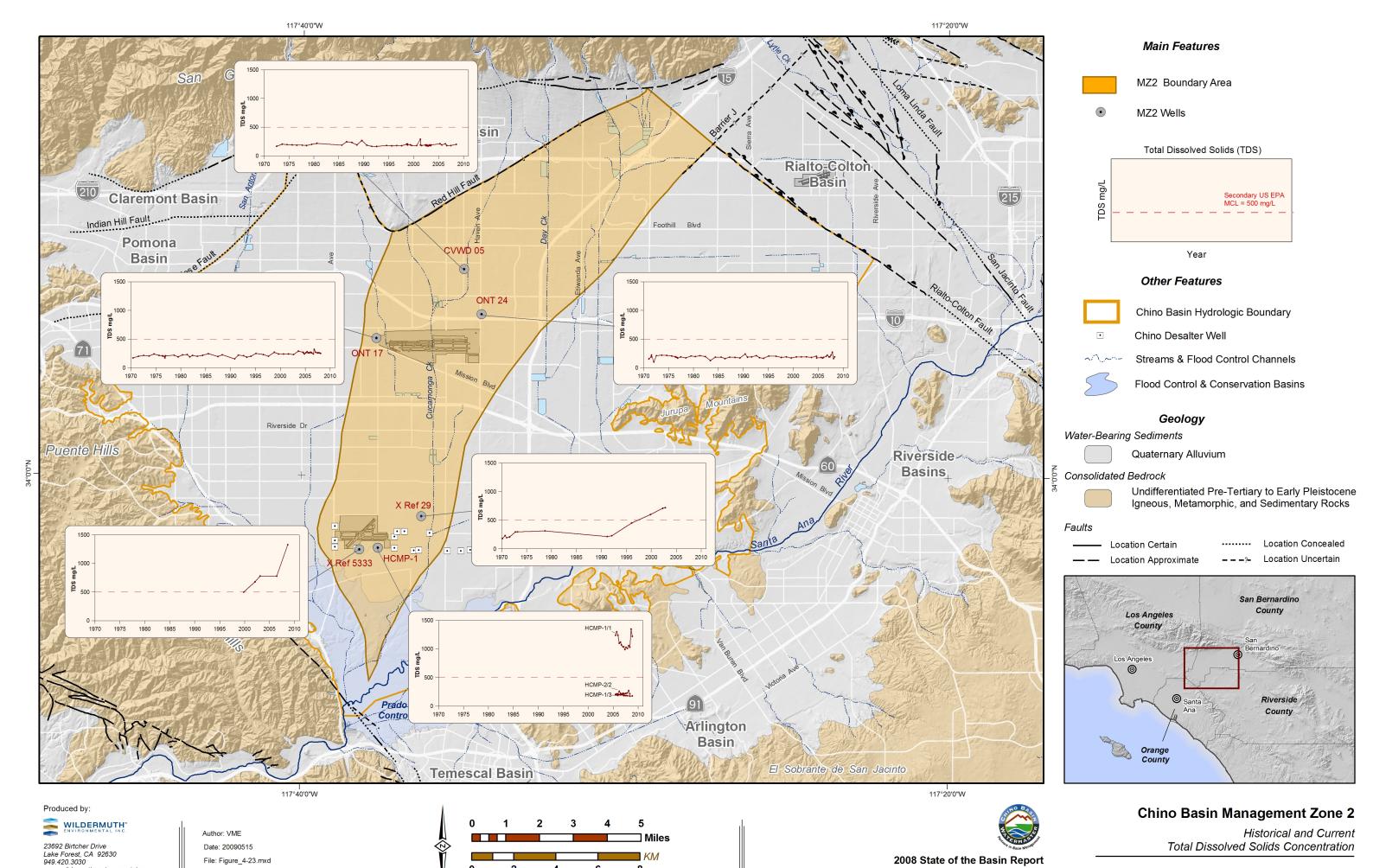
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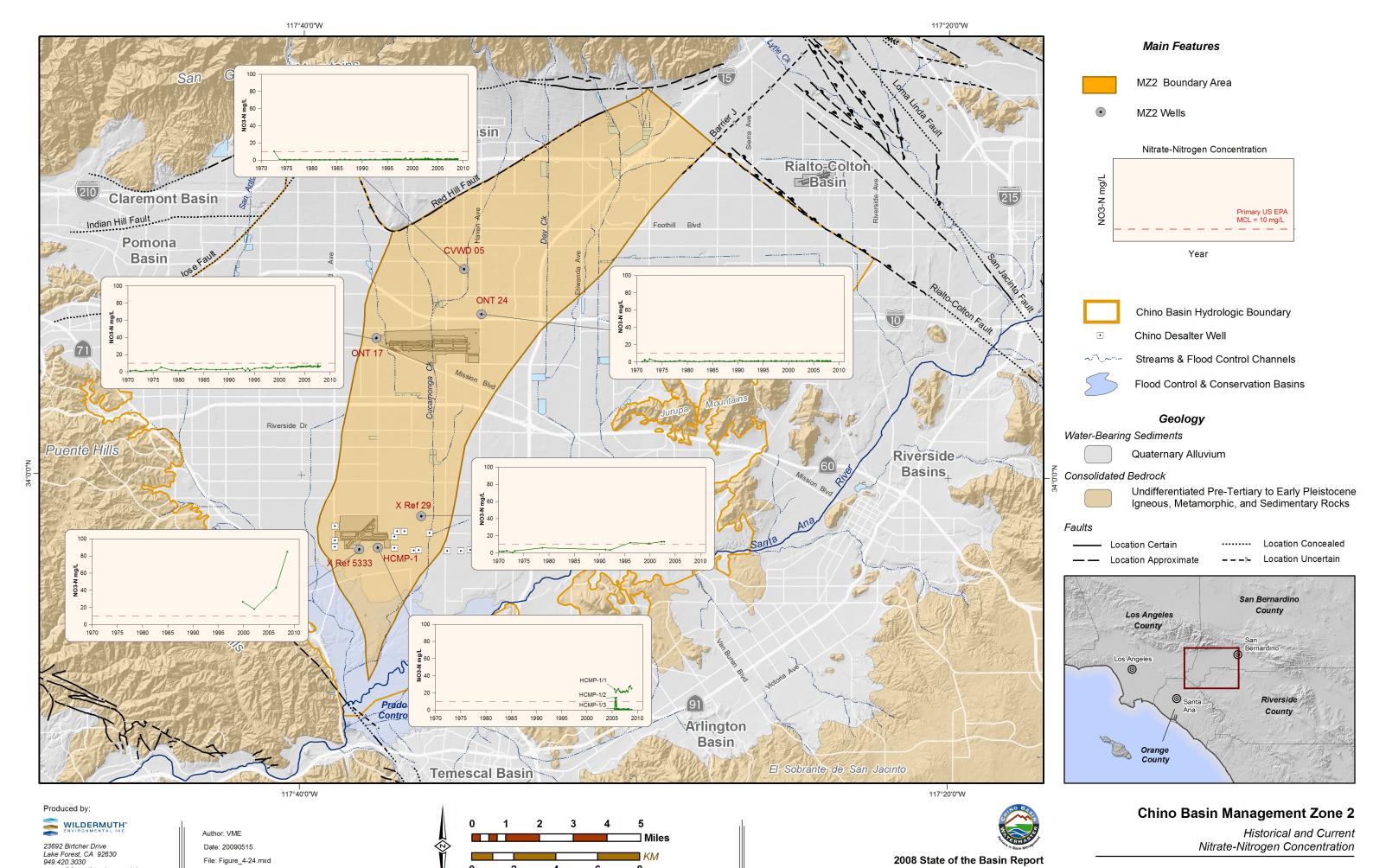
Figure 4-19

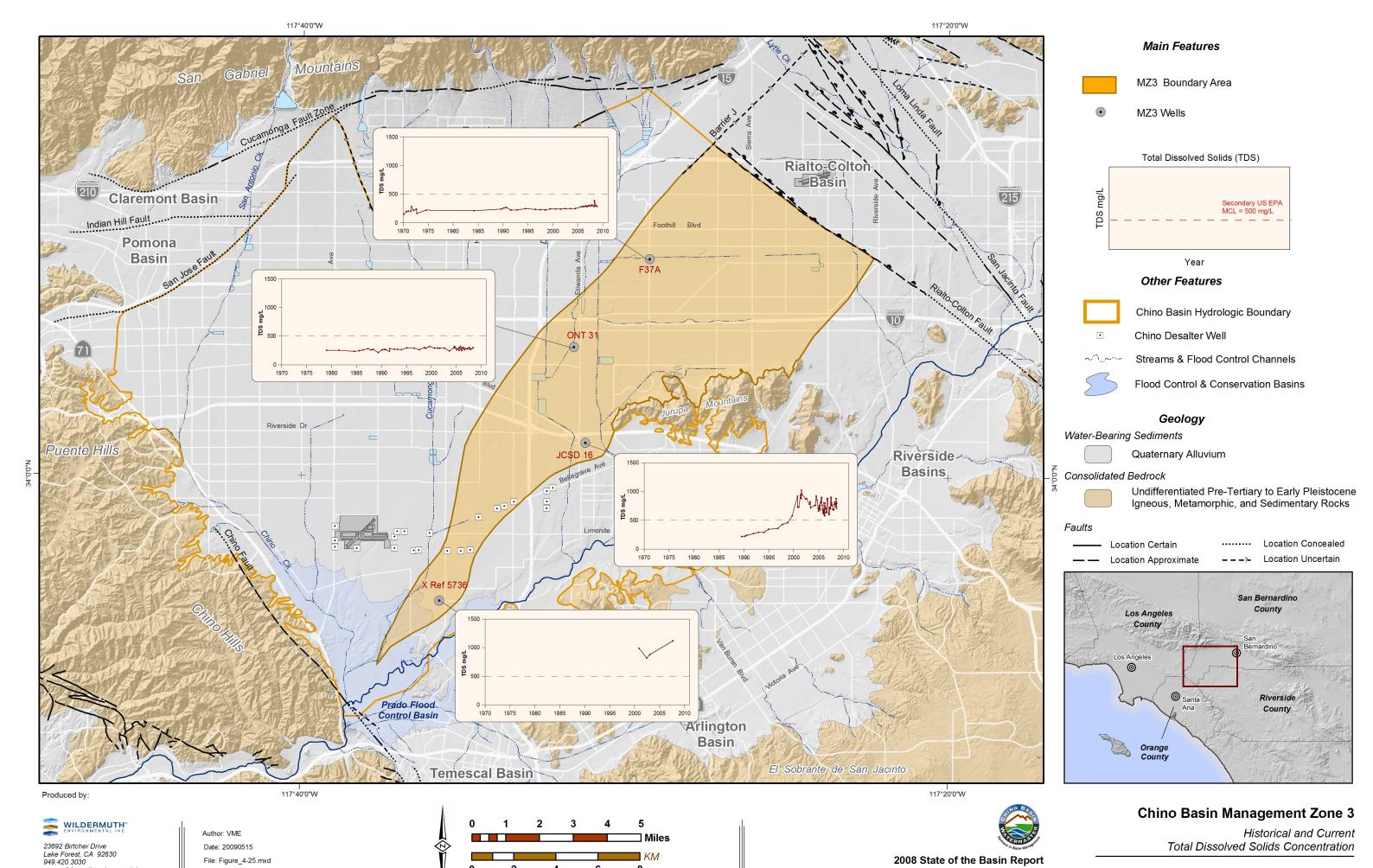


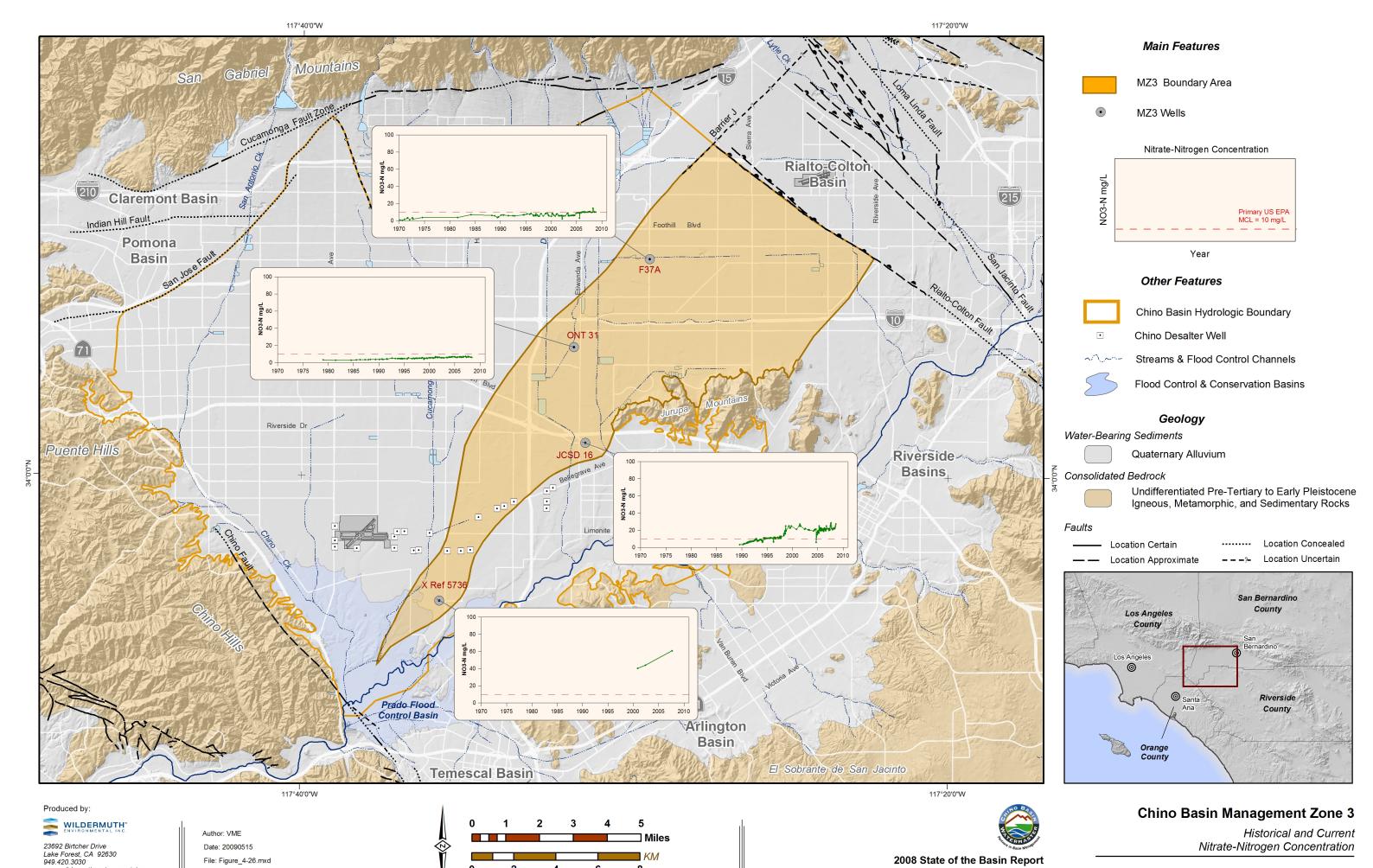


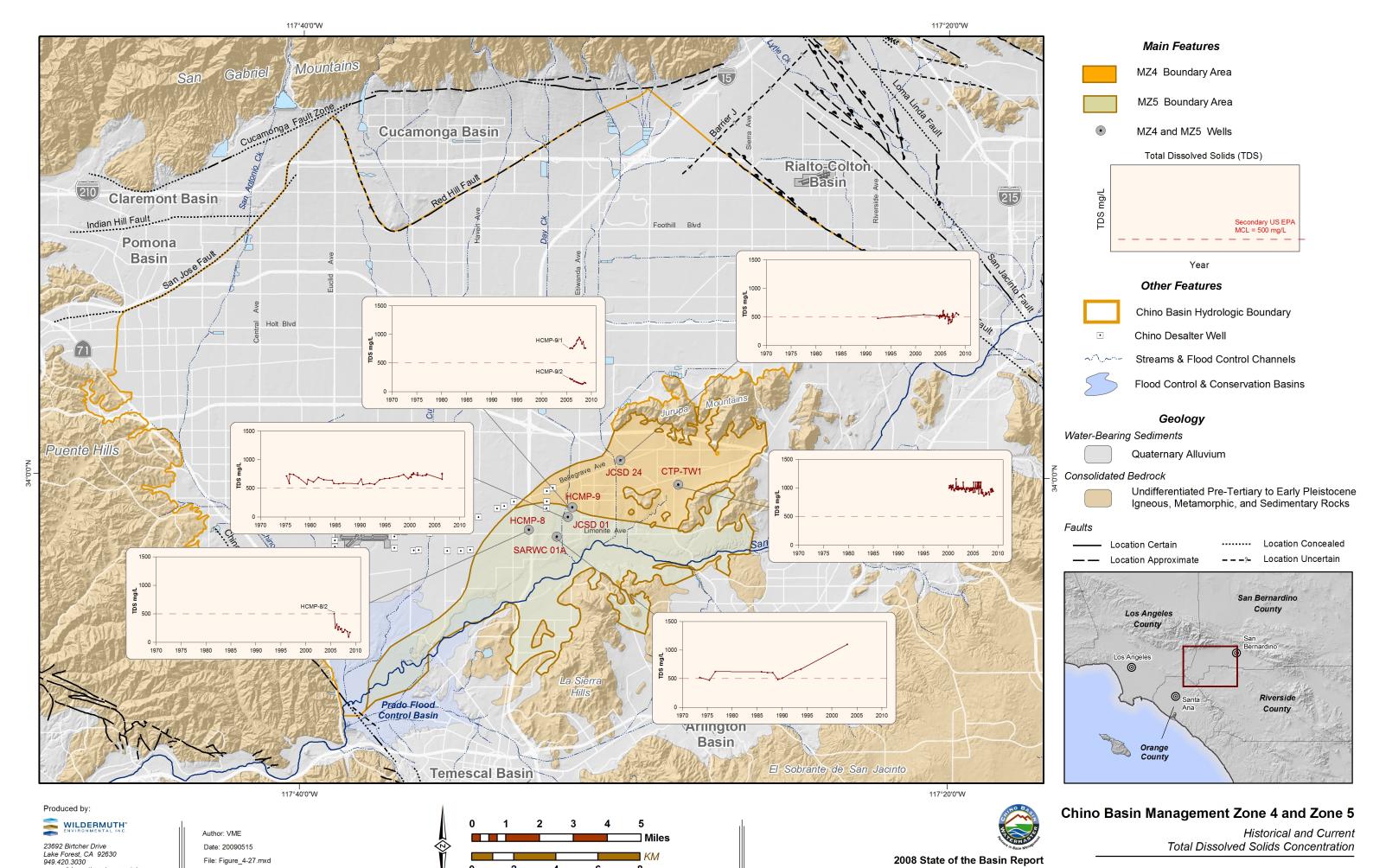


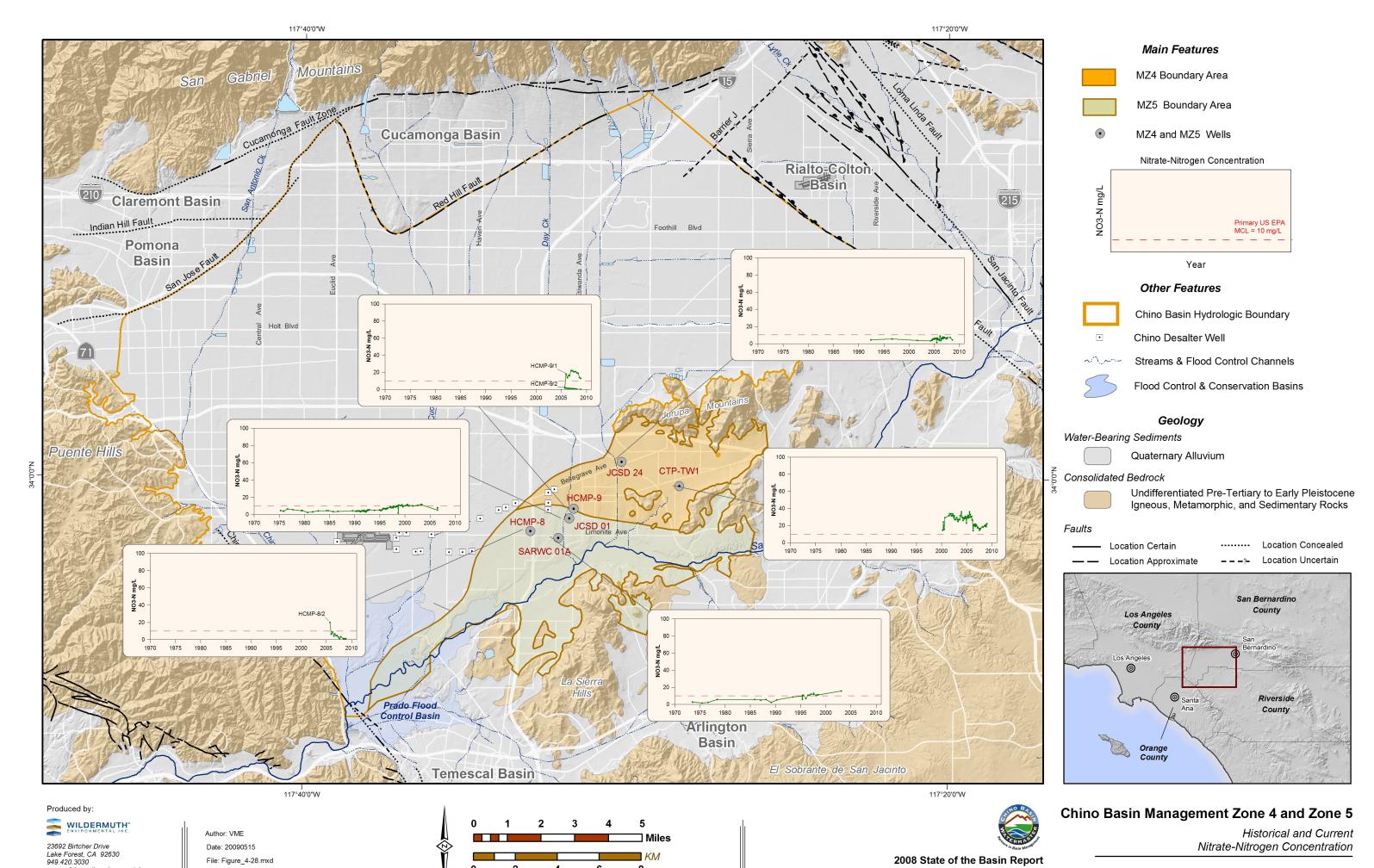












5.1 Background

One of the earliest indications of land subsidence in Chino Basin was the appearance of ground fissures in the City of Chino. These fissures appeared as early as 1973, but an accelerated occurrence of ground fissuring ensued after 1991 and resulted in damage to existing infrastructure (see Figure 5-1). The scientific studies that followed attributed the fissuring phenomenon to differential land subsidence caused by pumping of the underlying aquifer system and the consequent drainage and compaction of aquitard sediments.

5.1.1 OBMP Program Element 4

In 1999, the OBMP Phase I Report (WEI, 1999) identified pumping-induced drawdown and subsequent aquifer-system compaction as the most likely cause of land subsidence and ground fissuring observed in MZ1. Program Element 4 of the OBMP, *Develop and Implement a Comprehensive Groundwater Management Plan for Management Zone 1*, called for the development and implementation of an interim management plan for MZ1 that would:

- Minimize subsidence and fissuring in the short-term.
- Collect the information necessary to understand the extent, rate, and mechanisms of subsidence and fissuring.
- Formulate a management plan to abate future subsidence and fissuring or reduce it to tolerable levels.

In 2000, the Implementation Plan in the Peace Agreement called for an aquifer-system and land subsidence investigation in the southwestern region of MZ1 to support the development of a management plan for MZ1 (second and third bullets above). This investigation was titled the MZ1 Interim Monitoring Program (IMP). From 2001-2005, Watermaster developed, coordinated, and conducted the IMP under the guidance of the MZ1 Technical Committee, which is composed of representatives from all major MZ1 producers and their technical consultants. Specifically, the producers represented on the MZ1 Technical Committee include: the Agricultural Pool, the Cities of Chino, Chino Hills, Ontario, Pomona, and Upland; the Monte Vista Water District; the Southern California Water Company; and the State of California (CIM).

The main conclusions derived from the IMP were:

- 1. Groundwater production from the deep confined aquifer system in this area causes the greatest stress to the aquifer system. In other words, pumping of the deep aquifer system causes water level drawdowns that are much greater in magnitude and lateral extent than drawdowns caused by pumping of the shallow aquifer system.
- 2. Water level drawdowns due to pumping of the deep aquifer system can cause inelastic (permanent) compaction of the aquifer-system sediments, which results in permanent land subsidence. The initiation of inelastic compaction within the aquifer system was identified during this investigation when water levels fell below



- a depth of about 250 feet in the PA-7 piezometer at Ayala Park.
- 3. The current state of aquifer-system deformation in south MZ1 (in the vicinity of Ayala Park) is essentially elastic. Very little inelastic (permanent) compaction is now occurring in this area, which is in contrast to the recent past when about 2.2 feet of land subsidence, accompanied by ground fissuring, occurred from about 1987 to 1995.
- 4. During this study, a previously undetected barrier to groundwater flow was identified. This barrier is located within the deep aquifer system and is aligned with the historical zone of ground fissuring. Pumping from the deep aquifer system is limited to the area west of the barrier, and the resulting drawdowns do not propagate eastward across the barrier. Thus, compaction occurs within the deep system on the west side of the barrier but not on the east side, which causes concentrated differential subsidence across the barrier and creates the potential for ground fissuring.
- 5. InSAR and ground level survey data indicate that permanent subsidence in the central region of MZ1 (north of Ayala Park) has occurred in the past and continues to occur today. The InSAR data also suggest that the groundwater barrier extends northward into central MZ1. These observations suggest that the conditions that very likely caused ground fissuring near Ayala Park in the 1990s are also present in central MZ1 and should be studied in more detail.

The investigation methods, results, and conclusions (listed above) are described in detail in the MZ1 Summary Report (WEI, 2006b). The investigation provided enough information for Watermaster to develop Guidance Criteria for the MZ1 producers in the investigation area that, if followed, would minimize the potential for subsidence and fissuring during the completion of the MZ1 Subsidence Management Plan (MZ1 Plan). The Guidance Criteria formed the basis for the MZ1 Plan, which was developed by the MZ1 Technical Committee and approved by Watermaster in October 2007. In November 2007, the California Superior Court, which retains continuing jurisdiction over the Chino Basin Adjudication, approved the MZ1 Plan and ordered its implementation.

The MZ1 Plan includes a listing of Managed Wells subject to the plan, a map of the so-called Managed Area in southern MZ1, an initial threshold water level (Guidance Level) at an index well in the Managed Area (245 feet below the top of the PA-7 well casing at Ayala Park in Chino [ft-brp]), and a plan for ongoing monitoring and annual reporting.

5.1.2 OBMP Program Element 1

The OBMP Phase I Report also noted that land subsidence was occurring in other parts of the basin besides Chino. Program Element 1 (PE1) of the OBMP and the Implementation Plan, *Develop and Implement a Comprehensive Monitoring Program*, called for basin-wide analysis of land subsidence via ground-level surveys and InSAR and ongoing monitoring based on the analysis of the subsidence data. Through 2008, basin-wide monitoring has been based on the ground-level survey data and InSAR data collected as part of the IMP and the MZ1 Plan implementation.



5.2 Ground-Level Monitoring Program

Implementation of the MZ1 Plan began in 2008. The MZ1 Plan calls for (1) the continued scope and frequency of monitoring implemented during the IMP within the MZ1 Managed Area and (2) expanded monitoring of the aquifer system and land subsidence in other areas of the Chino Basin where the IMP indicated concern for future subsidence and ground fissuring. The expanded monitoring efforts outside of the MZ1 Managed Area are consistent with the requirements PE1.

Watermaster's current ground-level monitoring program includes:

- Piezometric Levels. Piezometric levels are an important part of the ground-level monitoring program because piezometric changes are the mechanism for aquifer-system deformation and land subsidence. Watermaster monitors piezometric levels at about 33 wells in MZ1. Currently, a pressure-transducer/data-logger is installed at each of these wells and records one water level reading every 15 minutes. And, Watermaster records depth-specific water levels at the piezometers located at the Ayala Park Extensometer facility every 15 minutes.
- Aquifer-System Deformation. Watermaster records aquifer-system deformation at the Ayala Park Extensometer facility (see Figure 5-1). At this facility, two extensometers, completed at 550 ft-bgs and 1,400 ft-bgs, record the vertical component of aquifer-system compression and/or expansion once every 15 minutes (synchronized with the piezometric measurements).
- Vertical Ground-Surface Deformation. Watermaster monitors vertical ground-surface
 deformation via the ground-level surveying and remote sensing (InSAR) techniques
 established during the IMP. Currently, ground-level surveys are being conducted in the
 MZ1 Managed Area once per year. InSAR is the only monitoring technique being
 employed outside the MZ1 Managed Area, and InSAR data is analyzed once per year.
- Horizontal Ground-Surface Deformation. Watermaster monitors horizontal ground-surface displacement across the eastern side of the subsidence trough and the adjacent area east of the barrier/fissure zone. These data, obtained by electronic distance measurements (EDMs), are used to characterize the horizontal component of land surface displacement caused by groundwater production on either side of the fissure zone. Currently, Watermaster is collecting EDMs at a semiannual frequency (Spring/Fall) between east/west aligned benchmarks on Eucalyptus, Edison, Schaefer, and Philadelphia Avenues.

5.3 Results of Ground-Level Monitoring Program

At the conclusion of each fiscal year, the MZ1 Plan requires that Watermaster produce an MZ1 Annual Report that includes the results of the past year's monitoring. The 2008 MZ1 Annual Report (currently in preparation) will be the first such report published by Watermaster and will focus primarily on the intensive monitoring being conducted in the MZ1 Managed Area.

The ground-level monitoring results described below will focus primarily on the ground-level



survey and InSAR monitoring being conducted across the entire Chino Basin (PE1).

5.3.1 InSAR

Figure 5-2 is a map of the Chino Basin that shows InSAR results for 2005-2008. The InSAR data are generally coherent and useful in the northern urbanized areas of the basin but are generally incoherent and not as useful in the southern agricultural areas (light brown areas in Figure 5-2). This pattern of "coherence" relative to land use is typical of InSAR data.

Figure 5-2 shows that ground motion during 2005-2008 was relatively minor (less than about -0.02 ft of subsidence) in the northeastern parts of the basin, such as Fontana and Rancho Cucamonga. However, in northwestern parts of the basin, land subsidence of over -0.14 ft and -0.12 ft have been measured by InSAR in Pomona and Ontario, respectively.

Figure 5-2 also shows that ground motion is influenced by geologic faults that cut through the aquifer system and act as barriers to groundwater flow. For instance, the land surface elevation has increased (uplift) in the southern portion of the Cucamonga Basin—just north of the Red Hill Fault. The San Jose Fault is clearly influencing the pattern of ground motion in the Claremont, Pomona, and Chino Basins. Of most concern, with respect to the potential for ground fissuring, is the differential ground motion across the San Jose Fault between the Pomona and Chino Basins.

Historically, the City of Chino has experienced the most land subsidence (e.g. over -2.0 ft of subsidence within the MZ1 Managed Area during 1987-1999), but for 2005-2008, the InSAR data indicate that land subsidence was relatively minor in this area (less than about -0.04 ft).

5.3.2 Ground-Level Surveys

Figure 5-3 is a map of the western half of Chino Basin that shows both the InSAR and ground-level survey results for 2005-2008. The ground-level survey data generally corroborate the patterns and magnitude of ground motion shown in the InSAR data with a few exceptions:

- The ground-level survey data indicate a greater magnitude of land subsidence in the MZ1 Managed Area (maximum subsidence = -0.10 ft) than the InSAR data (maximum subsidence = -0.05 ft).
- In some areas, the ground-level survey data indicate minor subsidence while the InSAR data indicate minor uplift. In these instances, the difference between the ground-level survey and InSAR data is generally less than about 0.05 ft.

One advantage of the ground-level survey data is that it can provide information on ground motion in areas where InSAR data is absent. See, for example, the area shown on Figure 5-3 near at the intersection of Euclid Avenue and Kimball Avenue where the Chino I Desalter wells pump groundwater from the deep aquifer system. The survey data indicated maximum land subsidence of -0.24 ft in this area during 2005-2008.



5.4 Analysis of Ground Surface Displacement

Historical ground motion data (shown in Figure 5-1) and recent ground motion data (shown in Figures 5-2 and 5-3) indicate that land subsidence concerns in the Chino Basin are confined to certain portions of MZ1 and MZ2. These "areas of subsidence concern" are delineated and labeled in Figures 5-2 and 5-3. Besides the MZ1 Managed Area, Watermaster has designated four additional areas of subsidence concern: the Central MZ1 Area, the Pomona Area, the Ontario Area, and the Southeast Area.

The recent land subsidence that has been occurring in each of these areas is mainly controlled by recent and/or historical changes in groundwater levels, which, in turn, are mainly controlled by pumping and recharge.

Below, the relationships between groundwater pumping, aquifer recharge, groundwater levels, and ground motion, which help to reveal cause and effect; the current state of ground motion; and the nature of current land subsidence (i.e. elastic and/or inelastic, differential, etc.), are discussed by area of concern.

5.4.1 MZ1 Managed Area

Within the MZ1 Managed Area, pumping of the deep confined aquifer system causes water level drawdowns that are much greater in magnitude and lateral extent than drawdowns caused by pumping of the shallow aquifer system. Artificial recharge in the northern portions of MZ1 appears to have no immediate impact on groundwater levels in the deep aquifer system in the MZ1 Managed Area. These conclusions were established during the IMP (WEI, 2006b) and are shown graphically in Figure 5-4.

Figures 5-4 and 5-5 also show vertical ground motion at the Deep Extensometer at Ayala Park and at a benchmark monument (137/53) at the corner of Schaefer Avenue and Central Avenue. About -2.5 ft of subsidence occurred in portions of the MZ1 Managed Area from 1987-2000, but very little inelastic subsidence has occurred since 2000, and no additional ground fissuring has been observed.

Another conclusion of the IMP was that groundwater-level drawdowns due to pumping of the deep aquifer system can cause inelastic (permanent) compaction of the aquifer-system sediments, which results in permanent land subsidence. The initiation of inelastic compaction within the aquifer system was identified during the IMP when water levels fell below a depth of about 250 feet in the PA-7 piezometer at Ayala Park. From 2005 to 2008, water levels at PA-7 did not decline below 250 ft-brp, and very little, if any, inelastic compaction was recorded in the MZ1 Managed Area. Data from the MZ1 Managed Area are further analyzed in the 2008 MZ1 Annual Report (in preparation).

The IMP also identified a previously undetected barrier to groundwater flow on the east side of the MZ1 Managed Area. This barrier is located within the deep aquifer system and is aligned with the historical zone of ground fissuring (see Figure 5-3). Pumping from the deep aquifer system has been limited to the area west of the barrier, and the resulting drawdowns have not propagated eastward across the barrier. Thus, historical compaction occurred within the deep system on the west side of the barrier but not on the east side. Concentrated



differential subsidence across the barrier is the most likely cause of the ground fissuring observed in the early 1990s. The rate of land subsidence decreased to almost zero in the MZ1 Managed Area in the mid-1990s, and no additional ground fissuring has been observed.

5.4.2 Central MZ1 Area

The Central MZ1 Area is located directly north of the MZ1 Managed Area (see Figure 5-3). Figures 5-6 and 5-7 display time histories of groundwater pumping, aquifer recharge, groundwater levels, and ground motion in the Central MZ1 Area.

The ground motion time histories for Central MZ1 is similar to that of the MZ1 Managed Area—as much as -2.2 ft of inelastic subsidence occurred at the corner of Philadelphia and Monte Vista Avenue from 1987-2000, but very little inelastic subsidence has occurred since 2000. This similarity suggests a relationship to the causes of land subsidence in the MZ1 Managed Area; however, there is very little historical groundwater level data in this area to confirm this relationship.

Most of the wells with historical groundwater level records are in the northern part of Central MZ1 (see Figure 5-3) where historical subsidence was not as pronounced. From about 1935 to 1978, groundwater levels in these wells declined by about 150 ft. Groundwater levels increase by about 50 ft during the 1980s and remained relatively stable until 2005. Since 2005, groundwater levels have increased by about 25 ft, which is likely due to decreased pumping and increased recharge in MZ1.

5.4.3 Pomona Area

The Pomona Area is located directly north of the Central MZ1 Area (see Figure 5-3). Figures 5-8 and 5-9 display time histories of groundwater pumping, aquifer recharge, groundwater levels, and ground motion in the Pomona Area.

The ground motion time histories of the Pomona Area is based solely on InSAR data from 1992 to 1995, 1995 to 2000, and 2005 to 2008. These data indicate that land subsidence has occurred continuously in this area, generally at a rate of about 0.07 ft/yr. The rate of subsidence appears to be decreasing gradually with time.

From about 1935 to 1978, groundwater levels in the Pomona Area declined by about 175 ft or more. Groundwater levels increased by about 50 to 100 ft during the 1980s. From about 1990 to 2004, groundwater levels declined again by about 25 to 50 ft. And from 2004 to 2008, groundwater levels increased by about 25 to 50 ft. The groundwater level changes from 1990 to 2008 appear to be closely related to pumping and recharge in MZ1.

The observed, continuous land subsidence cannot be explained entirely by the corresponding changes in groundwater levels during this time (1992-2008). A plausible explanation for the subsidence is that thick, slowly-draining aquitards are compacting in response to the historical drawdowns that occurred from 1935 to 1978 (see Figure 5-9).

Lastly, the InSAR data in Figure 5-3 shows a steep gradient of subsidence across the San Jose Fault, indicating the potential for the accumulation of horizontal strain in the shallow



sediments and the possibility of ground fissuring. Ground fissuring is the main subsidence-related threat to infrastructure.

5.4.4 Ontario Area

The Ontario Area is located east of the Central MZ1 and the Pomona Areas (see Figure 5-3). Figures 5-10 and 5-11 display time histories of groundwater pumping, aquifer recharge, groundwater levels, and ground motion in the Ontario Area.

The ground motion time histories of the Ontario Area is based solely on InSAR data from 1992 to 1995, 1995 to 2000, and 2005 to 2008. These data indicate that land subsidence has occurred continuously in this area, generally at a rate of about 0.06 ft/yr. The rate of subsidence appears to be decreasing gradually with time.

From about 1935 to 1978, groundwater levels in the Ontario Area declined by about 125 ft. Groundwater levels increased by about 10 to 20 ft during the early 1980s and have remained relatively stable since then.

The observed continuous land subsidence from 1992 to 2008 is not explained by the relatively stable groundwater levels. A plausible explanation for the subsidence is that thick, slowly-draining aquitards are compacting in response to the historical drawdowns that occurred from 1935 to 1978 (see Figure 5-11).

5.4.5 Southeast Area

The Southeast Area is located east of the MZ1 Managed Area (see Figure 5-3). Figures 5-12 and 5-13 display time histories of groundwater pumping, aquifer recharge, groundwater levels, and ground motion in the Southeast Area.

The ground motion time histories of the Southeast Area is based solely on ground-level surveys performed from 1987to 2008. These data indicate that land subsidence has occurred continuously and slowly in this area, generally at a rate of about 0.02 ft/yr. However, the data also indicate that from 2005 to 2008 about -0.24 ft of subsidence occurred near the western portion of the Chino I Desalter well field where these wells are pumping from and causing drawdown within the deep confined aquifer system.

There is very little historical groundwater level data for this area prior to about 1990. The data since 1990 indicate relatively stable groundwater levels.

The observed slow but continuous land subsidence from 1987 to 2008 is not explained by the relatively stable groundwater levels. A plausible explanation for the subsidence is that thick, slowly-draining aquitards are compacting in response to the historical drawdowns that likely occurred prior to 1990.

Lastly, the first ground fissures ever documented in the Chino Basin occurred in the Southeast Area in the early 1970s, but ground fissuring has not been observed in the Southeast Area since then.



5.5 Conclusions and Recommendations

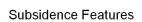
The conclusions and recommendations for Watermaster's basin-wide ground-level monitoring program are provided below.

- Land subsidence does not appear to be a concern in the eastern and northernmost portions of Chino Basin. In these areas, the underlying aquifer system is composed primarily of coarse-grained sediments that are not prone to compaction.
- Land subsidence and the potential for ground fissuring are major concerns in the western and southern portions of the Chino Basin. In these areas, the underlying aquifer system consists of interbedded, fine-grained sediment layers (aquitards) that can drain and compact when groundwater levels decline in the adjacent coarse-grained aquifers. Ground fissuring has occurred in the past where land subsidence was differential (i.e. steep gradient of subsidence). Ground fissuring is the main subsidence-related threat to infrastructure.
- Land subsidence has been persistent across most of the western and southern portions of the Chino Basin since, at least, 1987 when land subsidence monitoring began. In many of these areas, land subsidence continues even during periods of groundwater level recovery, indicating that thick, slowly-draining aquitards are compacting in response to the large historical drawdowns of 1935 to 1978.
- Pumping-induced drawdown has caused accelerated occurrences of land subsidence in the recent past, including subsidence in the City of Chino during the early 1990s and, currently, in the vicinity of the Chino I Desalter well field. Watermaster should anticipate similar occurrences of land subsidence in areas (1) that are prone to subsidence and (2) where drawdown will occur in the future.
- Watermaster will continue its basin-wide ground-level monitoring program, using InSAR and ground-level surveys. Watermaster will consider expanding the ground-level surveys to cover the area of the proposed Chino Creek Desalter Well Field. This is an area that is prone to subsidence, where drawdown may occur near where ground fissuring has occurred in the past, and where InSAR data is not currently available. Watermaster will also consider expanding the ground-level surveys to cover the Pomona and Ontario Areas. In general, InSAR data coverage is continuous and of high quality throughout both areas, so ground-level surveys would primarily provide supporting and confirmation data for the InSAR and would occur at a frequency of once every three to five years.
- Watermaster will consider installing low-cost piezometer/extensometer facilities at appropriate locations in all Areas of Subsidence Concern. This type of facility has been successfully constructed and tested at Ayala Park in Chino. Such facilities record the requisite data (1) to monitor land subsidence and groundwater levels at high resolution and accuracy, (2) to provide the information necessary to characterize the elastic and/or inelastic nature of any land subsidence occurring in an area, (3) to provide the information necessary to develop criteria to manage subsidence, and (4) to provide the information necessary to characterize aquifer and aquitard properties that could be used in a predictive computer-simulation model of subsidence.



- Watermaster will consider building and calibrating predictive computer-simulation models of subsidence across all Areas of Subsidence Concern in the Chino Basin. These models would provide information on the rates and ultimate magnitude of land subsidence that could be associated with various basin management planning scenarios (i.e. pumping and recharge patterns). This information would be valuable to affected Watermaster parties.
- Because ground fissuring caused by differential land subsidence is the main threat to infrastructure, Watermaster will periodically inspect for signs of ground fissuring in areas that are experiencing differential land subsidence. In addition, Watermaster will consider monitoring the horizontal strain across these zones of potential ground fissuring in an effort to better understand and manage ground fissuring.





0.0
Contours of Relative Change in Land Surface Altitude
as Measured by Leveling Surveys
1987 - 1999
(feet)

-2.2

+ 1.0 Relative Change in Land Surface Altitude 0.0 as Measured by InSAR Oct 1993 - Dec 1995 - 1.0 (feet)

No InSAR Data

Active Production Wells in MZ-1 by Owner

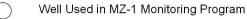
- OntarioPomona
- CIMChino Hills

117°40'0''W

- SAWC
- Chino HiChino
- Upland
- ChinoMVWD

• SCWC

Other Features





Proposed Central MZ-1 Piezometer

Chino Basin Desalter Well (Existing)

Management Zone 1 Boundary

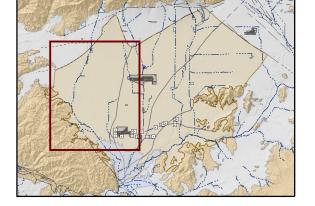
Prepared by: WILDERMUTH ENVIRONMENTAL INC. 23692 Birtcher Drive Lake Forest, CA 92630 949.420.3030 www.wildermuthenvironmental.com



Author: AEM

Date: 20090621

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Historical Land Surface Deformation in Management Zone 1

Leveling Surveys (1987-99) and InSAR (1993-95)

1 2 3 4 5 6 7

23692 Birtcher Drive Lake Forest, CA 92630 949.420.3030

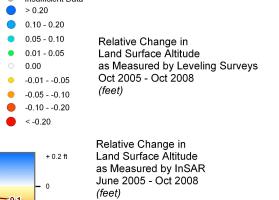
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Figure 5-2

2008 State of the Basin Report

Ground-Level Monitoring



Brown areas represent regions where InSAR data is absent (incoherent)

Water Level Wells (in Figures 5-4 to 5-13)
 Chino Basin Desalter Well (Existing)
 □ Proposed Chino Creek Desalter Well
 △ Survey and InSAR Measurement Points (in Figures 5-4 to 5-13)
 Chino Basin Management Zones
 Subsidence Areas of Interest
 MZ1 Managed Area



Vertical Ground Motion (2005-2008)

Prepared by:



- 0.2



Author: ETL
Date: 20090617
File: Figure_5-3.mxd

Leveling Surveys and InSAR in Western Chino Basin

Groundwater Levels versus Ground Levels in the MZ1 Managed Area - 1993 to 2009 Production & Recharge in MZ1
Wet Water Recharge **Ground Levels Groundwater Levels** Annual Recharge (acre-ft) - 2.5 at Wells (screened intervals) BM 137/53 Survey Measurements PA-7 (438-448 ft-bgs) **Groundwater Production** Deep Extensometer at Ayala Park CH-19 (340-1000 ft-bgs) C-06 (200-375 ft-bgs) -100 - 1.5 -200 Depth to Water (ft below reference point) 10,000 20,000 Vertical Ground Motion (ft) Annual Production (acre-ft)
70,000 60,000 50,000 40,000 30,000 20,000 10,000 450 -500 - -1.5 -550 -2 80,000 -600 Note: Discontinuities in the time series of ground levels are -2.5 represented by broken lines. The displacement that occurred during each discontinity is assumed to be zero, which may not be a valid assumption. ,0g1

Figure 5-4

Groundwater Levels versus Ground Levels in the MZ1 Managed Area - 1935 to 2009 100-150 200 250 Vertical Ground Motion (#) 600-650-Ground Levels BM 137/53 Survey Measurements 700 **Groundwater Levels** 750at Wells (screened intervals) PA-7 (438-448 ft-bgs) CH-19 (340-1000 ft-bgs) 800 Note: Discontinuities in the time series of ground levels are C-06 (200-375 ft-bgs) represented by broken lines. The displacement that occurred during each discontinity is assumed to be zero, which may not be a valid assumption. 850

Figure 5-5

Groundwater Levels versus Ground Levels in the Central MZ1 Area - 1993 to 2009 Production & Recharge in MZ1
Wet Water Recharge **Groundwater Levels Ground Levels** 90,000 - 2.5 at Wells (screened intervals) BM 125/49 Survey Measurements MV-02 (397-962 ft-bgs) **Groundwater Production** BM A-4 Survey Measurements -100 Annual Recharge (acre-ft) MV-24 (244-420 ft-bgs) Central-MZ1 InSAR Measurements C-10 (355-1090 ft-bgs) -150 - 1.5 -200 -250 Depth to Water (ft below reference point) 10,000 20,000 Vertical Ground Motion (ft) Annual Production (acre-ft)
70,000 60,000 50,000 40,000 30,000 20,000 10,000 -500 550 -1.5 -600 -2 80,000 -650 Note: Discontinuities in the time series of ground levels are -2.5 represented by broken lines. The displacement that occurred during each discontinity is assumed to be zero, which may not be a valid assumption.

Figure 5-6

Groundwater Levels versus Ground Levels in the Central MZ1 Area - 1935 to 2009 150 200 250 300 - 1.5 350-400 Depth to Water (ft below reference point) Vertical Ground Motion (#) 700 750-**Ground Levels** Central-MZ1 InSAR Measurements 800-BM A-4 Survey Measurements City BM 125/49 Survey Measurements 850-**Groundwater Levels** at Wells (screened intervals) 900 Chino 03 (230-450 ft-bgs) MV-02 (397-962 ft-bgs) 950 MV-14 (unknown) Note: Discontinuities in the time series of ground levels are -2.5 represented by broken lines. The displacement that occurred during MV-24 (244-420 ft-bgs) each discontinity is assumed to be zero, which may not be a valid assumption. 1000

Figure 5-7

Figure 5-8 Groundwater Levels versus Ground Levels in the Pomona Area - 1993 to 2009 -150 Production & Recharge in MZ1
Wet Water Recharge **Groundwater Levels Ground Levels** Annual Recharge (acre-ft) at Wells (screened intervals)

MV-19 (620-1230 ft-bgs) Pomona InSAR Measurements **Groundwater Production** -200 P-11 (168-550 ft-bgs) P-27 (472-849 ft-bgs) P-30 (565-875 ft-bgs) -250 - 1.5 300 -350 Depth to Water (it below reference point) 20,000 Vertical Ground Motion (#) 10,000 Annual Production (acre-ft)
70,000 60,000 50,000 40,000 30,000 20,000 10,000 -600 -650 - -1.5 -700 -2 80,000 -750 Note: Discontinuities in the time series of ground levels are -2.5 represented by broken lines. The displacement that occurred during each discontinity is assumed to be zero, which may not be a valid assumption.

Groundwater Levels versus Ground Levels in the Pomona Area - 1935 to 2009 150 200 250 300 350 400 Depth to Water (ft below reference point) Vertical Ground Motion (#) 750-**Ground Levels** Pomona InSAR Measurements 800-**Groundwater Levels** 850at Wells (screened intervals) MV-08 (225-447 ft-bgs) -2 900-MV-10 (unknown) MV-13 (203-475 ft-bgs) P-11 (168-550 ft-bgs) 950 Note: Discontinuities in the time series of ground levels are -2.5 represented by broken lines. The displacement that occurred during P-30 (565-875 ft-bgs) each discontinity is assumed to be zero, which may not be a valid assumption. 1000

Figure 5-9

Groundwater Levels versus Ground Levels in the Ontario Area - 1993 to 2009 -150 Production & Recharge in MZ1
Wet Water Recharge **Ground Levels Groundwater Levels** Annual Recharge (acre-ft) - 2.5 at Wells (screened intervals) Ontario InSAR Measurements C-14 (480-1200 ft-bgs) **Groundwater Production** -200 O-15 (474-966 ft-bgs) O-34 (522-1092 ft-bgs) -250 - 1.5 -300 -350 Depth to Water (it below reference point) Vertical Ground Motion (ft) Annual Production (acre-ft)
70,000 60,000 50,000 40,000 30,000 20,000 10,000 -600 -650 - -1.5 -700 -2 80,000 -750 Note: Discontinuities in the time series of ground levels are -2.5 represented by broken lines. The displacement that occurred during each discontinity is assumed to be zero, which may not be a valid assumption.

Figure 5-10

Groundwater Levels versus Ground Levels in the Ontario Area - 1930 to 2009 150 250 300 350-400 Depth to Water (ft below reference point) Vertical Ground Motion (#) 700 750-800-**Ground Levels** Ontario InSAR Measurements 850-**Groundwater Levels** at Wells (screened intervals) 900-O-05 (360-470 ft-bgs) O-15 (474-966 ft-bgs) O-34 (522-1092 ft-bgs) 950 Note: Discontinuities in the time series of ground levels are -2.5 represented by broken lines. The displacement that occurred during C-14 (480-1200 ft-bgs) each discontinity is assumed to be zero, which may not be a valid assumption. 1000

Figure 5-11

Groundwater Levels versus Ground Levels in the Southeast Area - 1993 to 2009 Production & Recharge in MZ1
Wet Water Recharge **Ground Levels Groundwater Levels** 90,000 - 2.5 at Wells (screened intervals) BM 137/61 Survey Measurements CH-18A (420-980 ft-bgs) **Groundwater Production** BM 133/61 Survey Measurements Annual Recharge (acre-ft) C-13 (290-720 ft-bgs) -100 - 1.5 -150 -200 Depth to Water (ft below reference point) Vertical Ground Motion (#) Annual Production (acre-ft)
70,000 60,000 50,000 40,000 30,000 20,000 10,000 450 -500 - -1.5 -550 -2 80,000 -600 Note: Discontinuities in the time series of ground levels are -2.5 represented by broken lines. The displacement that occurred during each discontinity is assumed to be zero, which may not be a valid assumption.

Figure 5-12

Groundwater Levels versus Ground Levels in the Southeast Area - 1930 to 2009 50 150 200 - 1.5 250-300 Depth to Water (it below reference point) Vertical Ground Motion (#) 650-700-**Ground Levels** BM 133/61 Survey Measurements 750 BM 137/61 Survey Measurements 800-**Groundwater Levels** at Wells (screened intervals) CH-18A (420-980 ft-bgs) 850 Note: Discontinuities in the time series of ground levels are -2.5 C-13 (290-720 ft-bgs) represented by broken lines. The displacement that occurred during each discontinity is assumed to be zero, which may not be a valid assumption. 900

Figure 5-13

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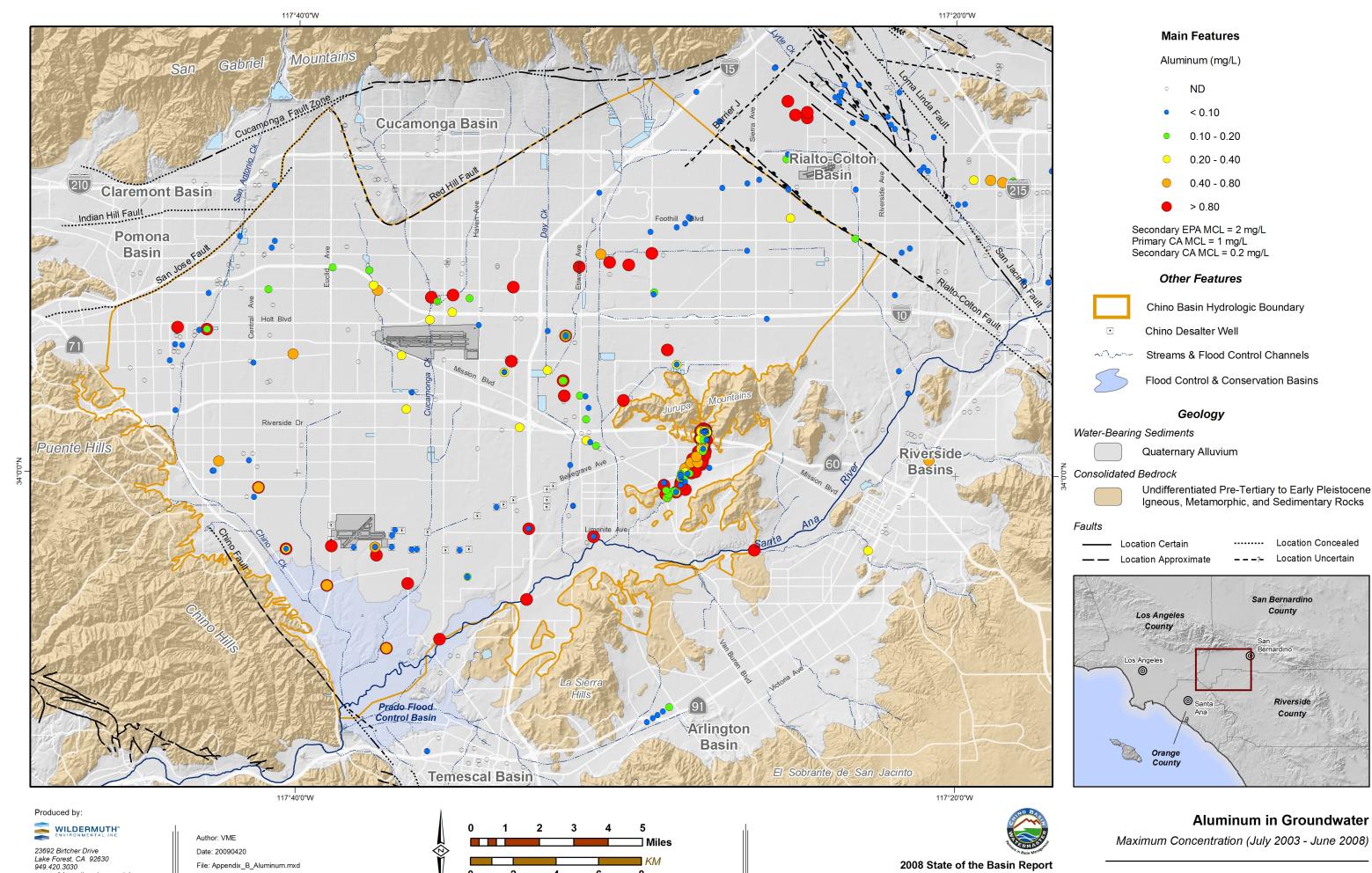


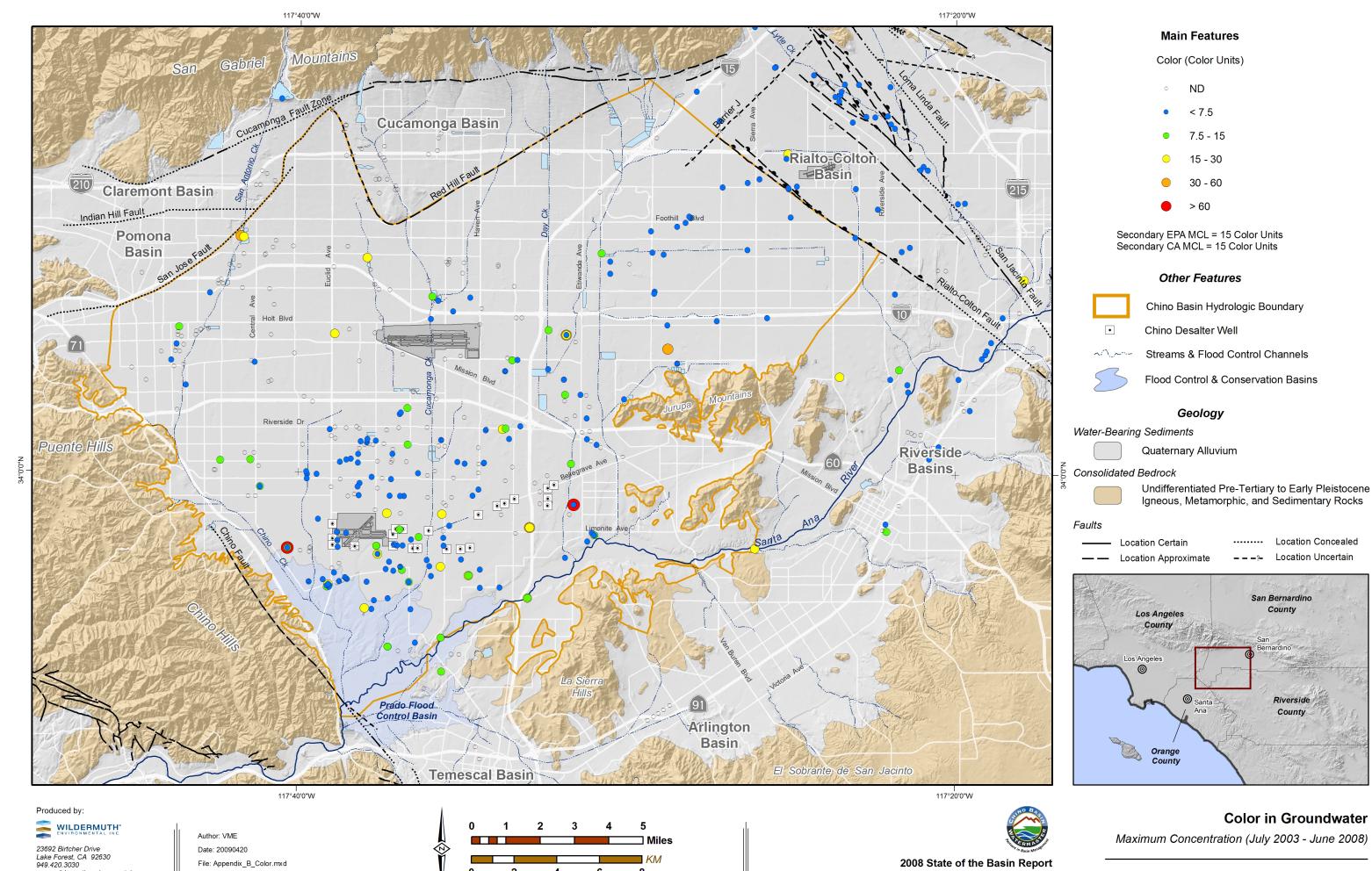
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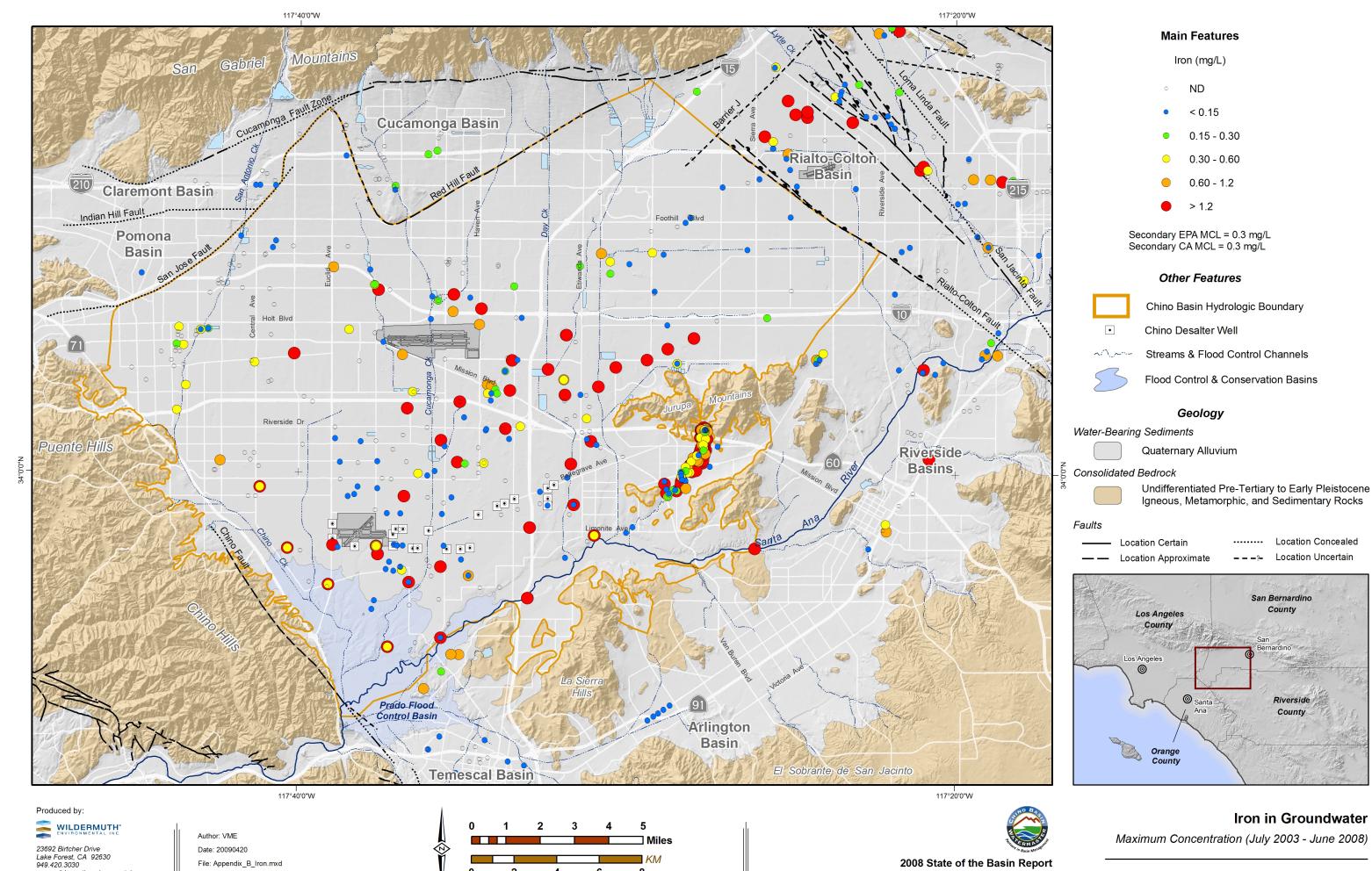


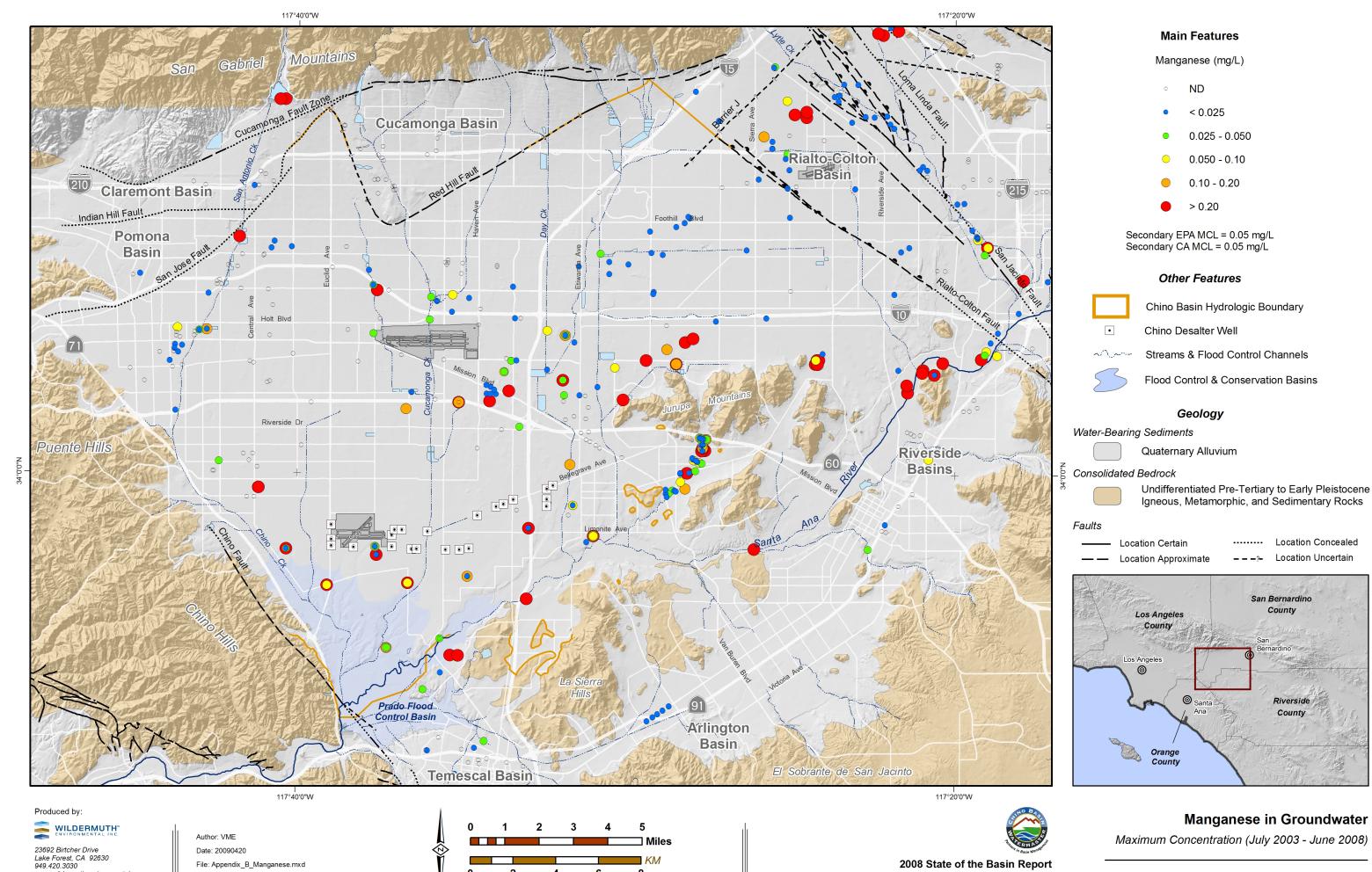
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Appendix A
Groundwater Level Map

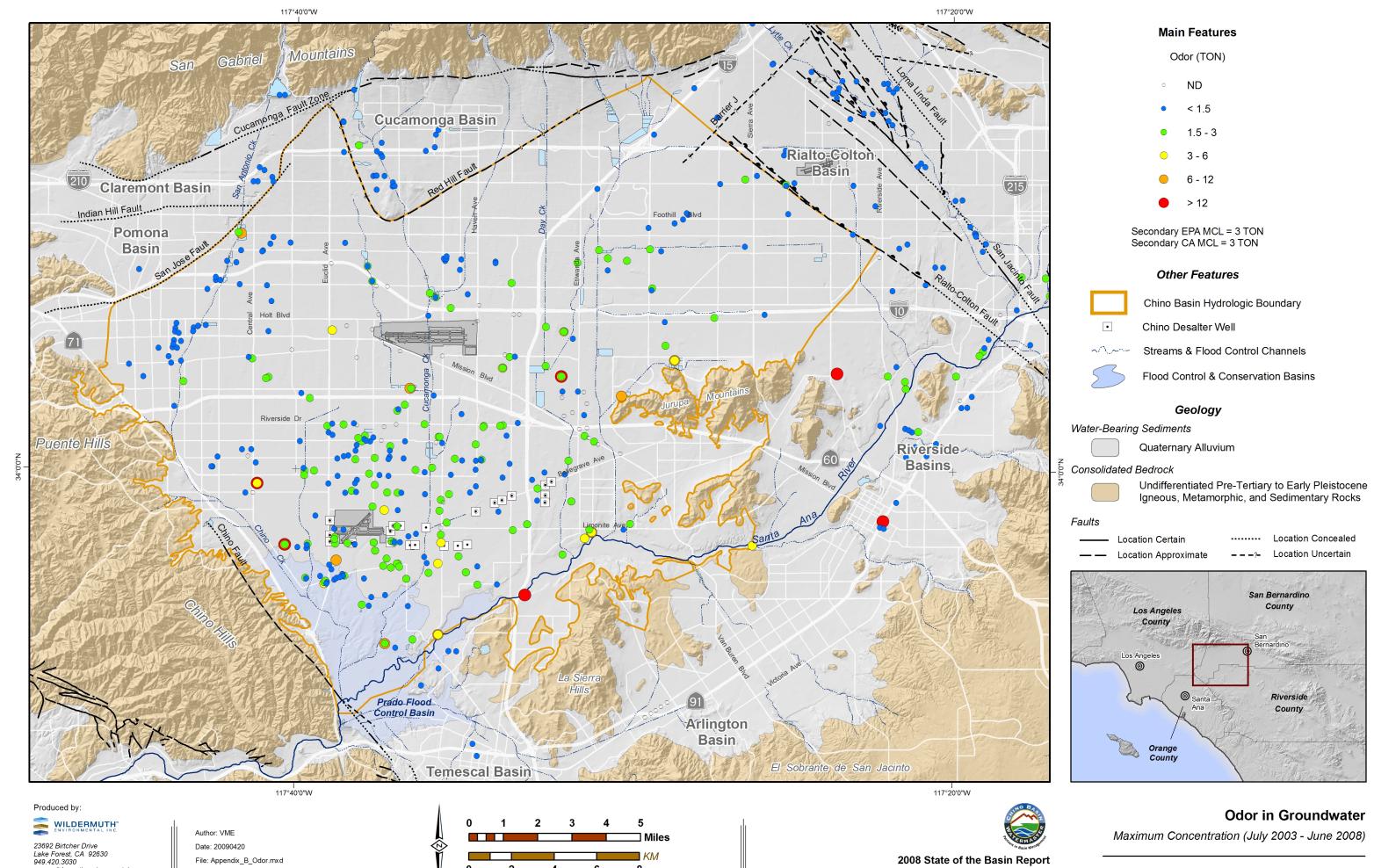


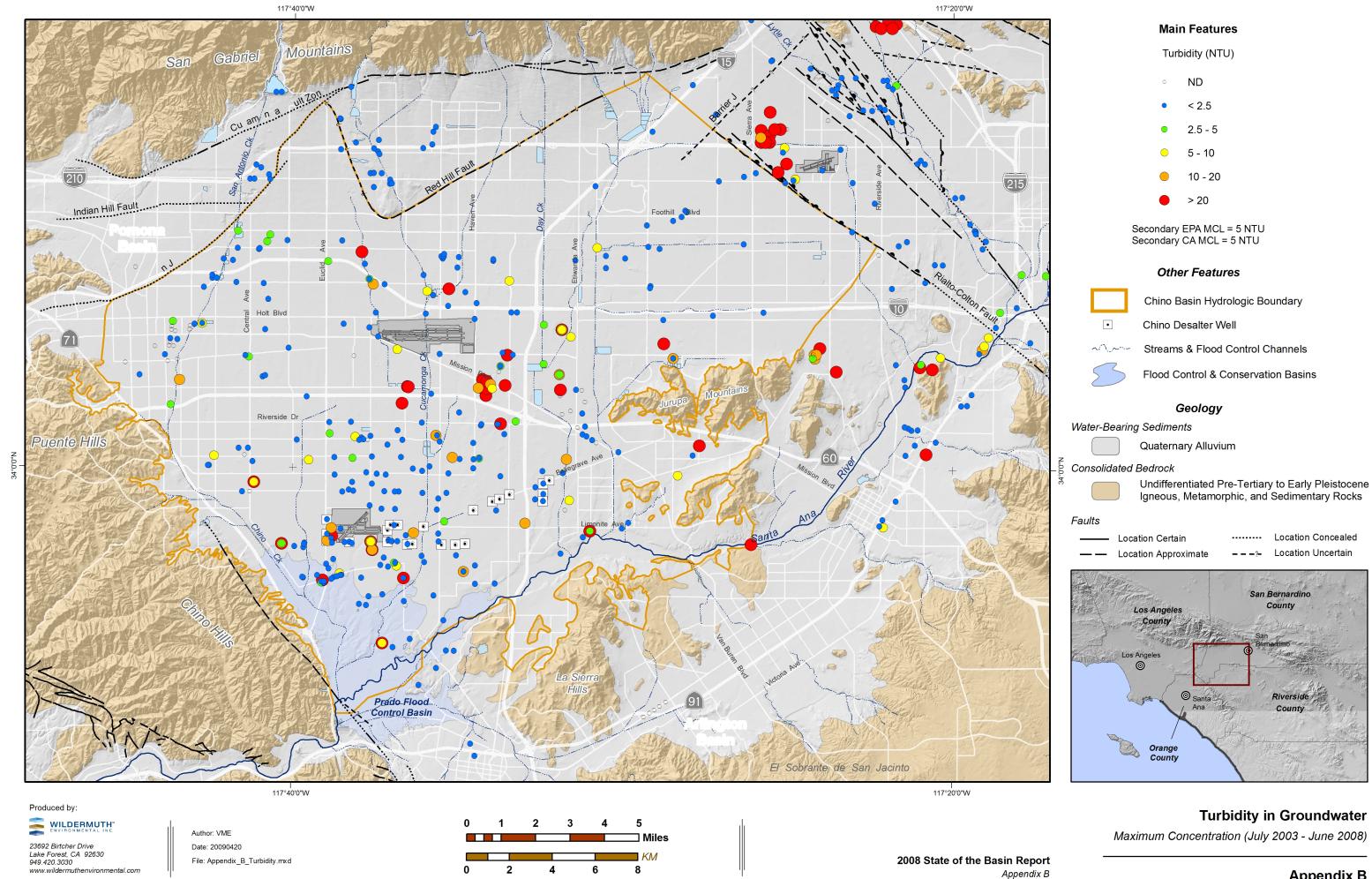














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Chemical		Unit	Primary E	PA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL	
1,1,1-Trichlo	proethane		ug/L	20	0	n/a	200	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.46	0.7	1.02	1.36	4.46	1.446	2641	499	5	0
1,1,2,2-Tetra	achloroethane		ug/L	n/	a	n/a	1	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
						2313	477	0	0
1,1,2-Trichlo	oro-1,2,2-trifluoroe	thane	ug/L	n/	a	n/a	1200	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.5	0.63	1.1	6.5	185	32.488	1694	396	6	0
1,1,2-Trichloroethane		ug/L	5	i	n/a	5	n/a	n/a	
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.11	0.45	0.81	2.3	3.8	1.293	2625	499	5	0
1,1-Dichloroethane		ug/L	n/a		n/a	5	n/a	n/a	
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.1	0.56	1.3	3.4	6013	23.667	2730	509	39	11
1,1-Dichloro	ethene		ug/L	7	•	n/a	6	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.19	1.94	5.4	11.8	190	13.667	2709	507	56	31
1,2,3-Trichlo	oropropane		ug/L	n/	a	n/a	n/a	n/a	0.005
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0	0.012	0.13	0.94	3.1	0.491	1192	375	25	23
1,2,4-Trichlo	orobenzene		ug/L	7	0	n/a	5	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.5	0.5	0.5		0.5	0.5	1008	285	1	0
1,2,4-Trimet	hylbenzene		ug/L	n/	a	n/a	n/a	n/a	330
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
						2062	440	0	0
1,2-Dibromo	o-3-chloropropane		ug/L	0.	2	n/a	0.2	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.01	0.11	0.16	0.24	0.639	0.185	880	301	16	4

WILDERMUTH™ ENVIRONMENTAL INC.

Chemical		Unit	Primary E	PA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL	
1,2-Dichloro	benzene		ug/L	60	0	n/a	600	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1200	# of Wells Sampled 292	# of Wells with Detects 0	# of Wells with Exceedances 0
1,2-Dichloro	ethane		ug/L	5		n/a	0.5	n/a	n/a
<i>Min</i> 0.1	1st Quartile 0.34	Median 0.45	3rd Quartile 0.6	<i>Maximum</i> 3.1	Average 0.611	# of Samples 2714	# of Wells Sampled 508	# of Wells with Detects 27	# of Wells with Exceedances 17
1,2-Dichloro	propane		ug/L	5	j	n/a	5	n/a	n/a
<i>Min</i> 0.12	1st Quartile 0.36	Median 0.5	3rd Quartile 1.1	Maximum 3.6	Average 0.933	# of Samples 2607	# of Wells Sampled 502	# of Wells with Detects 25	# of Wells with Exceedances 0
1,3,5-Trimethylbenzene		ug/L	n/	'a	n/a	n/a	n/a	330	
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1538	# of Wells Sampled 373	# of Wells with Detects 0	# of Wells with Exceedances 0
1,3-Dichloropropene		ug/L	n/a		n/a	0.5	n/a	n/a	
<i>Min</i> 94	1st Quartile 94	Median 94	3rd Quartile	<i>Maximum</i> 96.5	Average 95.25	# of Samples 790	# of Wells Sampled 238	# of Wells with Detects 2	# of Wells with Exceedances 2
1,4-Dichloro	benzene		ug/L	7	5	n/a	5	n/a	n/a
<i>Min</i> 0.13	1st Quartile 0.15	Median 0.17	3rd Quartile 0.21	Maximum 0.57	Average 0.215	# of Samples 1271	# of Wells Sampled 295	# of Wells with Detects	# of Wells with Exceedances 0
1,4-Dioxane			ug/L	n/	'a	n/a	n/a	n/a	3
<i>Min</i> 0.1	1st Quartile 0.29	Median 0.5	3rd Quartile 0.99	Maximum 46	Average 1.289	# of Samples 577	# of Wells Sampled 63	# of Wells with Detects 10	# of Wells with Exceedances
2,3,7,8-Tetra	chlorodibenzo-p-	dioxin	ug/L	3E-	∙05	n/a	3E-05	n/a	n/a
Min 0	1st Quartile 0	Median 0	3rd Quartile	<i>Maximum</i> 0	Average 0	# of Samples 192	# of Wells Sampled 98	# of Wells with Detects	# of Wells with Exceedances 0
2,4-Dichloro	phenoxyacetic ac	id	ug/L	7	0	n/a	70	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 227	# of Wells Sampled 118	# of Wells with Detects 0	# of Wells with Exceedances 0
2-Chlorotolu	ene		ug/L	n/	'a	n/a	n/a	n/a	140
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1531	# of Wells Sampled 364	# of Wells with Detects 0	# of Wells with Exceedances 0



Chemical	Chemical		Unit	Primary E	EPA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL
4-Chlorotolu	iene		ug/L	n/	'a	n/a	n/a	n/a	140
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1532	# of Wells Sampled 365	# of Wells with Detects 0	# of Wells with Exceedances 0
Alachlor			ug/L	2	2	n/a	2	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 262	# of Wells Sampled 129	# of Wells with Detects 0	# of Wells with Exceedances 0
Aluminum			mg/L	n/	'a	2	1	0.2	n/a
<i>Min</i> 0.005	1st Quartile 0.058	Median 0.2	3rd Quartile 1.1	Maximum 240	Average 3.145	# of Samples 1437	# of Wells Sampled 355	# of Wells with Detects 250	# of Wells with Exceedances 153
Antimony			ug/L	6	3	n/a	6	n/a	n/a
<i>Min</i> 0.159	1st Quartile 0.6	Median 0.8	3rd Quartile 1.1	Maximum 8.3	Average 1.066	# of Samples 1341	# of Wells Sampled 350	# of Wells with Detects 46	# of Wells with Exceedances
Arsenic			mg/L	0.0	01	n/a	0.05	n/a	n/a
Min 0	1st Quartile 0.002	Median 0.003	3rd Quartile 0.005	<i>Maximum</i> 0.14	Average 0.005	# of Samples 1565	# of Wells Sampled 381	# of Wells with Detects 247	# of Wells with Exceedances 24
Asbestos			MFL	7	,	n/a	7	n/a	n/a
<i>Min</i> 0.26	1st Quartile 0.26	Median 0.26	3rd Quartile	Maximum 0.26	Average 0.26	# of Samples 153	# of Wells Sampled 100	# of Wells with Detects 1	# of Wells with Exceedances 0
Atrazine			ug/L	3	3	n/a	1	n/a	n/a
<i>Min</i> 0.06	1st Quartile 0.06	Median 0.08	3rd Quartile 0.1	<i>Maximum</i> 1.04	Average 0.32	# of Samples 303	# of Wells Sampled 142	# of Wells with Detects	# of Wells with Exceedances
Barium			mg/L	2	2	n/a	1	n/a	n/a
Min 0	1st Quartile 0.042	Median 0.07	3rd Quartile 0.13	<i>Maximum</i> 160	Average 0.629	# of Samples 1396	# of Wells Sampled 354	# of Wells with Detects 291	# of Wells with Exceedances 10
Bentazon			ug/L	n/	'a	n/a	18	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 221	# of Wells Sampled 112	# of Wells with Detects 0	# of Wells with Exceedances 0
Benzene			ug/L	5	j	n/a	1	n/a	n/a
<i>Min</i> 0.11	1st Quartile 0.14	<i>Median</i> 0.16	3rd Quartile 0.52	<i>Maximum</i> 1.5	Average 0.4	# of Samples 2674	# of Wells Sampled 508	# of Wells with Detects 6	# of Wells with Exceedances



Chemical		Unit	Primary I	EPA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL	
Benzo(a)pyr	rene		ug/L	0.	.2	n/a	0.2	n/a	n/a
<i>Min</i> 0.02	1st Quartile 0.02	Median 0.02	3rd Quartile	Maximum 0.02	Average 0.02	# of Samples 265	# of Wells Sampled 131	# of Wells with Detects 1	# of Wells with Exceedances 0
Beryllium			mg/L	0.0	04	n/a	0.004	n/a	n/a
Min 0	1st Quartile 0	Median 0	3rd Quartile 0.001	Maximum 0.008	Average 0.001	# of Samples 1346	# of Wells Sampled 350	# of Wells with Detects 52	# of Wells with Exceedances 2
Boron			mg/L	n,	/a	n/a	n/a	n/a	1
<i>Min</i> -0.004	1st Quartile 0.1	<i>Median</i> 0.161	3rd Quartile 0.3	Maximum 2.5	Average 0.228	# of Samples 1260	# of Wells Sampled 299	# of Wells with Detects 105	# of Wells with Exceedances
Bromate			mg/L	0.0	01	n/a	0.01	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 2	# of Wells Sampled 1	# of Wells with Detects 0	# of Wells with Exceedances 0
Cadmium			mg/L	0.0	05	n/a	0.005	n/a	n/a
Min 0	1st Quartile 0	Median 0	3rd Quartile 0	Maximum 0.009	Average 0	# of Samples 1355	# of Wells Sampled 351	# of Wells with Detects 140	# of Wells with Exceedances
Carbofuran		ug/L	4	0	n/a	18	n/a	n/a	
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 210	# of Wells Sampled 116	# of Wells with Detects	# of Wells with Exceedances 0
Carbon Disu	ulfide		ug/L	n	/a	n/a	n/a	n/a	160
<i>Min</i> 0.28	1st Quartile 0.3	Median 0.54	3rd Quartile 6.6	<i>Maximum</i> 15.7	Average 3.862	# of Samples 1102	# of Wells Sampled 272	# of Wells with Detects 8	# of Wells with Exceedances 0
Carbon Tetra	achloride		ug/L	ţ	5	n/a	0.5	n/a	n/a
<i>Min</i> 0.16	1st Quartile 0.16	Median 0.9	3rd Quartile 1.2	Maximum 1.2	Average 0.753	# of Samples 2323	# of Wells Sampled 477	# of Wells with Detects	# of Wells with Exceedances 2
Chlorate			mg/L	n	/a	n/a	n/a	n/a	0.8
<i>Min</i> 0.021	1st Quartile 0.021	<i>Median</i> 0.061	3rd Quartile 0.063	Maximum 0.063	Average 0.048	# of Samples 3	# of Wells Sampled 2	# of Wells with Detects 2	# of Wells with Exceedances 0
Chlordane			ug/L	2	2	n/a	0.1	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 227	# of Wells Sampled 118	# of Wells with Detects 0	# of Wells with Exceedances 0

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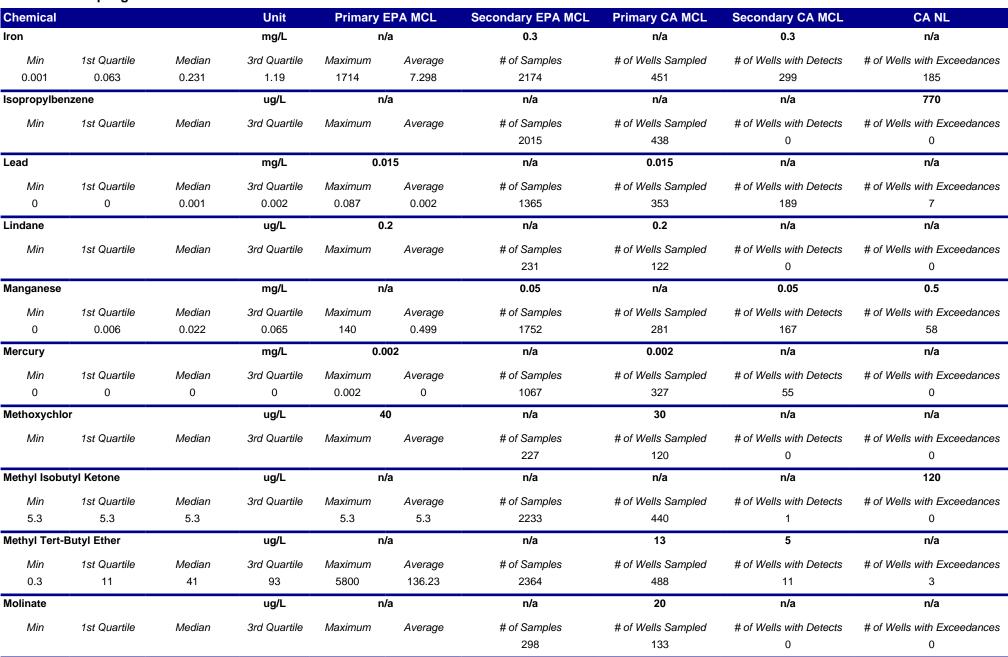
Chemical			Unit	Primary E	PA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL
Chloride			mg/L	n/	'a	250	n/a	250	n/a
<i>Min</i> 2.3	1st Quartile 12	Median 30	3rd Quartile 95	Maximum 2700	Average 68.323	# of Samples 2361	# of Wells Sampled 428	# of Wells with Detects 428	# of Wells with Exceedances 25
Chlorine			mg/L	4	ļ	n/a	4	n/a	n/a
<i>Min</i> 4.1	1st Quartile 50	Median 73	3rd Quartile 130	Maximum 486	Average 97.758	# of Samples 110	# of Wells Sampled 96	# of Wells with Detects 95	# of Wells with Exceedances 95
Chlorine Dio	oxide		mg/L	0.	8	n/a	0.8	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1	# of Wells Sampled 1	# of Wells with Detects 0	# of Wells with Exceedances 0
Chlorite			mg/L	1		n/a	1	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 4	# of Wells Sampled 3	# of Wells with Detects 0	# of Wells with Exceedances 0
Chlorobenzene		ug/L	100		n/a	70	n/a	n/a	
<i>Min</i> 0.28	1st Quartile 0.6	Median 0.79	3rd Quartile 1.4	Maximum 1.7	Average 0.962	# of Samples 2337	# of Wells Sampled 478	# of Wells with Detects 4	# of Wells with Exceedances 0
Chromium			ug/L	10	0	n/a	50	n/a	n/a
Min 0	1st Quartile 3.5	Median 6.5	3rd Quartile 13	<i>Maximum</i> 1500	Average 23.765	# of Samples 1762	# of Wells Sampled 372	# of Wells with Detects 329	# of Wells with Exceedances 30
Cis-1,2-Dich	loroethene		ug/L	70	0	n/a	6	n/a	n/a
<i>Min</i> 0.1	1st Quartile 0.7	Median 2.4	3rd Quartile 6.3	Maximum 71	Average 6.832	# of Samples 2690	# of Wells Sampled 509	# of Wells with Detects 43	# of Wells with Exceedances 10
Color			Assessment	n/	'a	15	n/a	15	n/a
<i>Min</i> 1	1st Quartile 3	Median 5	3rd Quartile 5	<i>Maximum</i> 100	Average 6.707	# of Samples 1483	# of Wells Sampled 377	# of Wells with Detects 182	# of Wells with Exceedances 21
Copper			mg/L	1.	3	1	1.3	1	n/a
<i>Min</i> 0	1st Quartile 0.001	<i>Median</i> 0.002	3rd Quartile 0.004	Maximum 150	Average 0.504	# of Samples 1768	# of Wells Sampled 370	# of Wells with Detects 277	# of Wells with Exceedances 8
Cyanide			ug/L	200		n/a	150	n/a	n/a
<i>Min</i> 8.46	1st Quartile 8.46	Median 8.46	3rd Quartile	Maximum 8.46	Average 8.46	# of Samples 450	# of Wells Sampled 173	# of Wells with Detects 1	# of Wells with Exceedances 0

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Chemical			Unit	Primary E	EPA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL
Dalapon			ug/L	20	0	n/a	200	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 227	# of Wells Sampled 118	# of Wells with Detects 0	# of Wells with Exceedances 0
Di(2-ethylhe	exyl)adipate		ug/L	400		n/a	400	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 260	# of Wells Sampled 128	# of Wells with Detects 0	# of Wells with Exceedances 0
Di(2-ethylhe	exyl)phthalate		ug/L	6	5	n/a	4	n/a	n/a
<i>Min</i> 0.77	1st Quartile 1.3	Median 3.3	3rd Quartile 8.3	Maximum 440	Average 36.405	# of Samples 261	# of Wells Sampled 124	# of Wells with Detects 9	# of Wells with Exceedances 4
Dichlorodifl	uoromethane		ug/L	n/	'a	n/a	n/a	n/a	1000
<i>Min</i> 0.17	1st Quartile 0.5	Median 0.8	3rd Quartile 2.5	Maximum 29	Average 3.07	# of Samples 2323	# of Wells Sampled 476	# of Wells with Detects 17	# of Wells with Exceedances 0
Dichlorome	thane		ug/L	5	i	n/a	5	n/a	n/a
<i>Min</i> 0.15	1st Quartile 0.17	Median 0.25	3rd Quartile 0.9	Maximum 3	Average 0.589	# of Samples 2468	# of Wells Sampled 482	# of Wells with Detects 53	# of Wells with Exceedances 0
Dinoseb			ug/L	7	,	n/a	7	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 227	# of Wells Sampled 118	# of Wells with Detects 0	# of Wells with Exceedances 0
Diquat			ug/L	2	0	n/a	20	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 198	# of Wells Sampled 108	# of Wells with Detects 0	# of Wells with Exceedances 0
Endothall			ug/L	10	00	n/a	100	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 215	# of Wells Sampled 109	# of Wells with Detects	# of Wells with Exceedances 0
Endrin			ug/L	2	2	n/a	2	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 231	# of Wells Sampled 122	# of Wells with Detects 0	# of Wells with Exceedances 0
Ethylbenzer	lbenzene ug/L 700		n/a	300	n/a	n/a			
<i>Min</i> 0.5	1st Quartile 0.6	Median 0.8	3rd Quartile 1.3	Maximum 1.7	Average 1.025	# of Samples 2380	# of Wells Sampled 481	# of Wells with Detects 8	# of Wells with Exceedances 0

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Chemical		Unit	Primary E	EPA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL	
Ethylene Dil	bromide		ug/L	0.0	05	n/a	0.05	n/a	n/a
<i>Min</i> 0.02	1st Quartile 0.02	Median 0.02	3rd Quartile 0.02	Maximum 0.02	Average 0.02	# of Samples 1227	# of Wells Sampled 360	# of Wells with Detects	# of Wells with Exceedances 0
Fluoride			mg/L	4	ļ	2	2	n/a	n/a
<i>Min</i> 0.05	1st Quartile 0.2	Median 0.3	3rd Quartile 0.7	Maximum 7.6	Average 0.538	# of Samples 1553	# of Wells Sampled 271	# of Wells with Detects 265	# of Wells with Exceedances 4
Foaming Ag	Foaming Agents		mg/L	n/	/a	0.5	n/a	0.5	n/a
<i>Min</i> 0.005	1st Quartile 0.06	<i>Median</i> 0.08	3rd Quartile 0.14	<i>Maximum</i> 18	Average 0.237	# of Samples 1140	# of Wells Sampled 226	# of Wells with Detects 76	# of Wells with Exceedances
Glyphosate			ug/L	70	00	n/a	700	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 196	# of Wells Sampled 109	# of Wells with Detects 0	# of Wells with Exceedances 0
Gross Alpha			pci/L	15		n/a	15	n/a	n/a
Min 0	1st Quartile 1.6	<i>Median</i> 2.91	3rd Quartile 4.94	Maximum 42	Average 4.283	# of Samples 440	# of Wells Sampled 127	# of Wells with Detects 93	# of Wells with Exceedances 7
Haloacetic A	Acids 5 (HAA5)		ug/L	6	0	n/a	60	n/a	n/a
<i>Min</i> 1.5	1st Quartile 8.9	<i>Median</i> 11.8	3rd Quartile 13.6	<i>Maximum</i> 90	Average 14.747	# of Samples 24	# of Wells Sampled 7	# of Wells with Detects 4	# of Wells with Exceedances
Heptachlor			ug/L	0.	4	n/a	0.01	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 232	# of Wells Sampled 122	# of Wells with Detects 0	# of Wells with Exceedances 0
Heptachlor	Epoxide		ug/L	0.	2	n/a	0.01	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 231	# of Wells Sampled 122	# of Wells with Detects 0	# of Wells with Exceedances 0
Hexachlorol	benzene		ug/L	1		n/a	1	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 271	# of Wells Sampled 137	# of Wells with Detects 0	# of Wells with Exceedances 0
Hexachloro	cyclopentadiene		ug/L	5	0	n/a	50	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 265	# of Wells Sampled 131	# of Wells with Detects 0	# of Wells with Exceedances 0







Chemical			Unit	Primary E	EPA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL
n-Butylbenz	zene		ug/L	n/	'a	n/a	n/a	n/a	260
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1531	# of Wells Sampled 364	# of Wells with Detects 0	# of Wells with Exceedances 0
N-Nitrosodi	methylamine		ug/L	n/a		n/a	n/a	n/a	0.01
<i>Min</i> 0.006	1st Quartile 0.006	<i>Median</i> 0.006	3rd Quartile	Maximum 0.006	Average 0.006	# of Samples 68	# of Wells Sampled 34	# of Wells with Detects 1	# of Wells with Exceedances 0
N-Nitrosodi	propylamine		ug/L	n/	'a	n/a	n/a	n/a	0.01
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 3	# of Wells Sampled 1	# of Wells with Detects 0	# of Wells with Exceedances 0
n-Propylbenzene		ug/L	n/	'a	n/a	n/a	n/a	260	
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1532	# of Wells Sampled 365	# of Wells with Detects 0	# of Wells with Exceedances 0
Naphthalene		ug/L	n/a		n/a	n/a	n/a	17	
<i>Min</i> 0.6	1st Quartile 0.6	Median 0.6	3rd Quartile 0.7	Maximum 0.7	Average 0.633	# of Samples 987	# of Wells Sampled 259	# of Wells with Detects	# of Wells with Exceedances 0
Nickel			mg/L	n/	'a	n/a	0.1	n/a	n/a
Min 0	1st Quartile 0.002	Median 0.003	3rd Quartile 0.007	Maximum 0.66	Average 0.013	# of Samples 1340	# of Wells Sampled 349	# of Wells with Detects 253	# of Wells with Exceedances 7
Nitrate-Nitro	ogen		mg/L	1	0	n/a	10	n/a	n/a
<i>Min</i> 0.009	1st Quartile 3.388	<i>Median</i> 7.677	3rd Quartile 15.806	Maximum 200	Average 12.759	# of Samples 8891	# of Wells Sampled 594	# of Wells with Detects 588	# of Wells with Exceedances 395
Nitrite-Nitro	gen		mg/L	1		n/a	1	n/a	n/a
<i>Min</i> 0	1st Quartile 0.05	<i>Median</i> 0.1	3rd Quartile 0.15	Maximum 35	Average 1.759	# of Samples 1827	# of Wells Sampled 402	# of Wells with Detects 124	# of Wells with Exceedances 6
Odor	1		TON	n/	'a	3	n/a	3	n/a
<i>Min</i> 1	1st Quartile 1	<i>Median</i> 1	3rd Quartile 2	Maximum 40	Average 1.69	# of Samples 1371	# of Wells Sampled 366	# of Wells with Detects 315	# of Wells with Exceedances 28
Oxamyl			ug/L	200		n/a	50	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 210	# of Wells Sampled 116	# of Wells with Detects 0	# of Wells with Exceedances 0

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Chemical			Unit	Primary E	PA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL
Pentachloro	phenol		ug/L	1		n/a	1	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 234	# of Wells Sampled 123	# of Wells with Detects 0	# of Wells with Exceedances 0
Perchlorate			ug/L	n/a		n/a	6	n/a	n/a
<i>Min</i> 0.81	1st Quartile 6	<i>Median</i> 11	3rd Quartile 20	<i>Maximum</i> 870	Average 21.406	# of Samples 2260	# of Wells Sampled 513	# of Wells with Detects 252	# of Wells with Exceedances 188
рН			рН	n/	a	8.5	n/a	n/a	n/a
Min 0	1st Quartile 7.3	Median 7.64	3rd Quartile 7.9	Maximum 770	Average 7.921	# of Samples 2319	# of Wells Sampled 394	# of Wells with Detects 394	# of Wells with Exceedances 14
Picloram	-		ug/L	50	0	n/a	500	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 227	# of Wells Sampled 118	# of Wells with Detects 0	# of Wells with Exceedances 0
Polychlorinated Biphenyls			ug/L	0.5		n/a	0.5	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 225	# of Wells Sampled 117	# of Wells with Detects 0	# of Wells with Exceedances 0
Propachlor			ug/L	n/	a	n/a	n/a	n/a	90
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 183	# of Wells Sampled 85	# of Wells with Detects 0	# of Wells with Exceedances 0
Ra 226 + Ra	228		pci/L	5	i	n/a	5	n/a	n/a
<i>Min</i> 0.16	1st Quartile 0.5	Median 0.5	3rd Quartile 0.57	Maximum 0.8	Average 0.513	# of Samples 20	# of Wells Sampled 15	# of Wells with Detects 6	# of Wells with Exceedances 0
Sec-Butylbe	nzene		ug/L	n/	'a	n/a	n/a	n/a	260
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1518	# of Wells Sampled 364	# of Wells with Detects 0	# of Wells with Exceedances 0
Selenium			mg/L	0.0)5	n/a	0.05	n/a	n/a
<i>Min</i> 0	1st Quartile 0.002	Median 0.004	3rd Quartile 0.006	Maximum 0.045	Average 0.005	# of Samples 1333	# of Wells Sampled 350	# of Wells with Detects 196	# of Wells with Exceedances 0
Silver			mg/L	n/a		0.1	n/a	0.1	n/a
<i>Min</i> 0	1st Quartile 0	Median 0	3rd Quartile 0	<i>Maximum</i> 0.014	Average 0	# of Samples 1369	# of Wells Sampled 350	# of Wells with Detects 80	# of Wells with Exceedances 0

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Chemical			Unit	Primary E	PA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL
Silvex			ug/L	50)	n/a	50	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 227	# of Wells Sampled 118	# of Wells with Detects 0	# of Wells with Exceedances 0
Simazine			ug/L	4		n/a	4	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.05	0.05	0.2	0.4	0.92	0.274	311	148	6	0
Specific Co	Specific Conductance (lab)		umhos/cm	n/	а	n/a	n/a	900	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
60	375	540	1100	1600000	3016.663	2124	335	335	121
Strontium-9	00		pci/L	n/	а	n/a	8	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
-0.35	0	0.103	0.3	1.2	0.217	63	19	18	0
Styrene			ug/L	10	0	n/a	100	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
						2291	478	0	0
Sulfate			mg/L	n/	a	250	n/a	250	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
2.4	17	50	120	1200	82.22	2913	527	527	41
TDS			mg/L	n/	a	500	n/a	500	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
48	260	380	760	4790	553.745	3945	425	425	221
Tert-Butyl A	Alcohol		ug/L	n/	a	n/a	n/a	n/a	12
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
2	2.1	9.7	22	150	37.16	968	232	3	1
Tert-Butylb	enzene		ug/L	n/	a	n/a	n/a	n/a	260
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
						1530	365	0	0
Tetrachloro	ethene	<u> </u>	ug/L	5	<u> </u>	n/a	5	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.14	1	1.8	5.7	182	7.975	3357	568	114	37



Chemical			Unit	Primary I	PA MCL	Secondary EPA MCL	Primary CA MCL	Secondary CA MCL	CA NL
Thallium			ug/L	2	2	n/a	2	n/a	n/a
<i>Min</i> -2.406	1st Quartile 0.14	<i>Median</i> 0.19	3rd Quartile 0.38	Maximum 30.72	Average 1.933	# of Samples 1260	# of Wells Sampled 349	# of Wells with Detects 41	# of Wells with Exceedances 6
Thiobencarl	b		ug/L	n/a		n/a	70	1	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 407	# of Wells Sampled 159	# of Wells with Detects	# of Wells with Exceedances 0
Toluene			ug/L	10	00	n/a	150	n/a	n/a
<i>Min</i> 0.11	1st Quartile 0.5	Median 0.71	3rd Quartile 2	Maximum 9.8	Average 1.694	# of Samples 2591	# of Wells Sampled 490	# of Wells with Detects 31	# of Wells with Exceedances 0
Total Xylene	е		ug/L	100	000	n/a	1750	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 1543	# of Wells Sampled 392	# of Wells with Detects	# of Wells with Exceedances 0
Toxaphene			ug/L	3	3	n/a	3	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples 227	# of Wells Sampled 118	# of Wells with Detects 0	# of Wells with Exceedances 0
Trans-1,2-D	ichloroethene		ug/L	10	00	n/a	10	n/a	n/a
<i>Min</i> 0.2	1st Quartile 0.28	Median 0.72	3rd Quartile 1.7	Maximum 7.73	Average 1.313	# of Samples 2703	# of Wells Sampled 509	# of Wells with Detects 12	# of Wells with Exceedances 0
Trichloroeth	hene		ug/L	Ę	i	n/a	5	n/a	n/a
<i>Min</i> 0.13	1st Quartile 1.8	Median 3.8	3rd Quartile 18	Maximum 5620	Average 64.883	# of Samples 3412	# of Wells Sampled 569	# of Wells with Detects 241	# of Wells with Exceedances 115
Trichloroflu	oromethane		ug/L	n,	'a	n/a	150	n/a	n/a
<i>Min</i> 0.07	1st Quartile 0.3	Median 0.42	3rd Quartile 0.62	<i>Maximum</i> 19	Average 1.663	# of Samples 2042	# of Wells Sampled 420	# of Wells with Detects 18	# of Wells with Exceedances 0
Trihalometh	nanes		ug/L	8	0	n/a	80	n/a	n/a
<i>Min</i> 0.5	1st Quartile 1.6	Median 4.8	3rd Quartile 64.5	Maximum 87.3	Average 28.432	# of Samples 618	# of Wells Sampled 215	# of Wells with Detects 23	# of Wells with Exceedances
Tritium			pci/L	n/a		n/a	20000	n/a	n/a
<i>Min</i> -199	1st Quartile -12.6	Median 25.7	3rd Quartile 287	Maximum 596	<i>Average</i> 118.69	# of Samples 65	# of Wells Sampled 18	# of Wells with Detects 18	# of Wells with Exceedances 0



Chemical Turbidity		Unit NTU	Primary EPA MCL 5		Secondary EPA MCL n/a	Primary CA MCL	Secondary CA MCL 5	CA NL n/a	
									Min
0	0.21	0.52	2.6	2880	21.599	1699	360	320	78
Uranium			pci/L	n/a		n/a	20	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.48	1.58	2.77	5.48	20.5	4.319	175	54	53	1
Vanadium			mg/L	n/a		n/a	n/a	n/a	0.05
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.001	0.009	0.013	0.025	0.31	0.02	817	290	286	25
Vinyl Chloride			ug/L	2	2	n/a	0.5	n/a	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
						2389	483	0	0
Zinc		n		mg/L n/a		5	n/a	5	n/a
Min	1st Quartile	Median	3rd Quartile	Maximum	Average	# of Samples	# of Wells Sampled	# of Wells with Detects	# of Wells with Exceedances
0.001	0.003	0.006	0.014	9.853	0.061	1804	369	264	1

Primary EPA MCL	Primary EPA MCLs are federally enforceable limits for chemicals in drinking water and are set as close as feasible to the corresponding EPA MCLG.					
Secondary EPA	Secondary EPA MCLs apply to chemicals in drinking water that adversely affect its odor, taste, or appearance Secondary EPA MCLs are not based on direct health effects associated with the chemical. Secondary MCLs are consdered desireable goals and are not federally enforceable.					
Primary CA MCL	Primary CA MCLs are analogous to Primary EPA MCLs and are enforceable at the state level . If the California DHS has adopted a more stringent primary MCL than the EPA MCL, the primary CA MCL sould be enforceable.					
Secondary CA	Secondary CA MCLs are analogous to Secondary EPA MCLs and are applicable at the state level. If the California DHS has adopted a more stringent secondary MCL than the EPA MCL, the secondary CA MCL would be applied.					
CA NL	California Notification Levels are health -based criteria similar to US EPA Health Advisories. CA NLs are not enforceable, but are levels at which the California Department of Health Services strongly urges water purveyors to take corrective actions.					

